



UNIVERSITY  
OF WOLLONGONG  
AUSTRALIA

University of Wollongong  
Research Online

---

Faculty of Science, Medicine and Health - Papers

Faculty of Science, Medicine and Health

---

2013

# Accuracy of micrometeorological techniques for detecting a change in methane emissions from a herd of cattle

Johannes Laubach  
*Landcare Research NZ*

Mei Bai  
*University of Wollongong, mb81@uow.edu.au*

Cesar S. Pinares-Patino  
*Agresearch Grasslands Research Centre*

Frances A. Phillips  
*University of Wollongong, francesp@uow.edu.au*

Travis A. Naylor  
*University of Wollongong, naylor@uow.edu.au*

*See next page for additional authors*

---

## Publication Details

Laubach, J., Bai, M., Pinares-Patino, C. S., Phillips, F. A., Naylor, T. A., Molano, G., Cardenas Rocha, E. A. & Griffith, D. W. T. (2013). Accuracy of micrometeorological techniques for detecting a change in methane emissions from a herd of cattle. *Agricultural and Forest Meteorology*, 176 50-63.

Research Online is the open access institutional repository for the University of Wollongong. For further information contact the UOW Library:  
research-pubs@uow.edu.au

---

# Accuracy of micrometeorological techniques for detecting a change in methane emissions from a herd of cattle

## Abstract

Micrometeorological techniques are effective in measuring methane (CH<sub>4</sub>) emission rates at the herd scale, but their suitability as verification tools for emissions mitigation depends on the uncertainty with which they can detect a treatment difference. An experiment was designed to test for a range of techniques whether they could detect a change in weekly mean emission rate from a herd of cattle, in response to a controlled change in feed supply. The cattle were kept in an enclosure and fed pasture baleage, of amounts increasing from one week to the next. Methane emission rates were measured at the herd scale by the following techniques: (1) an external tracer-ratio technique, releasing nitrous oxide (N<sub>2</sub>O) from canisters on the animals' necks and measuring line-averaged CH<sub>4</sub> and N<sub>2</sub>O mole fractions with Fourier-transform infra-red (FTIR) spectrometers deployed upwind and downwind of the cattle, (2) a mass-budget technique using vertical profiles of wind speed and CH<sub>4</sub> mole fraction, (3) a dispersion model, applied separately to CH<sub>4</sub> mole fraction data from the FTIR spectrometers, the vertical profile, and a laser system measuring along four paths surrounding the enclosure. For reference, enteric CH<sub>4</sub> emissions were also measured at the animal scale on a daily basis, using an enteric tracer-ratio technique (with SF<sub>6</sub> as the tracer). The animal-scale technique showed that mean CH<sub>4</sub> emissions increased less than linearly with increasing feed intake. The herd-scale techniques showed that the emission rates followed a diurnal pattern, with the maximum about 2 h after the feed was offered. The herd-scale techniques could detect the weekly changes in emission levels, except that the two vertical-profile techniques (mass-budget technique and dispersion model applied to profile) failed to resolve the first step change. The weekly emission rates from the external tracer-ratio technique and the dispersion model, applied to data from either the two FTIR paths or the four laser paths, agreed within  $\pm 10\%$  with the enteric tracer-ratio technique. By contrast, the two vertical-profile techniques gave 33–68% higher weekly emission rates. It is shown with a sensitivity study that systematically uneven animal distribution within the enclosure could explain some of this discrepancy. Another cause for bias was the data yield of the vertical-profile techniques being higher at day-time than at night-time, thus giving more weight to times of larger emission rates. The techniques using line-averaged mole fractions were less sensitive to the exact locations of emission sources and less prone to data loss from unsuitable wind directions; these advantages outweighed the lack of a method to calibrate CH<sub>4</sub> mole fractions in situ.

## Keywords

herd, emissions, methane, cattle, change, accuracy, detecting, techniques, micrometeorological, GeoQuest

## Disciplines

Medicine and Health Sciences | Social and Behavioral Sciences

## Publication Details

Laubach, J., Bai, M., Pinares-Patino, C. S., Phillips, F. A., Naylor, T. A., Molano, G., Cardenas Rocha, E. A. & Griffith, D. W. T. (2013). Accuracy of micrometeorological techniques for detecting a change in methane emissions from a herd of cattle. *Agricultural and Forest Meteorology*, 176 50-63.

---

**Authors**

Johannes Laubach, Mei Bai, Cesar S. Pinares-Patino, Frances A. Phillips, Travis A. Naylor, German Molano, Edgar A. Cardenas Rocha, and David W. T Griffith

# Accuracy of micrometeorological techniques for detecting a change in methane emissions from a herd of cattle

Johannes Laubach<sup>A,\*</sup>, Mei Bai<sup>B,\*\*</sup>, Cesar S. Pinares-Patiño<sup>C</sup>, Frances A. Phillips<sup>B</sup>, Travis A. Naylor<sup>B</sup>, German Molano<sup>C</sup>, Edgar A. Cárdenas Rocha<sup>C,D</sup>, David W. T. Griffith<sup>B</sup>

<sup>A</sup>*Landcare Research, P.O. Box 40, Lincoln 7640, New Zealand*

<sup>B</sup>*Centre for Atmospheric Chemistry, University of Wollongong, Wollongong NSW 2522, Australia*

<sup>C</sup>*AgResearch, Grasslands Research Centre, Tennent Drive, Palmerston North 4442, New Zealand*

<sup>D</sup>*Facultad de Medicina Veterinaria y Zootecnia, Universidad Nacional de Colombia, Carrera 45 No. 26-85, Bogotá, Colombia*

\*Corresponding author:

Tel.: +64-3-321 9999

Fax: +64-3-321 9998

email: LaubachJ@LandcareResearch.co.nz

\*\*present address: CSIRO Animal, Food & Health Sciences, Townsville QLD 4811, Australia

*Submitted to Agricultural and Forest Meteorology 15 November 2012*

*Revised version, submitted 23 February 2013*

## 1 **Abstract**

2 Micrometeorological techniques are effective in measuring methane (CH<sub>4</sub>) emission  
3 rates at the herd scale, but their suitability as verification tools for emissions mitigation  
4 depends on the uncertainty with which they can detect a treatment difference. An  
5 experiment was designed to test for a range of techniques whether they could detect a  
6 change in weekly mean emission rate from a herd of cattle, in response to a controlled  
7 change in feed supply. The cattle were kept in an enclosure and fed pasture baleage, of  
8 amounts increasing from one week to the next. Methane emission rates were measured  
9 at the herd scale by the following techniques: 1) an external tracer-ratio technique,  
10 releasing nitrous oxide (N<sub>2</sub>O) from canisters on the animals' necks and measuring line-  
11 averaged CH<sub>4</sub> and N<sub>2</sub>O mole fractions with Fourier-transform infra-red (FTIR)  
12 spectrometers deployed upwind and downwind of the cattle, 2) a mass-budget  
13 technique using vertical profiles of wind speed and CH<sub>4</sub> mole fraction, 3) a dispersion  
14 model, applied separately to CH<sub>4</sub> mole fraction data from the FTIR spectrometers, the  
15 vertical profile, and a laser system measuring along four paths surrounding the  
16 enclosure. For reference, enteric CH<sub>4</sub> emissions were also measured at the animal scale  
17 on a daily basis, using an enteric tracer-ratio technique (with SF<sub>6</sub> as the tracer). The  
18 animal-scale technique showed that mean CH<sub>4</sub> emissions increased less than linearly  
19 with increasing feed intake. The herd-scale techniques showed that the emission rates  
20 followed a diurnal pattern, with the maximum about two hours after the feed was  
21 offered. The herd-scale techniques could detect the weekly changes in emission levels,  
22 except that the two vertical-profile techniques (mass-budget technique and dispersion  
23 model applied to profile) failed to resolve the first step change. The weekly emission  
24 rates from the external tracer-ratio technique and the dispersion model, applied to data  
25 from either the two FTIR paths or the four laser paths, agreed within ±10 % with the  
26 enteric tracer-ratio technique. By contrast, the two vertical-profile techniques gave 33  
27 to 68 % higher weekly emission rates. It is shown with a sensitivity study that  
28 systematically uneven animal distribution within the enclosure could explain some of  
29 this discrepancy. Another cause for bias was the data yield of the vertical-profile  
30 techniques being higher at day-time than at night-time, thus giving more weight to  
31 times of larger emission rates. The techniques using line-averaged mole fractions were  
32 less sensitive to the exact locations of emission sources and less prone to data loss  
33 from unsuitable wind directions; these advantages outweighed the lack of a method to  
34 calibrate CH<sub>4</sub> mole fractions in situ.

