

1-1-2012

Optimization of colorimetric DET technique for the in situ, two-dimensional measurement of iron(II) distributions in sediment porewaters

William W. Bennett
Griffith University

Peter R. Teasdale
Griffith University

David T. Welsh
Griffith University

Jarad G. Panther
Griffith University

Dianne F. Jolley
University of Wollongong, djolley@uow.edu.au

Follow this and additional works at: <https://ro.uow.edu.au/scipapers>



Part of the [Life Sciences Commons](#), [Physical Sciences and Mathematics Commons](#), and the [Social and Behavioral Sciences Commons](#)

Recommended Citation

Bennett, William W.; Teasdale, Peter R.; Welsh, David T.; Panther, Jarad G.; and Jolley, Dianne F.: Optimization of colorimetric DET technique for the in situ, two-dimensional measurement of iron(II) distributions in sediment porewaters 2012, 490-495.
<https://ro.uow.edu.au/scipapers/4260>

Optimization of colorimetric DET technique for the in situ, two-dimensional measurement of iron(II) distributions in sediment porewaters

Keywords

iron, measurement, dimensional, two, situ, technique, det, porewaters, sediment, colorimetric, distributions, optimization, ii, CMMB

Disciplines

Life Sciences | Physical Sciences and Mathematics | Social and Behavioral Sciences

Publication Details

Bennett, W. W., Teasdale, P. R., Welsh, D. T., Panther, J. G. & Jolley, D. F. (2012). Optimization of colorimetric DET technique for the in situ, two-dimensional measurement of iron(II) distributions in sediment porewaters. *Talanta*, 88 (N/A), 490-495.

1 Optimization of colorimetric DET technique for the in
2 situ, two-dimensional measurement of iron(II)
3 distributions in sediment porewaters

4 *William W. Bennett^a, Peter R. Teasdale^{a*}, David T. Welsh^a, Jared G. Panther^a and Dianne F. Jolley*

5 *b*

6 ^a Environmental Futures Centre, Griffith University, Gold Coast campus, QLD 4222, Australia

7 ^b School of Chemistry, University of Wollongong, Wollongong, NSW 2522, Australia

8
9 *Corresponding Author: p.teasdale@griffith.edu.au

10 Ph: +617 55528358, Fax: +617 55528067

16 **Abstract**

17 The recently developed colorimetric diffusive equilibration in thin films (DET) technique for the in
18 situ, high-resolution measurement of iron(II) in marine sediments is optimized to allow measurement of
19 the higher iron concentrations typical of sediment porewaters. Computer imaging densitometry (CID)
20 is used to analyze the retrieved samplers following exposure to ferrozine, a colorimetric reagent
21 selective for iron(II). The effect of ferrozine concentration, image processing parameters and ionic
22 strength are investigated to improve the applicability of this technique to a wider range of aquatic
23 systems than reported in the first publications of this approach. The technique was optimized to allow
24 detection of up to 2000 $\mu\text{mol L}^{-1}$ iron(II), a four-fold increase on the previous upper detection limit of
25 500 $\mu\text{mol L}^{-1}$. The CID processing of the scanned color image was also optimized to adjust the
26 sensitivity of the assay as required; by processing the image with different color channel filters, the
27 sensitivity of the assay can be optimized for lower concentrations (up to 100 $\mu\text{mol L}^{-1}$) or higher
28 concentrations (up to 2000 $\mu\text{mol L}^{-1}$) of iron(II), depending on the specific site characteristics. This
29 process does not require separate sampling probes or even separate scans of the DET gels as the color
30 filter and grayscale conversion is done post-image capture. The optimized technique is very simple to
31 use and provides highly representative, high-resolution (1 mm) two-dimensional distributions of
32 iron(II) in sediment porewaters. The detection limit of the optimized technique was $4.1 \pm 0.3 \mu\text{mol L}^{-1}$
33 iron(II) and relative standard deviations were less than 6%.

34
35 **Keywords:** DET; CID; iron; biogeochemistry; ferrozine; high-resolution

36

37

38

39 **1.0 Introduction**

40 The study of sediment biogeochemistry is a challenging and complex area due to heterogeneous
41 chemical distributions. Traditional sediment sampling techniques require the removal and processing of
42 sediment cores [1-3]. Porewater profiles are typically obtained by slicing and centrifugation of
43 sediment samples, resulting in large volumes of sediment porewater being mixed and thus the
44 measured concentrations representing an average of the sampled volume [4, 5]. This process can
45 introduce sampling errors, particularly where reactive chemical species may be present in the same
46 sample (e.g. iron(II) and sulfide) [3-7]. Furthermore, these traditional methods of porewater analysis
47 fail to provide data at sufficient spatial resolution to allow investigation of heterogeneous solute
48 distributions or concentration gradients of sediment porewater solutes, which often occur on the
49 millimeter scale [1, 2, 8]. Diffusion-based techniques, such as diffusive equilibration in thin films
50 (DET), avoid the disadvantages of traditional sediment sampling as they measure porewater profiles *in*
51 *situ*, thus representing the actual chemical distributions present in the sediment [1, 4]. Additionally,
52 they allow analysis at high spatial resolution (millimeter to sub-millimeter) [9], in two-dimensions [10]
53 and only sample a very small volume of sediment porewater; approximately 10 μL for a gel analyzed at
54 1 square millimeter resolution (based on a 10 mm depletion distance; see Harper et al. [11] for details),
55 thereby allowing small-scale heterogeneity to be observed and mechanistic interactions to be assessed
56 and interpreted [8]. The interpretation of DET measurements has been comprehensively discussed by
57 Harper and co-workers [11] and Davison and co-workers [12] and the techniques have been utilized for
58 the measurement of trace metals in both fresh [13, 14] and marine [15, 16] sediment porewaters.

59

60 *In situ* sampling is particularly important for redox-sensitive analytes such as iron, which may be
61 present as reduced iron(II) in the sediment and is rapidly oxidized to iron(III) upon exposure to oxygen
62 [6, 7, 17, 18]. The biogeochemical cycling of iron is of particular importance due to its role as both a
63 source and sink for many other important solutes including phosphorus and arsenic, both of which are

64 present as oxyanions in water, are strongly bound to solid-phase iron oxides and are remobilized via
65 reductive dissolution of iron(III) minerals [19-21]. Understanding the biogeochemistry of iron is
66 integral to determining the cycling of these solutes and how other factors such as eutrophication-
67 induced or seasonal anoxia can affect their fluxes both to and from the sediment.

68

69 A colorimetric DET technique recently developed by Jezequel and co-workers [22] and Robertson and
70 co-workers [6] allows the two-dimensional measurement of iron(II) distributions in sediment
71 porewaters. A colorimetric reagent, selective for iron(II), was equilibrated with a hydrogel that was
72 then placed onto the hydrogel of a DET sampler, immediately following removal from the sediment.
73 The colored complex that formed within the two gels was analyzed by scanning on a conventional
74 flatbed scanner and converting the grayscale intensities to iron(II) concentrations utilizing a calibration
75 curve [6]. This approach, known as computer imaging densitometry (CID), was first applied to
76 environmental analysis for the measurement of dissolved sulfide [23] by a diffusive gradients in thin
77 films (DGT) technique. This technique has been applied to the investigation of iron(II) and sulfide
78 distributions in mesocosms containing bioturbators [24] and to study the effect of seagrass on iron(II)
79 and sulfide distributions in natural sediments [25]. A colorimetric DET technique similar to that
80 described in this study has also been developed for the measurement of phosphate distributions in
81 sediment porewaters [26]. These diffusive techniques utilizing CID have proven to be useful and
82 simple techniques for quantitative imaging of sediment solute distributions in one or two dimensions.
83 Unfortunately, the maximum measurement range of the colorimetric DET techniques developed for
84 iron(II) to date is $0.6 - 500 \mu\text{mol L}^{-1}$ [6], which is not sufficiently large to allow the measurement of
85 the full range of iron(II) concentrations typically seen in many freshwater sediment porewaters where
86 iron(II) is not constrained by sulfide [27]. Furthermore, the technique has not been evaluated with
87 respect to temperature, ionic strength ranges or reagent concentration. In this study, we optimized the

88 colorimetric DET technique for iron(II) by modifying the concentration of colorimetric reagent and the
89 CID parameters to allow the quantification of a much larger range of iron(II) concentrations.