35

36 **Keywords:** Cattle CH<sub>4</sub> emissions, Gas dispersion, Atmospheric surface layer, Tracer-ratio  
37 techniques, Mass-budget technique, Backward-Lagrangian stochastic model

38

## 39 **1. Introduction**

40 Methane (CH<sub>4</sub>) emissions from ruminant livestock constitute about 30 % of New  
41 Zealand's and 12 % of Australia's greenhouse gas emissions. For any future practice or  
42 technology to mitigate these emissions, it must be verified at the “herd scale” (or  
43 “paddock scale”), under representative farming conditions, that the expected emissions  
44 reduction is achieved. In New Zealand (NZ) and Australia, cattle and sheep are farmed  
45 outdoors year-round. To measure CH<sub>4</sub> emissions outdoors, micrometeorological  
46 techniques are effective at the herd scale (Laubach et al., 2008). These can potentially  
47 verify small changes in emission rates, provided that the uncertainty with which such  
48 changes are detected is accurately known. Here, we report an experiment measuring  
49 the emissions from a herd of beef cattle that was designed to specify this uncertainty  
50 for a suite of micrometeorological techniques. These include: a mass-budget technique  
51 using vertical profiles of wind speed and CH<sub>4</sub> mole fraction ([CH<sub>4</sub>]), a backward-  
52 Lagrangian stochastic (BLS) dispersion model using the same [CH<sub>4</sub>] profiles, the same  
53 BLS model using line-averaged [CH<sub>4</sub>] data gathered with two types of instruments,  
54 and an external tracer-ratio technique, releasing nitrous oxide (N<sub>2</sub>O) co-located with  
55 the CH<sub>4</sub> emission sources and measuring line-averaged N<sub>2</sub>O and CH<sub>4</sub> mole fractions  
56 upwind and downwind of the sources. The last technique was classified by Harper et  
57 al. (2011) as “non-micrometeorological”, since it does not require any meteorological  
58 information to compute emission rates; however, as its feasibility relies on atmospheric  
59 transport, we consider it more appropriate to include it among the  
60 “micrometeorological” techniques. With these, it also shares the spatial and temporal  
61 scales at which it operates. To obtain non-micrometeorological reference values of  
62 CH<sub>4</sub> emissions on a daily basis, an enteric tracer-ratio technique was employed,  
63 commonly known as the SF<sub>6</sub> tracer-ratio technique. This technique operates at the  
64 “animal scale”, i.e. individual animals. Details and references for all techniques are  
65 given in later sections.

66 Laubach et al. (2008) already reported an experiment at the same site, with  
67 similar animals, and using the same techniques except for the external tracer-ratio  
68 technique. In that experiment, the animals were freely grazing in rectangular strips that  
69 were changed daily. This represented usual farming practice in NZ but had two

70 disadvantages: the instrumentation had to be moved frequently to be in suitable  
71 locations for capturing the emissions, and the feed intake of the animals could not be  
72 measured. Knowing the feed intake is desirable because it is a major factor  
73 determining CH<sub>4</sub> emissions. The “methane yield”  $Y_m$ , defined as the ratio of CH<sub>4</sub>  
74 emissions per dry-matter intake (DMI) where both variables are expressed in units of  
75 combustion energy, is recommended for inventory purposes (IPCC, 2006; Lassey,  
76 2007). To overcome the two disadvantages in the present experiment, the cattle were  
77 held in a grass-free area and were fed known quantities of pasture baleage. The  
78 experiment was run for three weeks. Within each week, daily feed rations were held  
79 constant, then increased for the following week in order to produce a measurable step  
80 change in the herd’s CH<sub>4</sub> emissions. The objective was to test for each technique  
81 whether it was possible to detect this step change on the basis of weekly averages,  
82 which in turn required quantification of the uncertainty of these averages. Factors  
83 determining this uncertainty are not only measurement accuracy, but also for each  
84 technique its data yield, i.e. the number of runs meeting specific quality criteria, and its  
85 sensitivity to the spatial distribution of sources (which is always assumed  
86 homogeneous across a prescribed area, except for the external tracer-ratio technique,  
87 where no such assumption is needed).

88 A side issue, inadvertently discovered from consistency checks between the  
89 different CH<sub>4</sub> instruments, is a strong temperature dependence of the “GasFinder” CH<sub>4</sub>  
90 laser, previously unreported in the micrometeorological literature. This result is  
91 presented and empirically corrected for in the main text; the causes are discussed in the  
92 Appendix.

93

94

## 95 **2. Experimental design**

96 The experiment was conducted in November 2008. The site (40.336° S, 175.465° E)  
97 was located on the Aorangi Research Farm, ca. 20 km inland from the west coast of  
98 the North Island of NZ, near the city of Palmerston North. It is ideally suited for  
99 micrometeorological techniques and tracer dispersion studies because the surrounding  
100 terrain is flat for several km in all directions and there is a predominant wind direction,  
101 from W (which includes frequent afternoon sea breezes when synoptic winds are  
102 weak). The cattle were managed in two groups, of 30 and 31 animals, respectively,  
103 with identical treatments as described in 2.2.

104

## 105 **2.1 Site preparation and setup geometry**

106 A paddock area of about 200 m by 200 m had been sprayed with herbicide prior to the  
107 experiment, so that the ground was initially bare; by the end, a thin cover of herbage  
108 had grown back. In the NW quarter of this area, a rectangle of 80 m by 55 m was  
109 fenced to contain the cattle herd. A 7 m tall mast to collect vertical profiles of wind  
110 speed and CH<sub>4</sub> mole fraction was erected at the midpoint of the E boundary of the  
111 fenced area, and with additional fencing of a semicircle with 20 m radius, the cattle  
112 were kept at a minimum distance of 20 m from the mast. The nominal source area  
113 (rectangle minus semicircle) thus covered 3772 m<sup>2</sup>. This area was subdivided into two  
114 equal-size enclosures (Fig. 1) to allow handling of the cattle in two separate groups and  
115 to limit clustering of the animals at feeding time.

116 The surrounding terrain consisted mainly of flat paddocks, with no significant  
117 flow obstacles to the W, S and E for at least 500 m. To the N, there was a water ditch  
118 dropping 2 m below ground level at 50 m distance from the profile mast, and a  
119 shelterbelt at ca. 150 m distance. The bare ground extended ca. 100 m from the mast to  
120 the W, S and E, and 45 m to the N.