90

91 **2.0 Experimental**

92 **2.1 Reagents, materials and solutions**

93 Ultra-pure deionized water (Milli-Q Element, Millipore) was used for the preparation of all solutions.
94 Bisacrylamide-cross linked hydrogels for the DET technique were prepared as described previously
95 [6]. Ferrous iron stock solutions were prepared by dissolving ammonium iron(II) sulfate (AR Grade,
96 Merck) in 0.01 mol L⁻¹ hydroxylamine hydrochloride (ACS Grade, Alfa Aesar). All dilutions of the
97 ferrous iron stock solution were made in 0.01 mol L⁻¹ hydroxylamine hydrochloride in 0.01 mol L⁻¹
98 sodium chloride (AR Grade, Merck). Hydroxylamine was added to prevent oxidation of iron(II) to
99 iron(III) during experiments. The DET colorimetric reagent consisted of 0.01 mol L⁻¹ ferrozine (3-(2-
100 pyridyl)-5,6-diphenyl-1,2,4-triazine-4',4''-disulfonic acid sodium salt, Fluka) in an acetate buffer
101 consisting of 0.1 mol L⁻¹ acetic acid (AR Grade, Merck) and 0.1 mol L⁻¹ sodium acetate (AR Grade,
102 Chem. Supply). DET sediment probe assemblies were purchased from DGT Research Ltd., washed in
103 10% (v/v) HNO₃ (AR Grade, Merck) for 24 h and then rinsed in deionized water prior to assembly as
104 described previously [6]. An iron(II) quality control standard was prepared from a NIST-certified 1000
105 mg L⁻¹ iron standard (High Purity Standards, Charleston) reduced to iron(II) by addition of
106 hydroxylamine hydrochloride. An aliquot of 1 mol L⁻¹ hydroxylamine hydrochloride was added to the
107 1000 mg L⁻¹ iron standard in a ratio of 1:10 and left to react for at least 1 h before use. This standard
108 was then diluted to 100 μmol L⁻¹ (5.58 mg L⁻¹) in 0.01 mol L⁻¹ hydroxylamine hydrochloride and 0.01
109 mol L⁻¹ sodium chloride.

110

111

112

113 **2.2 Computer imaging densitometry (CID)**

114 CID analysis was done with a commercially available flatbed scanner (Canon LiDE 200) and image
115 processing software (GIMP v2.6, available free of charge from <http://www.gimp.org>). Once exposed to
116 the staining gel, the iron(II) within the DET gel reacted with the ferrozine reagent forming a magenta
117 colored complex. To ensure minimal loss in profile fidelity during the staining procedure, it is
118 important that the staining gel is well drained following removal from the ferrozine reagent so that
119 iron(II) reacts with ferrozine within the gel structure, rather than with excess ferrozine reagent present
120 on the surface of the gel. The optimized staining time for this technique was 10 minutes (Section 3.2).
121 The stained gels were scanned in color at 300 dpi resolution, saved as a TIFF file and then processed in
122 GIMP. Processing of the images involved conversion to grayscale so that color variation was
123 eliminated and the grayscale intensity of the image was the only variable being measured. The channel
124 mixer function in GIMP can be used to convert images to grayscale using a red, green or blue color
125 filter, or a combination thereof. The optimization of this step is discussed in Section 3.1. Following the
126 selection of an appropriate color filter, the image was converted into grayscale, resized so that one pixel
127 equaled one square millimeter and saved as a TIFF image file. The maximum resolution of this
128 technique (one square millimeter) is limited by the thickness of the diffusive layer; the use of thinner
129 diffusive gels will permit analysis at higher resolution [28]. Grayscale images were imported into
130 ImageJ (available free of charge from <http://rsbweb.nih.gov/ij/>), where grayscale values were measured
131 and exported to Microsoft Excel. Blank gels were analysed in the same way and blank grayscale values
132 were subtracted from the sample grayscale values prior to calculation of concentration.

133

134 **2.3 Ferrozine concentration, color channel and staining time**

135 The method was optimized by varying two important parameters: the concentration of ferrozine in the
136 staining gel and the color channel used for grayscale conversion. The technique described by Robertson
137 and co-workers [6] relied on the reaction of iron(II) (that is equilibrated within the retrieved

138 bisacrylamide hydrogel sampling gel) with a second hydrogel equilibrated with 0.001 mol L⁻¹ ferrozine
139 in 0.1 mol L⁻¹ acetic acid (staining gel), which was overlain onto the DET sampling gel immediately
140 after retrieval. The iron(II) diffuses from the retrieved gel into the staining gel, forming a magenta
141 colored complex with ferrozine in a 3:1 molar ratio (ferrozine: iron) [29]. Robertson and co-workers [6]
142 reported an upper measurement limit of 500 μmol L⁻¹ for the technique. In the present study we
143 increased the ferrozine concentration ten-fold to 0.01 mol L⁻¹ and evaluated this against the procedure
144 used by Robertson and co-workers [6].

145

146 Calibration standards were prepared at 0, 10, 20, 50, 75, 100, 200, 500, 1000, 2000 μmol L⁻¹ in 0.01
147 mol L⁻¹ sodium chloride and 0.01 mol L⁻¹ hydroxylamine hydrochloride, to prevent oxidation of
148 iron(II). Gel discs (n=3 for each calibration standard) were equilibrated in the calibration solutions for
149 at least 1 h, removed and placed on a ferrozine staining gel and covered with a clear acetate
150 transparency sheet. The calibration curves were scanned at 5, 10, 15 and 20 minutes, and processed as
151 described in Section 2.2.

152

153 To determine the optimal color channel for grayscale conversion of the acquired image, calibration
154 standards were analyzed and converted to grayscale utilizing either a red, green, or blue color channel
155 filter. The selection of the best color channel filter was based on the requirements of sensitivity and
156 upper measurement limits.

157

158 **2.4 Effect of ionic strength**

159 Gel discs were equilibrated with 100 or 1000 μmol L⁻¹ iron(II) prepared in 0.01 mol L⁻¹ hydroxylamine
160 hydrochloride and 0.01, 0.1 or 0.7 mol L⁻¹ sodium chloride. The gels were then stained and scanned as
161 described in Section 2.2.

162

163 **2.5 Deployment in a freshwater sediment mesocosm**

164 Freshwater sediment was collected from the Coomera River on the Gold Coast, Queensland, Australia,
165 sieved to 1 mm to remove biota and large debris and 13 L placed into a plastic mesocosm (30 cm Ø, 30
166 cm height) with 7 L of overlying water collected from the same site. The mesocosm was allowed to
167 stabilize for ~3 months prior to deployment. A DET sediment probe with a 0.8 mm bisacrylamide
168 hydrogel was deployed in the mesocosm for 24 h. Upon retrieval the probe was briefly rinsed with
169 deionized water and the bisacrylamide hydrogel cut out of the front window using a razor blade. This
170 gel was then analyzed for iron(II) as per Section 2.2. The time between removal and staining of the
171 probe was less than 1 minute to minimize the oxidation of iron(II) in air.

172

173 **3.0 Results and Discussion**

174 **3.1 Ferrozine concentration and color channel**

175 The existing colorimetric iron(II) DET technique has an upper measurement range of 500 $\mu\text{mol L}^{-1}$. We
176 propose this could be due to an insufficient concentration of ferrozine reagent (0.001 mol L^{-1}), which is
177 required in a 3:1 ratio of ferrozine to iron [29]. Assuming that all iron(II) in the sample gel reacted with
178 all ferrozine in the staining gel, a concentration of 0.001 mol L^{-1} ferrozine would limit measurement to
179 $\sim 330 \mu\text{mol L}^{-1}$ iron(II). By increasing the concentration of ferrozine reagent to 0.01 mol L^{-1} , higher
180 concentrations of iron(II) should theoretically be measurable. Figure 1 shows the results of calibration
181 curves made using 0.01 mol L^{-1} ferrozine reagent for a concentration range up to 2000 $\mu\text{mol L}^{-1}$ iron(II)
182 with a staining time of 10 minutes (see Section 3.2).

183

184 The original colorimetric iron(II) DET technique utilized a green channel filter during grayscale
185 conversion to maximize the sensitivity of the measurement. While this does improve the sensitivity of

186 the technique, it also means the maximum grayscale values are reached earlier in the concentration
187 range. In this case, grayscale intensity plateaus at iron(II) concentrations $>500 \mu\text{mol L}^{-1}$ for the
188 calibration curve processed with a green filter. By utilizing a different color channel filter, it is possible
189 to adjust the sensitivity to allow higher concentrations of iron(II) to be measured. The red channel filter
190 results in a calibration curve that can be used to measure concentrations as high as $2000 \mu\text{mol L}^{-1}$
191 iron(II) with an acceptable degree of accuracy. Although this approach results in some sensitivity loss
192 at the low end of the concentration range (see Table 1 for detection limits), this is not expected to be an
193 issue for most deployments. Processing the calibration data with a blue channel filter resulted in a very
194 similar result to the red channel filter, but with linearity to $100 \mu\text{mol L}^{-1}$ opposed to $200 \mu\text{mol L}^{-1}$ for
195 the red channel. Therefore, the blue channel calibration was not investigated any further.

196

197 The reason for the different sensitivities when using the different color channel filters is related to the
198 magenta color of the ferrozine complex. Magenta is composed primarily of the blue and red
199 wavelengths of light, with a relatively small proportion of green wavelengths. When a single color is
200 converted to grayscale, the more intense the color, the higher the grayscale value it will have, resulting
201 in grayscale intensity closer to white. When applying a green filter, the red and blue wavelengths that
202 form the majority of the magenta color do not contribute to the grayscale intensity, so the images are
203 darker (not as close to white) than if the red or blue wavelengths were included. This means that the
204 grayscale values approaches the black end of the spectrum over a smaller range of color intensities
205 (iron(II) concentrations), which increases sensitivity but limits the maximum measurement range.
206 When a red filter is used, the red wavelengths of light, which form a larger proportion of the magenta
207 color than green wavelengths, contribute to the grayscale value so that black is approached over a
208 larger range of color intensities (iron(II) concentrations), effectively increasing the measurement range
209 but decreasing the sensitivity.