121 Two types of line-averaging optical sensors (described below) were employed  
122 (Fig. 1). One was a CH<sub>4</sub>-specific laser system with four paths that were mounted  
123 outside the cattle area, with one path along each side of the rectangle, at 5 to 10 m  
124 distance from the fence. Path lengths were between 53 and 57 m, measured with 0.1 m  
125 accuracy, and path heights above ground were 1.85, 1.37, 1.38 and 1.07 m ( $\pm 0.02$ ) m  
126 for the W, S, E and N path, respectively. The other instrument type was a Fourier-  
127 transform infra-red (FTIR) spectrometer, measuring mole fractions of multiple gas  
128 species simultaneously along an open path. Two identical FTIRs were set up parallel to  
129 the W and E sides of the rectangle, at about 5 and 12 m distance to W and E,  
130 respectively. The path lengths were 88.9 and 89.5 m and the heights above ground  
131 were 1.39 ( $\pm 0.05$ ) m.

132 Wind direction, atmospheric stability and velocity statistics (required as inputs  
133 for the dispersion model) were measured with a sonic anemometer (81000V, RM  
134 Young, Traverse City, Michigan, USA), mounted on top of a telescope mast, 3.85 m  
135 above ground and 13 m NNW of the profile mast.

136

## 137 2.2 Animals and feed

138 The experiment involved 61 one-year-old Friesian-Hereford crossbred steers. They  
139 were fed perennial ryegrass/white clover baleage that had been prepared on-farm, with  
140 an approximate content of white clover of 20 % (DM basis). The chemical  
141 composition of the feed was determined by near-infrared spectroscopy.

142 Prior to the start of measurements, the animals were acclimatised to the  
143 management conditions over a period of 10 d by gradually increasing the baleage  
144 fraction in the diet while decreasing herbage allowance. At the same time, the animals  
145 were accustomed to wear the gas collection gear required for the SF<sub>6</sub> tracer-ratio  
146 technique. Starting on the 6th day of acclimatisation, baleage constituted 100 % of the  
147 diet. The cattle were fed at three increasing feeding levels (low, medium and high)  
148 over three consecutive periods of one week duration each. The first three days of each  
149 week were considered as adjustment periods to the new feeding level, and the last four  
150 days were used to conduct CH<sub>4</sub> emissions measurements with the animal-scale  
151 technique. Feeding levels were set with the intention of making 1.0, 1.5, and 2.0 times  
152 maintenance energy requirements available to the steers during Weeks 1, 2 and 3,  
153 respectively. The maintenance energy requirement was estimated following CSIRO  
154 (2007). In order to allow for increasing fractions of feed wastage with increasing  
155 feeding level, the feed amounts offered were set to 1.10, 1.73 and 2.60 times  
156 maintenance requirements. The animals were weighed at the start of the experiment  
157 and at the end of each one-week feeding period. Drinking water was freely available to  
158 them, from troughs placed centrally in each enclosure.

159 The feed bales weighed around 450 kg. Every morning at 09:00 h the feed was  
160 distributed on the ground inside the two enclosures, using a forage mixer feed-out  
161 wagon equipped with a scale. The accuracy of the wagon's scale was calibrated against  
162 a load cell at the start of the experiment and found to agree within ±10 kg for a bale. In  
163 each enclosure, the feed was laid out approximately in a "B" shape, which maximised  
164 the areal spread of the feed while avoiding that the feed-out wagon had to cross the  
165 supply pipes to the water troughs (Fig. 1).

166 Estimation of feed intake on the basis of metabolisable energy algorithms was  
167 not considered reliable, because the 7 d long feeding period at each level was too short  
168 to accurately account for liveweight change due to the potentially overwhelming  
169 influence of variable gut fill when the animals were weighed. Alternatively, feed  
170 intake estimates were to be based on the difference between feed offered and refused.  
171 However, refusals were trampled and mixed with faeces and soil, and attempts to

172 collect representative samples proved impractical. Consequently, the proportion of  
173 refused feed was estimated daily by visual assessment.

174

### 175 **2.3 Profile mast instrumentation**

176 The vertical profile of [CH<sub>4</sub>] was measured with an analyser using off-axis integrated  
177 cavity output spectroscopy (DLT-100, Los Gatos Research, Mountain View,  
178 California, USA). Air was drawn continuously from seven intakes via tubes into  
179 ballast volumes. A datalogger-controlled valve-switching system sequentially selected  
180 each of the intake lines for sampling by the analyser. Each switching cycle lasted  
181 20 min, allocating 171 s of sampling time to each intake, of which the first 140 s were  
182 discarded as flushing time to purge the sample cell of air from the previous intake.  
183 Five of the intakes were mounted on the mast, at heights of 0.62, 1.22, 2.24, 4.13 and  
184 7.18 m ( $\pm 0.02$ ) m. The other two were placed to the W and SE away from the cattle  
185 area (Fig. 1), at 1.94 and 1.77 m height, respectively, so that for any wind direction one  
186 of them would be suitably located to provide the local background mole fraction  
187 (upwind of the herd).

188 Wind speed, relative humidity (RH) and temperature were measured at the same  
189 five heights as [CH<sub>4</sub>]: wind speed by cup anemometers (A101M, Vector Instruments,  
190 Rhyl, Clwyd, UK) with matched calibrations, humidity by capacitive RH sensors  
191 (MP100A, Rotronic, Bassersdorf, Switzerland) and temperature by thermocouples  
192 inside the same aspirated double radiation shields as the RH sensors.

193

### 194 **2.4 Open-path instrumentation**

195 The four-path CH<sub>4</sub> laser system (GasFinder MC, Boreal Laser, Edmonton, Alberta,  
196 Canada) consists of a central control unit, four remote heads (Fig. 2a) and four retro-  
197 reflectors. The central unit houses the laser source, of wavelength 1653 nm, a sealed  
198 reference cell filled with a known amount of CH<sub>4</sub>, an optical multiplexer and the  
199 controlling and processing electronics. The wavelength is modulated across a narrow  
200 band in order to sample the shape of the absorption line. The IR light is ducted via  
201 fibre-optical cables to the remote heads, switching between them in a cycle of ca. 18 s  
202 duration. The head emits the light along the open path towards the retro-reflector and  
203 collects the returning light on a photodiode. The electrical output signal of the  
204 photodiode is transmitted back to the central unit via a coaxial cable. There, the light

205 intensity measurements, as scanned across the waveband, are converted to CH<sub>4</sub> mole  
206 fractions with a regression algorithm that compares the measured line shape to that of  
207 the reference cell (Boreal Laser, 2005). These mole fraction data were collected on a  
208 laptop computer and subjected to post-processing to remove dubious data (guided by  
209 quality flags and light level data provided by the instrument), apply correction factors  
210 for signal dampening along the coaxial cable length, and form 20-min averages aligned  
211 with the switching cycles of the vertical profile system.

212 The two identical open-path multi-gas FTIR spectrometer systems were  
213 constructed at the University of Wollongong (Bai, 2010). The spectrometer is a  
214 Matrix-M IRcube (Bruker Optik, Ettlingen, Germany). Light from an air-cooled SiC  
215 infrared source is modulated by the interferometer and passes via a beam splitter (ZnSe  
216 window) into a modified 10" Schmidt-Cassegrain telescope (LX200R, Meade  
217 Instrument Corp., Irvine, California, USA). From there, the primary beam is  
218 transmitted across the measurement path, returned by a 0.3 m diameter retro-reflector  
219 array (PLX, Deer Park, New York, USA) and received by a HgCdTe detector (Infrared  
220 Associates, Stuart, Florida, USA) which sits on the same tripod-mounted optical bench  
221 as the source, interferometer and telescope (Fig. 2b). A Zener-diode thermometer (type  
222 LM335) and a barometer (PTB110, Vaisala, Helsinki, Finland) are integrated in the  
223 source/detector array to provide real-time air temperature and pressure data for the  
224 analysis of the measured spectra. The spectrometer is controlled by software developed  
225 at the University of Wollongong. The interferometer performs about 80 scans min<sup>-1</sup>  
226 across the waveband of interest (700 – 5000 cm<sup>-1</sup>), which were in this experiment  
227 averaged to 2-min spectra. From the spectra, the mole fractions of CH<sub>4</sub>, N<sub>2</sub>O, CO<sub>2</sub>, CO  
228 and H<sub>2</sub>O are retrieved by a non-linear optimisation algorithm (MALT software),  
229 developed by Griffith (1996). The algorithm uses absorption line strengths of these  
230 gases as listed in the HITRAN database (Rothman et al., 2005) and constructs the  
231 combination of mole fractions that provides the best match between the transmission  
232 spectrum expected from this combination and the observed spectrum (Smith et al.,  
233 2011; Griffith et al., 2012). Only the CH<sub>4</sub> and N<sub>2</sub>O mole fractions are used in this  
234 experiment.

235

236

### 237 **3. Techniques to determine CH<sub>4</sub> emission rates**

238 The various techniques to determine the CH<sub>4</sub> emission rates are listed in Table 1. The  
239 first technique (Section 3.1) was applied to individual animals, providing daily  
240 averages. The others (Sections 3.2 to 3.4) all integrated over emissions from the herd,  
241 as the emitted CH<sub>4</sub> was transported and dispersed by the wind. They were applied  
242 using 20-min averages of the CH<sub>4</sub> mole fraction, relative to moist air, and of the  
243 meteorological variables.

244

#### 245 **3.1 Enteric tracer-ratio technique (SF<sub>6</sub>)**

246 The SF<sub>6</sub> tracer-ratio technique (Johnson et al., 1994) was employed to estimate CH<sub>4</sub>  
247 emissions from individual animals. For this purpose, on the first day of the pre-trial  
248 acclimatisation period of 10 d, individually calibrated brass permeation tubes  
249 containing the SF<sub>6</sub> tracer were dosed *per os* into the reticulo-rumen of each  
250 participating animal. For the last four days of each feeding level period, daily breath  
251 and background air samples were collected from each steer, using a yoke fitted around  
252 his neck (Fig. 3). Although the animals were out of the paddock for 2 to 3 h every  
253 morning (time spent in transit to and from the management yards and for handling),  
254 changing the sampling yoke took only about 1 min per animal, thus the collection  
255 period was effectively 24 h. Samples were transported to the laboratory for analysis of  
256 SF<sub>6</sub> and CH<sub>4</sub> mole fractions using a gas chromatograph (GC-2010, Shimadzu, Kyoto,  
257 Japan), as described by Pinares-Patiño et al. (2007). Daily CH<sub>4</sub> emissions from  
258 individual animals were calculated from the ratio of mole fraction elevations of CH<sub>4</sub>  
259 and SF<sub>6</sub> (above the respective background-air mole fractions) and the known  
260 permeation rate of SF<sub>6</sub> from the permeation tubes (Johnson et al., 1994).

261

#### 262 **3.2 External tracer-ratio technique (N<sub>2</sub>O)**

263 The basic approach of the N<sub>2</sub>O tracer-ratio technique is the same as for the SF<sub>6</sub>  
264 technique: a tracer gas is released at a known rate co-located with the CH<sub>4</sub> emission  
265 from the animals, the mole fractions of both gases are measured post-emission in the  
266 same volume, and then the ratio of the two mole fractions (after subtracting  
267 atmospheric background from each) is equated with the ratio of their emission rates.  
268 The main difference is that with the N<sub>2</sub>O technique the gases are not collected in a  
269 container, but measured in situ, some distance downwind, by the open-path FTIR. The  
270 path length is long enough to capture the emissions from many, if not all, animals

271 simultaneously, which makes the technique effective at the herd scale. Along with this  
272 spatial integration comes high temporal resolution, not achievable with the animal-  
273 scale technique.

274 Nitrous oxide was chosen because it is inert, non-toxic, commercially available  
275 and its spectrum documented in HITRAN, which makes it easily measurable by the  
276 FTIR in the same wavebands as CH<sub>4</sub>. Also, it could be released at rates much larger  
277 than from any environmental sources (soil and animal excreta), so that the measured  
278 downwind-upwind [N<sub>2</sub>O] differences were clearly attributable to the manufactured  
279 release. In a previous experiment, Griffith et al. (2008) released the N<sub>2</sub>O from a  
280 perforated pipe along the upwind fence line. Here, the release location was much more  
281 closely matched to that of the CH<sub>4</sub> by using pressurised canisters carried by the  
282 animals near their mouths (Fig 3). These were 12 oz paintball gas canisters (Catalina  
283 cylinders, Garden Grove, California, USA), which were filled with ca. 0.3 kg liquid  
284 compressed N<sub>2</sub>O and fitted to the animals on a daily schedule, along with the SF<sub>6</sub>  
285 collection yokes to which they were attached. Due to the long preparation time  
286 required to fill the N<sub>2</sub>O canisters, only half of the cattle (15 in each group) were  
287 equipped with them. This was unlikely to introduce additional error because animals  
288 with and without release canister distributed themselves randomly across the same  
289 area. Each canister was fitted with a tap and restricting orifice to regulate the release  
290 rate to about 10 g h<sup>-1</sup>. The release rate's time average was determined for each canister  
291 by weighing it, before and after it was carried by the animal, and the release rate's  
292 temporal evolution was constructed by taking into account its systematic decrease with  
293 canister pressure as well as its temperature dependence. Technical details of the  
294 canister filling procedure and the computation of the N<sub>2</sub>O release rate are given in Bai  
295 (2010). The CH<sub>4</sub> emission rate was then calculated from the total N<sub>2</sub>O release rate and  
296 the path-integrated CH<sub>4</sub> and N<sub>2</sub>O mole fractions measured downwind from the herd.

297

### 298 **3.3 Micrometeorological mass-budget technique**

299 The mass-budget technique, sometimes also named integrated horizontal-flux  
300 technique, was first applied to animal emissions by Harper et al. (1999) and Leuning et  
301 al. (1999). Here, it was implemented using the vertical 5-point profiles of CH<sub>4</sub> mole  
302 fraction, measured with the Los Gatos analyser, and of wind speed, measured with the  
303 cup anemometers. From these profiles and the background mole fraction, measured  
304 upwind of the cattle, the integrated horizontal flux of CH<sub>4</sub> was computed. To convert

305 this flux to an areal emission rate, it was divided by the distance across the cattle area,  
306 from the profile mast in the upwind direction. Corrections for cross-wind variation of  
307 this distance, for horizontal flux contributions above the top measurement height, and  
308 for turbulent backflow were applied as described in Laubach and Kelliher (2004).

309

### 310 **3.4 BLS technique**

311 The principle of the backwards-Lagrangian stochastic (BLS) technique is to employ a  
312 dispersion model to simulate trajectories of air parcels backwards in time from a  
313 location where a gas concentration was measured, and to analyse statistically where  
314 these trajectories intersect with a source volume (or area) in which the concentration of  
315 the simulated air parcels would have been altered. For a given flow field (mean wind,  
316 direction, and certain turbulent parameters), this trajectory-mapping delivers an  
317 unambiguous linear relationship between measured concentration and emission rate.