210

211 In order to use Figure 1 to calibrate the technique, a curve needed to be fitted to the data. A logarithmic
212 function fitted the majority of the data well but due to the large range of concentrations, did not provide
213 an accurate fit for the lower concentrations. Therefore, a new approach for calibration was taken,
214 whereby the data was split into two separate calibration curves (Figure 2).

215

216 The grayscale intensity followed a linear relationship up to $200 \mu\text{mol L}^{-1}$ iron(II), after which point a
217 logarithmic relationship fitted the data up to the maximum of $2000 \mu\text{mol L}^{-1}$ iron(II). This split
218 calibration curve allows a more accurate calibration to be performed compared to trying to fit a single
219 curve to the entire data set. For blank-corrected grayscale intensity values of ≤ 46 the linear calibration
220 should be used, whereas for grayscale intensity values above this the logarithmic calibration should be
221 used.

222

223 The two calibration curves meet at the $200 \mu\text{mol L}^{-1}$ calibration point, corresponding to a grayscale
224 intensity of 45.8. To determine how well the calibration curves agree at this point, the predicted
225 concentration for this grayscale intensity value was calculated from both calibration curve equations.
226 The linear calibration curve predicted an iron(II) concentration of $197.8 \mu\text{mol L}^{-1}$ and the logarithmic
227 calibration curve predicted an iron(II) concentration of $201.1 \mu\text{mol L}^{-1}$ for the same grayscale value.
228 This is a difference of approximately 2%, indicating that both calibration curves accurately predict the
229 iron(II) concentration at the crossover point of the calibration range, and thus that both calibration
230 curves can be used in conjunction to measure a wide range of iron(II) concentrations.

231

232 By processing the calibration curve using a green channel filter, as described by Robertson and co-
233 workers [6], the sensitivity of the measurement is increased. If concentrations of iron(II) are $< 100 \mu\text{mol}$

234 L^{-1} (corresponding to grayscale intensity on the red channel image of <24) the image can be
235 reprocessed to grayscale using the green channel filter and the green channel calibration curve applied
236 (Figure 3). The higher sensitivity of the green channel is evident from the slope of the regression line
237 for the green channel filter calibration, which is more than three-fold higher than the linear calibration
238 curve obtained using the red channel filter. The ability to adjust the sensitivity of the measurement by
239 simply reprocessing the image with a different color channel filter results in a versatile and useful
240 measurement technique that can be adapted to a wide range of concentrations.

241

242 **3.2 Staining time and temperature**

243 The effect of staining time was evaluated by scanning the calibration curves at five-minute intervals for
244 30 minutes. For concentrations less than $200 \mu\text{mol L}^{-1}$ the grayscale values of the red color channel
245 calibration curve did not change after 10 minutes of staining time, indicating that the color reaction was
246 complete by this stage. For concentrations higher than $200 \mu\text{mol L}^{-1}$ the color reaction took up to 25
247 minutes to go to completion, as indicated by no further change in grayscale intensity. However, as long
248 as the staining time is controlled accurately during the analysis, the color reaction is not required to be
249 complete to allow quantitative measurement of iron(II). After 10 minutes of staining time, the grayscale
250 intensity of the highest calibration standard is $\sim 80\%$ of the value after 25 minutes. Therefore, given that
251 diffusive relaxation will increase over time thereby decreasing the fidelity of the measured iron(II)
252 distributions (see Harper et al. [11] for details), a 10 minute staining time was chosen to give the best
253 compromise between color development and minimizing diffusive relaxation of the iron(II)
254 distribution. Similar results were observed for the green color channel calibration curve, with no further
255 color development after 10 minutes of staining time, indicating that 10 minutes is an appropriate
256 staining time to use regardless of the calibration curve utilized. This differs from Robertson and co-
257 workers [6], who found an optimal staining time of 15 minutes, probably due to the higher

258 concentration of ferrozine reagent used in this study that ensured rapid reaction times and a sufficient
259 excess of the ferrozine reagent.

260

261 During the initial development of this optimized method the effect of temperature on color
262 development was evaluated by preparing, staining and analyzing calibration curves at 15, 25 and 35°C.