318 The BLS model used here, developed by Flesch et al. (1995; 2004), is distributed  
319 as a user-friendly software (WindTrax 2.0, [www.thunderbeachscientific.com](http://www.thunderbeachscientific.com)). The  
320 model can accommodate concentration measurements at a point, such as realised by an  
321 intake to a gas analyser, or along a line, such as the path of an optical sensor measuring  
322 absorption. It is here applied separately to the three types of CH<sub>4</sub>-measuring  
323 instruments: the closed-path analyser (vertical profile), the open-path laser system  
324 (4 paths) and the open-path FTIR instruments (2 paths). While the three simulations  
325 differ in their concentration input data, they assume identical CH<sub>4</sub> source distributions  
326 and identical flow fields. The CH<sub>4</sub> source is defined as a ground-level area source with  
327 the dimensions of the cattle enclosure. The flow field is specified, as in Laubach  
328 (2010), by wind speed from the highest cup anemometer, the stability parameter and  
329 the standard deviations of the three wind components from the sonic anemometer (at  
330 3.85 m), and roughness length. Roughness length was not allowed to vary randomly  
331 from run to run; rather, its evolution for the whole experiment was determined as a  
332 function of time and wind direction. Over time, the roughness length increased slowly  
333 as vegetation grew back, from 0.8 cm (bare soil) to 1.7 cm. The directional analysis  
334 showed systematically larger values from the N sector than from the other directions,  
335 consistent with the presence of ditch and shelterbelt to the N. The directional  
336 dependence was fitted with a squared cosine function for directions within  $\pm 90^\circ$  from  
337 N, added to the time-dependent roughness length for the other sectors. The maximum  
338 roughness length obtained for N winds was 9.1 cm.

339

340

## 341 **4. Calibration checks and corrections to CH<sub>4</sub> mole fractions**

### 342 **4.1 Closed-path analyser**

343 Every 6 h, sampling of vertical [CH<sub>4</sub>] profiles with the closed-path analyser (Los  
344 Gatos) was interrupted for 20 min to perform an automated calibration check, using  
345 two gas cylinders that provided air with near-zero and near-ambient CH<sub>4</sub> mole  
346 fractions. These checks showed that throughout the campaign, [CH<sub>4</sub>] was reproducible  
347 within ±3 ppb standard deviation (±0.18 % of ambient). Testing for temperature  
348 dependence, a linear regression slope of 0.07 ppb K<sup>-1</sup> was found, which is a factor 12  
349 smaller than reported by Tuzson et al. (2010) for the same type of instrument. With  
350 R<sup>2</sup> = 0.03, this temperature dependence was not significantly different from zero and  
351 thus neglected. Instead, the observed differences between subsequent calibration  
352 checks were used to correct the mole fraction data in post-processing, assuming a  
353 linear drift in time. Within a 20-min run, this drift correction had negligible effect on  
354 mole fraction differences between intakes, but it provided minor adjustments to the  
355 diurnal courses.

356

### 357 **4.2 Open-path FTIR instruments**

358 For the open-path instruments, in-situ calibration checks were not possible. Instead, a  
359 series of consistency checks was conducted, first between paths of the same system,  
360 then between these and the closed-path instrument. For these checks, only the periods  
361 of cattle absence were used, expecting that then all instruments would effectively  
362 sample the local background CH<sub>4</sub> mole fraction – not necessarily identical to  
363 hemispheric background, but influenced only by sources (or sinks) far enough away  
364 that the position differences between the compared instruments did not matter.

365 As the periods of cattle absence were in the morning, strongly stable  
366 stratification from the residual nocturnal boundary could occur on some days, with  
367 elevated CH<sub>4</sub> concentrations that potentially showed a vertical structure. To be able to  
368 identify the influence of stable versus well-mixed conditions, the [CH<sub>4</sub>] data from the  
369 two FTIR instruments are compared by plotting their ratio as a function of the wind  
370 speed at 7.18 m (Fig. 4). As a change in alignment can change the instrument  
371 calibration, the data are split into two groups, before and after a realignment of the W

372 path that was necessary following disturbance to the instrument while diagnosing an  
373 electronic failure. Before the realignment, the ratio was consistently near 1, except for  
374 some runs of low wind speed which are considered as stably stratified and not well-  
375 mixed. The mean ( $\pm$ SD) for wind speed  $> 2 \text{ m s}^{-1}$  was  $1.007 (\pm 0.009, n = 67)$ . After  
376 the realignment, the ratio was  $1.034 (\pm 0.006, n = 30)$ . The SD values indicate a random  
377 error of order 5 to 10 ppb at ambient  $[\text{CH}_4]$ . This estimate would include an error  
378 contribution from the horizontal distance between the instruments; it is thus  
379 compatible with several independent instrument precision checks that yielded  
380 estimates between 2 and 4 ppb (Bai, 2010).

381 Similar tests were done for the  $\text{N}_2\text{O}$  mole fractions, required for the external  
382 tracer-ratio technique (not shown). These yielded W/E ratios of 1.02 ( $\pm 0.01$ ) before  
383 and 1.03 ( $\pm 0.01$ ) after realignment, respectively. The SD values indicate a random  
384 error of 2 ppb at ambient  $[\text{N}_2\text{O}]$ , which again – due to the error from the horizontal  
385 separation – is compatible with independent precision checks, yielding 0.4 ppb (Bai,  
386 2010).

387 To compare the two FTIR to the closed-path analyser,  $[\text{CH}_4]$  from each of the  
388 former was divided by  $[\text{CH}_4]$  from the nearest intake of the latter, selecting the same  
389 cattle-free periods as before. The nearest path for the W path FTIR was the background  
390 intake at 1.94 m height, and for the E path, the second-lowest intake from the profile  
391 mast, at 1.22 m height. Since the open-path and closed-path data differ in three  
392 respects: height, averaging volume (“line” vs. “point”), and sampling duration (shorter  
393 for closed-path due to the switching cycle), one should not expect perfect agreement.  
394 However the ratios showed remarkable consistency, at about 0.97 for the E path and  
395 about 0.98 and 1.01 for the W path before and after the realignment, respectively  
396 ( $\pm 0.01$  for each ratio). The ratios were independent of temperature and humidity. Any  
397 ratios less than 0.94 were associated with wind speed  $< 2 \text{ m s}^{-1}$ . This indicates that the  
398 absolute mole fractions of the two FTIR instruments were 2 to 3 % too low, equivalent  
399 to an absolute error of 40 to 60 ppb at background, and they did not drift.

400 After these checks, the mole fractions from the W path were corrected against  
401 those from the E path, by factors determined as the slopes of period-wise linear  
402 regressions (of runs without cattle presence). This ensured matched calibrations  
403 between the two FTIRs, which are crucial for accurate determination of upwind-  
404 downwind differences, and in turn, emission rates. The FTIR data were not corrected  
405 against the closed-path data. Thus, it is expected that the 3 % difference between the E

406 FTIR and the closed-path analyser carries through to the emission rate computations  
407 with the BLS technique.

408

### 409 **4.3 Four-path laser system**

410 For the open-path laser (GasFinder MC), the same consistency checks were carried out  
411 as for the FTIR. In Fig. 5, consistency between paths is assessed, by showing the ratios  
412 of [CH<sub>4</sub>] from the W, S and N path, respectively, to [CH<sub>4</sub>] from the E path, against  
413 wind speed. For each ratio, the spread is considerably larger than for the W/E FTIR  
414 ratio in Fig. 4, and wind speed does not affect the spread, indicating that the variability  
415 reflects instrument precision, not true [CH<sub>4</sub>] variations. The means ( $\pm$ SD), for 126 runs  
416 of cattle absence, were 1.015 ( $\pm$ 0.039) for W/E, 0.963 ( $\pm$ 0.038) for S/E, and  
417 1.033 ( $\pm$ 0.041) for N/E, respectively. The SD values indicate a random error of order  
418 48 ppb, at ambient [CH<sub>4</sub>], for each path. This is compatible with the manufacturer's  
419 specification of 2 ppm m precision for the path-integrated mole fraction, equivalent to  
420 36 ppb for a 55 m long path. The laser system is thus one magnitude less precise than  
421 the FTIR instruments and the closed-path analyser.