263 This range of temperatures had a negligible effect on color development, with the slopes of the linear
264 calibration curves from 0 – 200 $\mu\text{mol L}^{-1}$ agreeing to within 3%. For analysis temperatures outside of
265 the tested range the authors recommend performing a field calibration as described in Section 3.4.

266

267 **3.3 Quality control, method detection limits and reproducibility**

268 Method detection limits (3σ) were calculated based on the standard deviation of grayscale values from
269 blank gels analyzed at $1 \text{ mm}^2 = 1 \text{ pixel}$ resolution ($n=100$ pixels). Mean values were obtained by
270 replicating this analysis three times. These method detection limits are higher than those reported by
271 Robertson and co-workers ($0.6 \mu\text{mol L}^{-1}$) [6] due to a more representative MDL calculation being used
272 in this study. Robertson and co-workers calculated their MDL based on the variation between three gels
273 that were stained, scanned and then their grayscale intensity values averaged over the area of the gel,
274 effectively eliminating the variation in grayscale intensity values on the single pixel scale. This
275 technique is most commonly used on the single pixel scale to obtain high-resolution measurements,
276 therefore, the MDL should represent this by taking the variation of each pixel into account, as the
277 authors have done in this case.

278

279 A $100 \mu\text{mol L}^{-1}$ quality control standard prepared from a separate, NIST-certified 1000 mg L^{-1} iron
280 standard solution was also analysed to investigate the accuracy and reproducibility of the technique
281 (Table 1).

282

283 The QC recovery is an average from four gels, with the error of the mean representing the standard
284 deviation between gels. The reproducibility, however, represents the mean variation of grayscale
285 intensity within each QC gel analyzed at $1 \text{ mm}^2 = 1 \text{ pixel}$ resolution, which better represents the
286 reproducibility of the technique when deployed *in situ* and analyzed as a high-resolution, two-
287 dimensional technique. Reproducibility was excellent both between and within scanned gels. The QC
288 recoveries indicate that both approaches are capable of accurately measuring iron(II) concentrations.

289

290 The MDL is higher when processing and analyzing the data based on the red channel calibration
291 compared to the green channel calibration, which is to be expected based on the different sensitivities
292 of the two methods. These detection limits are sufficiently low to allow analysis of a wide range of
293 iron(II) concentrations in sediment porewaters, especially considering the representative nature of the
294 DET measurement [1, 4] and the high-resolution, two-dimensional capabilities. The reproducibility is
295 also higher when using the green channel calibration, although both are sufficiently low to ensure
296 precise results. Additionally, since data from the same scanned image can be analysed using both color
297 filter channels, the image or specific parts of the image can be analysed using the calibration most
298 appropriate to the iron(II) concentration range present.

299

300 The interference of the colorimetric measurement of ferrous iron by major inorganic ions has been
301 previously investigated by Stookey [29]. They found that of the major ions tested, only divalent cobalt
302 and monovalent copper caused a minor positive interference; copper present at five times the ferrous
303 iron concentration resulted in a 15% overestimation and cobalt present at eight times the ferrous iron
304 concentration resulted in a 5% overestimation [29]. This minor positive interference will be negligible
305 in sediment porewater where iron concentrations will typically be much higher than most other metal
306 ions.

307 **3.4 Effect of ionic strength**

308 The effect of ionic strength on the colorimetric DET technique was evaluated by analysis of hydrogel
309 discs equilibrated in 100 or 1000 $\mu\text{mol L}^{-1}$ iron(II) standard solutions prepared at 0.01, 0.1 and 0.7 mol
310 L^{-1} sodium chloride (Table 2).

311

312 A minor effect of ionic strength is observed, which is more pronounced at 100 $\mu\text{mol L}^{-1}$ iron(II) than at
313 1000 $\mu\text{mol L}^{-1}$ iron(II). The difference between the highest and lowest recoveries is 13% and 6% for
314 the 100 and 1000 $\mu\text{mol L}^{-1}$ iron(II) standards, respectively. This difference is low and so calibrations at
315 individual ionic strengths are unnecessary. However, given the ease-of-use of this technique,
316 calibrations curves could be analyzed in the field alongside samples at matrix-matched ionic strengths
317 for improved accuracy. This would involve the preparation of stable iron(II) standard solutions
318 (prepared in a reducing agent such as hydroxylamine hydrochloride) in which polyacrylamide
319 hydrogels are equilibrated; these could then be removed, stained and analysed alongside retrieved gel
320 probes on-site. The effect of matrix pH on this technique is not a factor requiring optimization due to
321 the acetate buffer used in the analysis and the wide pH working range of the ferrozine reagent (pH 4 –
322 9) [29].