422 In Fig. 5, the differences of the mean ratios from 1 cannot be explained by height  
423 differences (the W path was the highest, N the lowest, and S and E were at equal  
424 height). Like for the FTIR, the E path was selected as the reference path, and the [CH<sub>4</sub>]  
425 data from the other three paths were corrected against it. This was done separately for  
426 the three feeding-level periods, to ensure the [CH<sub>4</sub>] differences between paths were  
427 bias-free within each such period.

428 As for the FTIR, mole fractions from the W and E paths of the laser system were  
429 compared to those from the nearest intakes of the closed-path analyser. Both horizontal  
430 and vertical distances between path and intake location were smaller than for the FTIR.  
431 Therefore, it was considered acceptable to include runs with and without cattle  
432 presence in the comparison. To ensure sufficient atmospheric mixing, only runs with  
433 wind speed  $> 2 \text{ m s}^{-1}$  were selected. The path/intake [CH<sub>4</sub>] ratios reveal a significant  
434 temperature dependence (Fig. 6), with a slope of  $-9.4 \times 10^{-3} \text{ K}^{-1}$  for both. The ratio  
435 equals 1 at 23 ( $\pm$ 2) °C (equal to the reference temperature, 296 K, of the HITRAN  
436 database – see Appendix), while at 0 °C the open-path laser would overestimate [CH<sub>4</sub>]  
437 by about 20 %. It was further tested whether the [CH<sub>4</sub>] ratios depended on specific  
438 humidity or pressure, with negative result. This test was repeated with the data binned  
439 into 3 K wide temperature classes, to prevent that the temperature dependence would

440 overwhelm any trends with the other variables, but no dependence on humidity or  
441 pressure could be identified. By contrast, when the data were binned into five classes  
442 of specific humidity, or pressure, then within each class there was still a temperature  
443 dependence, represented by a slope that agreed within  $\pm 30\%$  with that in Fig. 6. For  
444 all subsequent analyses, the  $[\text{CH}_4]$  data from the laser system are therefore linearly  
445 corrected for temperature, using the slope from Fig. 6 as the adjustment factor.

446 To our knowledge, such a temperature dependence of the GasFinder has not been  
447 reported in the micrometeorological literature before. Possible physical causes are  
448 discussed in the Appendix.

449

450

## 451 **5. Results and Discussion**

### 452 **5.1 Animal behaviour**

453 On the majority of mornings, the cattle were taken off-site between 7:15 and 8:00 h, to  
454 have their gas collection yokes changed. During their absence, the feed was laid out in  
455 the enclosures. When the cattle returned, between 10:00 and 10:30 h, they began  
456 feeding immediately, and used up the available rations within a few hours. The rest of  
457 the day they spent mostly drinking, resting and ruminating, moving around less than  
458 grazing animals typically would. It was clear from observations, of the cattle  
459 themselves as well as the distribution of their excreta, that they stayed preferably in the  
460 W parts of the enclosures, frequently bunching near the W fence line (and  
461 occasionally, around the water troughs). This behaviour was very different from the  
462 freely-grazing behaviour observed by Laubach and Kelliher (2004; 2005b) and  
463 Laubach et al. (2008), when the paddock area was by and large evenly covered by the  
464 cattle, for most of the time. The uneven spatial animal distribution in the present  
465 experiment had implications for the emission rate determinations. These are discussed  
466 in Section 5.5.

467

### 468 **5.2 Feed intake and emission rates at the animal scale**

469 On average, the baleage contained 399 ( $\pm 26$ ) g dry matter (DM) per kg wet weight. On  
470 a DM basis, it contained 46.5 ( $\pm 5.5$ ) % neutral detergent fibre, 19.0 ( $\pm 3.4$ ) % crude  
471 protein, 9.4 ( $\pm 2.3$ ) % soluble sugars, 3.3 ( $\pm 0.5$ ) % lipids and 13.8 ( $\pm 3.0$ ) % ash. The  
472 estimated metabolisable energy content of the feed was 10.5 ( $\pm 0.4$ ) MJ (kg DM)<sup>-1</sup>.

473           The average liveweight (LW) of the cattle increased with level of feeding from  
474 331 kg at low level to 352 kg at the high level (Table 2). On a DM basis, feed on offer  
475 increased by 58 % from the low to the medium level of feeding and by 52 % from the  
476 medium to the high level. Visual assessment of feed refused showed negligible  
477 amounts (< 1 %) for Weeks 1 and 2. In Week 3, the soiled feed residues were  
478 considerably larger and estimated at 10 %. The resulting estimates of dry matter intake  
479 (DMI) were 4.3, 6.7 and 9.3 kg DM d<sup>-1</sup> animal<sup>-1</sup> for the low, medium and high feeding  
480 level, respectively (Table 2), and hence the relative increases from one feeding level to  
481 the next were 58 and 38 %, respectively.

482           The CH<sub>4</sub> emission rates were obtained with the SF<sub>6</sub> tracer-ratio technique, for the  
483 last 4 d of each week. The first 3 d were considered as an adjustment period to the new  
484 feeding level (as diet composition did not change and feeding level increased, rather  
485 than decreased, a period of 3 d sufficed). The average emission rates were  
486 70.8 (±13.5), 89.7 (±11.1) and 119.1 (±16.4) g CH<sub>4</sub> d<sup>-1</sup> animal<sup>-1</sup> for Weeks 1, 2 and 3,  
487 respectively (Table 2), where the numbers are means (± SD) of the 61 animals. Even in  
488 Week 3, when DMI was relatively generous, the CH<sub>4</sub> emission rate was low in  
489 comparison to those observed previously for freely-grazing steers at the same location  
490 (Laubach et al., 2008). The standard deviations in Table 2 convert to standard errors of  
491 the mean ≤ 2 g CH<sub>4</sub> d<sup>-1</sup> animal<sup>-1</sup>. We assume the weekly mean CH<sub>4</sub> emission rates to  
492 be “true” within this limit and use them as reference values for the other techniques.  
493 The accuracy of the SF<sub>6</sub> technique for estimating mean emissions of CH<sub>4</sub> has been  
494 proved by comparison to animal chamber measurements, both for sheep (Hammond et  
495 al., 2009) and cattle (Grainger et al., 2007).

496           The CH<sub>4</sub> emission rate increased with increase in feeding level, by 27 % from  
497 the low to the medium feeding level and by 33 % from medium to high. Similarly, the  
498 estimates of CH<sub>4</sub> emissions expressed per unit of LW increased with increasing  
499 feeding levels, by 24% between low and medium feeding levels and by 28% between  
500 medium and high feeding levels. These increases were significant (P < 0.01), but less  
501 than a proportional increase with DMI. Hence, the estimated CH<sub>4</sub> yield per unit of feed  
502 intake decreased, by 19 % from the low to the medium feeding level and by 4 % from  
503 medium to high (Table 2). Decreasing CH<sub>4</sub> yields at increasing feeding levels were  
504 already reported by Blaxter and Clapperton (1965), and have been attributed to a faster  
505 rate of passage of feed through the alimentary tract (Benchaar et al., 2001; Pinares-  
506 Patiño et al., 2003).