323

324 **3.5 Sediment deployment of optimized technique**

325 To demonstrate the effectiveness of the optimized technique it was used to obtain an iron(II) porewater
326 distribution in a freshwater sediment mesocosm (Figure 4).

327

328 The results show iron(II) first appearing at 20 mm depth, followed by an increase over the following
329 100 mm of the profile. The concentration reaches a maximum of 720 $\mu\text{mol L}^{-1}$ at a depth of 120 mm,
330 which is above the maximum measurement range of the original ferrozine DET technique but within

331 the range of the optimized technique described in this study. This demonstrates the advantage of the
332 optimized technique, as it is able to measure the higher iron(II) concentrations typical of freshwater
333 sediments. The two-dimensional distribution provides a higher degree of representativeness compared
334 to the averaged depth profile, showing the lateral and vertical variation in iron(II) porewater
335 concentration. In this case, the variation is low due to the homogenization of the sediment prior to
336 establishment of the mesocosm. However, in natural sediments colonized by infauna or rooted plants
337 porewater solute concentrations can exhibit a very high degree of lateral variation [6, 24, 26, 30]
338 making the use of methods such as the described technique for iron(II), which are able to measure
339 solute distributions in two-dimensions, essential.

340

341 **4.0 Conclusions**

342 The optimized colorimetric DET technique for the measurement of porewater iron(II) described in this
343 study provides a simple and effective method of investigating iron(II) biogeochemistry. The expanded
344 measurement range allows this technique to be used in freshwater sediment where iron(II)
345 concentrations are typically much higher than in marine sediments. Furthermore, this technique has
346 numerous potential applications when coupled with the diffusive gradients in thin films (DGT) or
347 complimentary colorimetric DET techniques for determining co-distributions of porewater solutes at
348 the same location in the sediment in either one or two dimensions [6, 24]. The representativeness, ease-
349 of-use and rapid data acquisition makes this optimized technique an important tool for the analysis of
350 iron(II) biogeochemistry and the interpretation of analytes with closely linked biogeochemical cycles.