507 The CH<sub>4</sub> emissions per feed intake are unusually low, compared to the mean for  
508 ryegrass-fed cattle in NZ, of 19.1 ( $\pm$ 3.70) g CH<sub>4</sub> (kg DMI)<sup>-1</sup> (Hammond et al., 2009).  
509 The observed CH<sub>4</sub> yields, of 4.6, 3.7 and 3.6 % of gross energy intake (GEI), are closer  
510 to the IPCC default value for concentrate-fed cattle, 3.0 %, than to the default value for  
511 grazing cattle, 6.5 % (IPCC, 2006, p. 10-30). This is likely to be due to the baleage  
512 used in this study, in contrast to fresh grass being the usual cattle diet in New Zealand.  
513 It is well established (McDonald et al., 1991; Huhtanen and Jaakkola, 1993; González  
514 et al., 2007) that ensiling reduces the ruminal degradability of dry matter. A computer  
515 simulation by Benchaar et al. (2001) took this effect into account, and also that CH<sub>4</sub>  
516 production decreases as pH and acetate/propionate ratio decrease. It predicted that CH<sub>4</sub>  
517 emission from feeding lucerne silage would be lower than that from hay (3.7 vs. 5.4 %  
518 of GEI). In light of this result, the observed CH<sub>4</sub> yields appear plausible. Yet, a deeper  
519 discussion of this finding is beyond the scope of this study.

520

### 521 **5.3 Emission rates at the herd scale**

522 For any of the herd-scale techniques, the run-to-run variations of the obtained CH<sub>4</sub>  
523 emission rate,  $Q_c$ , were considerable. This was the compounded effect of several  
524 sources of variability: true emission rate changes in response to the animals' digestion  
525 processes, variability of source locations relative to the instruments as the animals  
526 moved around, influence of variations in wind speed and direction on the effective  
527 footprint of the various techniques, influence of wind speed and direction on the  
528 magnitude of the concentration differences to be resolved, and random instrument  
529 error (precision). Data availability also differed between techniques. This was partly  
530 due to scheduled calibration checks and occasional malfunctions, e.g. dew on optical  
531 components, failure of electronic components (resolved by replacement). In addition,  
532 wind direction affected availability of the different techniques in different ways. The  
533 techniques involving the vertical profile mast (mass-budget and BLS-profile) required  
534 the mast to be downwind of at least a few steers, so gave meaningful emission rates  
535 only for directions between 210 and 330 °. The techniques involving the two FTIR  
536 instruments required one of them to be up- and the other downwind, so data were  
537 accepted only if wind direction was within  $\pm$ 50 ° of either W or E. By contrast, the  
538 four-path laser system could be used for all wind directions. Periods of low wind speed  
539 were excluded for all techniques, when friction velocity was less than 0.1 m s<sup>-1</sup>.

540           Given the different operational constraints for each technique, and the overall  
541 goal to assess the suitability of each to detect a change between weekly emission rates,  
542 the results were first compared on the basis of weekly diurnal courses. To construct  
543 these, for each week and technique the 20-min emission rates were sorted by time of  
544 day, into 1-h wide bins, and then bin-averaged. The results are discussed in the  
545 following subsections.

546

### 547 *5.3.1 Diurnal pattern of emissions*

548 Despite obvious discrepancies between techniques, a few features of the emission  
549 pattern appear robust (Fig. 7). From about 10:00 h, when the cattle entered their  
550 enclosures,  $Q_c$  increased steeply until it reached a maximum typically around 13:00 h.  
551 In Weeks 1 and 2,  $Q_c$  began declining after that, according to all techniques, while in  
552 Week 3 the techniques disagreed as to whether  $Q_c$  further increased or decreased until  
553 18:00 h. This pattern suggests that the maximum  $\text{CH}_4$  emissions typically occurred  
554 within 2 h of the time of maximum feeding activity. Feeding always started soon after  
555 10:00 h and in Weeks 1 and 2 was practically finished by noon, while it was stretched  
556 out longer in Week 3, when food amounts on offer exceeded the animals'  
557 requirements. Similar phase relationships between feeding time and maximum-  
558 emission time were observed for feedlot cattle (Loh et al., 2008; Gao et al., 2011) and  
559 sheep in test chambers (Lassey et al., 2011). From late afternoon throughout the night,  
560  $Q_c$  generally decreased, reaching the lowest levels from about 7:00 to 9:00 h, before  
561 the new day's feed ration was offered.

562           With all herd-scale techniques,  $Q_c$  from 10:00 to about 22:00 h ( $\pm 3$  h) was  
563 usually larger than the daily average obtained from the  $\text{SF}_6$  tracer-ratio technique, and  
564 during the other half of the diurnal cycle,  $Q_c$  was smaller than the average. This  
565 reflected that most of the feeding and ruminating occurred during the daylight hours.  
566 Similar activity patterns are common for free-ranging animals.

567

### 568 *5.3.2 Profile-based techniques*

569 The mass-budget and BLS technique that used the vertical  $[\text{CH}_4]$  profile from the  
570 closed-path analyser detected virtually the same temporal pattern of emission rates  
571 (Fig. 7). On a run-to-run basis, the two techniques agreed typically within 10 %, as  
572 was found in other experiments (Laubach and Kelliher, 2005a; Gao et al., 2009;  
573 Laubach, 2010).

574 From one week to the next, the daytime observations (10:00 to 18:00 h) clearly  
575 showed increasing  $Q_c$ . For the night-time hours, this cannot be stated. On many days  
576 the wind direction turned from W in the afternoon, via N in the evening, to NE at  
577 night, and then wind speed usually dropped, with friction velocity falling below the  
578 acceptance threshold. Since the profile-based techniques relied on westerly wind  
579 directions, night-time data availability was poor, and the available data suffered from  
580 low replication rates. This is particularly evident in Week 1 (Fig. 7, left panel), where  
581 lack of points for some hours, and a lack of error bars for others, indicated that none or  
582 only one run per bin, respectively, was available.

583 Consistently for all three weeks, and almost all times of day, these techniques  
584 gave higher  $Q_c$  than the other techniques. The likely cause for this is that for most of  
585 the time, the animals were not spread evenly across the enclosures. Rather, they  
586 preferred to stay near the W fence line. Consequently, for the wind directions most  
587 frequent and best-suited for the profile-based techniques, the actual animal density in  
588 the effective source area tended to be larger than the nominal mean animal density  
589 across the enclosures, which was used for converting emission rates per area to  
590 emission rates per animal. Hence, the latter were frequently overestimated. This point  
591 is further elaborated in Section 5.5.

592

### 593 *5.3.3 BLS with open-path instruments*

594 During the nights of Week 1, the laser system showed the highest data availability of  
595 all herd-scale techniques, while no data were available from the FTIR either due to  
596 unsuitable wind direction or, during one night, failure of the instrument W of the  
597 cattle. During day-time in Week 1, and throughout Weeks 2 and 3, there was generally  
598 reasonable agreement in the temporal pattern of  $Q_c$  retrieved by BLS from the FTIR  
599 and from the laser system. The laser system gave larger variability throughout,  
600 indicated both by the error bars in Fig. 7 and the variations from hour to hour. This  
601 may appear counter-intuitive, given that the laser system consisted of four paths almost  
602 completely surrounding the emission sources, so it should have provided the most  
603 representative coverage of the emissions plume for any wind direction. However, the  
604 more erratic behaviour of  $Q_c$  from the laser system can be explained by its poorer  
605 measurement precision.