351

352

353

354

355

356 **5.0 References**

- 357 [1] W. Davison, H. Zhang, G.W. Grime, *Environ. Sci. Technol.*, 28 (1994) 1623-1632.
358 [2] P.R. Teasdale, G.E. Batley, S.C. Apte, I.T. Webster, *Trends Anal. Chem.*, 14 (1995) 250-256.
359 [3] E. Viollier, C. Rabouille, S. Aplitz, E. Breuer, G. Chaillou, K. Dedieu, Y. Furukawa, C. Grenz, P.
360 Hall, F. Janssen, *J. Exp. Mar. Biol. Ecol.*, 285 (2003) 5-31.
361 [4] S.E. Bufflap, H.E. Allen, *Water Res.*, 29 (1995) 165-177.
362 [5] A. Stockdale, W. Davison, H. Zhang, *Earth-Sci. Rev.*, 92 (2009) 81-97.
363 [6] D. Robertson, P.R. Teasdale, D.T. Welsh, *Limnol. Oceanogr. Methods*, 6 (2008) 502-512.
364 [7] P.M. Chapman, F. Wang, J.D. Germano, G. Batley, *Mar. Pollut. Bull.*, 44 (2002) 359-366.
365 [8] H. Zhang, W. Davison, R.J.G. Mortimer, M.D. Krom, P.J. Hayes, I.M. Davies, *Sci. Total Environ.*,
366 296 (2002) 175-187.
367 [9] G.R. Fones, W. Davison, G.W. Grime, *Sci. Total Environ.*, 221 (1998) 127-137.
368 [10] S.M. Shuttleworth, W. Davison, J. Hamilton-Taylor, *Environ. Sci. Technol.*, 33 (1999) 4169-4175.
369 [11] M.P. Harper, W. Davison, W. Tych, *Environ. Sci. Technol.*, 31 (1997) 3110-3119.
370 [12] W. Davison, G. Fones, M. Harper, P. Teasdale, H. Zhang, J. Buffle, G. Horvai, *Dialysis, DET, and*
371 *DGT: In situ Diffusional Techniques for Studying Water, Sediments and Soils.*, in: J. Buffle, G. Horvai
372 (Eds.) *In Situ Monitoring of Aquatic Systems: Chemical Analysis and Speciation*, John Wiley & Sons
373 Ltd., Chichester, 2000, pp. 495-569.
374 [13] Y. Gao, M. Leermakers, M. Elskens, G. Billon, B. Ouddane, J.C. Fischer, W. Baeyens, *Sci. Total*
375 *Environ.*, 373 (2007) 526-533.
376 [14] Y. Gao, M. Leermakers, C. Gabelle, P. Divis, G. Billon, B. Ouddane, J.C. Fischer, M. Wartel, W.
377 Baeyens, *Sci. Total Environ.*, 362 (2006) 266-277.
378 [15] G.R. Fones, W. Davison, O. Holby, B.B. Jorgensen, B. Thamdrup, *Limnol. Oceanogr.*, (2001)
379 982-988.
380 [16] S. Tankere-Muller, H. Zhang, W. Davison, N. Finke, O. Larsen, H. Stahl, R.N. Glud, *Mar. Chem.*,
381 106 (2007) 192-207.
382 [17] B. Morgan, O. Lahav, *Chemosphere*, 68 (2007) 2080-2084.
383 [18] S.L. Simpson, G.E. Batley, *Environ. Toxicol. Chem.*, 22 (2003) 424-432.
384 [19] N. Belzile, A. Tessier, *Geochim. Cosmochim. Acta*, 54 (1990) 103-109.
385 [20] P. Bose, A. Sharma, *Water Res.*, 36 (2002) 4916-4926.
386 [21] T.F. Rozan, M. Taillefert, R.E. Trouwborst, B.T. Glazer, S. Ma, J. Herszage, L.M. Valdes, K.S.
387 Price, G.W. Luther III, *Limnol. Oceanogr.*, (2002) 1346-1354.
388 [22] D. Jézéquel, R. Brayner, E. Metzger, E. Viollier, F. Prévot, F. Fiévet, *Estuar. Coast. Shelf. Sci.*, 72
389 (2007) 420-431.
390 [23] P. Teasdale, S. Hayward, W. Davison, *Anal. Chem.*, 71 (1999) 2186-2191.
391 [24] D. Robertson, D.T. Welsh, P.R. Teasdale, *Environ. Chem.*, 6 (2009) 60-69.
392 [25] J. Deborde, G. Abril, A. Mouret, D. Jézéquel, G. Thouzeau, J. Clavier, G. Bachelet, P. Anschutz,
393 *Mar. Ecol. Prog. Ser.*, 355 (2008) 59.
394 [26] A. Pages, P.R. Teasdale, D. Robertson, W.W. Bennett, J. Schäfer, D.T. Welsh, *Chemosphere*,
395 (2011).
396 [27] R.J. Donahoe, C. Liu, *Environ. Geol.*, 33 (1998) 143-153.
397 [28] W. Davison, G.W. Grime, J.A.W. Morgan, K. Clarke, *Nature*, 352 (1991) 323-325.
398 [29] L.L. Stookey, *Anal. Chem.*, 42 (1970) 779-781.
399 [30] A. Widerlund, W. Davison, *Environ. Sci. Technol.*, 41 (2007) 8044-8049.

400

401

402 **Table Captions**

403

404 **Table 1. Method detection limits, quality control recoveries and reproducibility of the optimized**
405 **colorimetric iron(II) DET technique. Values are means \pm 1 standard deviation (n=3-4).**

406

407 **Table 2. Recoveries (%) of iron(II) measured using the optimized colorimetric DET technique at various**
408 **ionic strengths. Values are means \pm 1 standard deviation (n=3).**

409

410

411

412

413

414

415

416

417

418

419

420

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435 **Figure Captions**

436

437 **Figure 1. Calibration curve for colorimetric iron(II) DET technique with 0.01 mol L⁻¹ ferrozine, based on**
438 **grayscale conversion using a green (▲), blue (■) or red (●) color filter. Error bars represent 1 standard**
439 **deviation of the mean (n=3).**

440

441 **Figure 2. Split calibration curve for colorimetric iron(II) DET technique showing linear relationship for**
442 **lower concentrations (bottom) and logarithmic relationship for higher concentrations (top). Converted to**
443 **grayscale with red color channel filter. Error bars represent 1 standard deviation of the mean (n=3).**
444 **Staining time was 10 minutes.**

445

446 **Figure 3. Calibration curve for colorimetric iron(II) DET technique based on grayscale conversion with**
447 **green color channel filter. Error bars represent 1 standard deviation of the mean (n=3).**

448

449 **Figure 4. One-dimensional depth profile and two-dimensional distribution of porewater iron(II)**
450 **concentration in a homogenized freshwater sediment mesocosm at 1 mm resolution. Values of the one-**
451 **dimensional depth profile are means of 13 individual horizontal measurements at 1 mm resolution and**
452 **error bars represent plus or minus one standard deviation.**

453