606 The BLS results from both open-path systems showed an increase of  $Q_c$  from  
607 one week to the next. Each week, the hourly averages from the laser system were

608 spread roughly from 0.3 times to 1.7 times the average  $Q_c$  from the SF<sub>6</sub> tracer-ratio  
609 technique. The corresponding FTIR data spanned roughly between night-time minima  
610 (when available) of 0.5 times  $Q_c$  and day-time maxima of 1.5 times  $Q_c$  from the SF<sub>6</sub>  
611 technique.

612

#### 613 *5.3.4 External tracer-ratio technique*

614 During Week 1,  $Q_c$  from the N<sub>2</sub>O tracer-ratio technique agreed closely with that from  
615 BLS using the same [CH<sub>4</sub>] input data from the FTIR (Fig. 7). During Week 2, the N<sub>2</sub>O  
616 technique gave systematically lower  $Q_c$  than BLS, and for many hourly bins the lowest  
617 of all herd-scale techniques. The same is true at day-time in Week 3, but not at night.  
618 In effect, this technique gave the smallest variations with time of day, tracking the  
619 average  $Q_c$  from the SF<sub>6</sub> technique more closely than the other herd-scale techniques.  
620 This can be explained by the fact that the tracer ratio is independent of the flow field  
621 parameters, hence these parameters do not contribute to its error. The sampling errors  
622 of the external tracer-ratio technique are those of the mole fractions and the mean N<sub>2</sub>O  
623 release rate, while for BLS they are those of the mole fractions and a number of wind  
624 and turbulence parameters.

625

### 626 **5.4 Weekly mean emission rates**

627 To obtain mean CH<sub>4</sub> emission rates for each feeding level, with each technique, the  
628 mean diurnal courses were averaged. This gave more representative estimates than  
629 simply averaging all available runs, which would have weighted day-time more  
630 strongly than night-time (because of the more frequent occurrence of unsuitable calm  
631 periods at night). In some instances, especially in Week 1, some night-time hours were  
632 not covered by the data, which made some residual bias towards day-time inevitable.  
633 Random errors of the weekly means were obtained by propagating the standard errors  
634 of the mean for the hourly bins (root-mean-squares estimate). Fig. 8 displays the  
635 results, discussed in the following two subsections.

636

#### 637 *5.4.1 Comparison of absolute mean emission rates*

638 For all three weeks, the techniques using vertical profile data gave the largest mean  
639 emission rates. In Week 1, these were 68 and 56 % higher than  $Q_c$  from the SF<sub>6</sub> tracer-  
640 ratio technique, for mass-budget and BLS, respectively. In Weeks 2 and 3, they

641 exceeded the SF<sub>6</sub> technique consistently by 35 (±2) %. As noted in 5.3.2, the likely  
642 main cause for these biases was the systematically uneven animal distribution,  
643 resulting in a discrepancy between the actual animal density (in the effective source  
644 area of these techniques) and the nominal mean animal density. Another contributing  
645 factor was probably the low data yield at night, which could have led to potentially  
646 erratic results for some night-time hours. In Week 1, when several night-time hours  
647 had no data coverage at all, that would also have led to bias in the diurnal average.

648 The BLS technique with the four-path laser gave weekly mean  $Q_c$  that were  
649 18 % lower than from the SF<sub>6</sub> technique in Week 1, and 8 and 6 % higher (not  
650 significantly different) in Weeks 2 and 3, respectively. The low average in Week 1  
651 appears to be caused mostly by low hourly values between 13:00 and 15:00, which do  
652 not fit with the diurnal patterns observed in the other weeks and by the other  
653 techniques (Fig. 7). Some of the contributing runs with low  $Q_c$  had unstable  
654 stratification and SE to S winds; it is possible that the N laser path, at only 1.07 m  
655 height, was too close to the ground to accurately sample the emissions plume in  
656 convective conditions.

657 The BLS technique with FTIR data exceeded  $Q_c$  from the SF<sub>6</sub> technique by 25  
658 and 21 % in Weeks 1 and 2, and agreed within < 1 % (no significant difference) in  
659 Week 3. The N<sub>2</sub>O tracer-ratio technique exceeded  $Q_c$  from the SF<sub>6</sub> technique by 18 %  
660 in Week 1, by 3 % (not significant) in Week 2, and fell short of it by 10 % in Week 3  
661 (different at the 95 % confidence level). For both techniques, the overestimate in  
662 Week 1 is explained by the lack of night-time data, causing the weekly mean to over-  
663 represent the times of feeding activity. For the BLS technique in Week 2, there is no  
664 obvious cause for an overestimate. Valid FTIR data were obtained for either W or E  
665 winds; the fraction of W winds was 90, 80 and 74 % for Week 1, 2 and 3, respectively,  
666 so any effects related to wind direction are unlikely to explain differences between  
667 weeks in the BLS technique's performance. For the N<sub>2</sub>O tracer-ratio technique in  
668 Week 3, it can be seen in Fig. 7 that the day-time emission rates were substantially  
669 lower than those recorded by the other techniques. A possible cause for low emission  
670 rate estimates is an underestimate of the N<sub>2</sub>O release rate, obtained from weighing of  
671 the canisters and using the time and temperature dependence determined by Bai  
672 (2010). However, the release rates during Week 3 showed very consistent diurnal  
673 courses. It thus remains unclear whether the N<sub>2</sub>O tracer-ratio technique underestimated  
674 the day-time emissions for this week. If it did, that would have caused the weekly  
675 average to be an underestimate.

676 With the exceptions as discussed, it appears that all techniques using open-path  
677 [CH<sub>4</sub>] measurements delivered weekly mean  $Q_c$  that agreed with the enteric tracer-  
678 ratio technique within 10 %, which is the commonly assumed magnitude of  
679 uncertainty for micrometeorological flux measurement techniques. It should be noted  
680 that the herd-scale techniques include CH<sub>4</sub> emissions from the rectum, while the SF<sub>6</sub>  
681 tracer-ratio technique does not. Comparisons of the SF<sub>6</sub> technique to animal-chamber  
682 measurements by McGinn et al. (2006) and Grainger et al. (2007) indicated lower  
683 emission estimates for the SF<sub>6</sub> technique of 4 % and 6 %, respectively, which could in  
684 part be attributed to rectal emissions missed by this technique. Accounting for a small  
685 underestimate by the SF<sub>6</sub> technique would not change the finding that, for weekly  
686 averages, the three open-path techniques agreed well with it. For the two vertical-  
687 profile techniques, the differences to the SF<sub>6</sub> technique would be somewhat reduced.

688

#### 689 *5.4.2 Suitability to resolve a change in mean emission rate*

690 According to the SF<sub>6</sub> tracer-ratio technique, the mean CH<sub>4</sub> emission rate increased by  
691 27 % from Week 1 to 2, and by 33 % from Week 2 to 3. The mass-budget technique  
692 failed to detect the first change, but identified an increase of the right magnitude  
693 (34 %) for the second. The BLS technique using vertical profile data recorded CH<sub>4</sub>  
694 emission increases for both feed level changes, of 10 % and 34 % respectively, but the  
695 change from Week 1 to 2 was not significant. For both these techniques, the failure to  
696 detect the first step change must be attributed to sparse data coverage and large  
697 variability during night-time hours, since Fig. 7 shows that the day-time emission rates  
698 increased for both techniques.

699 The other techniques (N<sub>2</sub>O tracer-ratio and BLS with FTIR or four-path laser)  
700 detected both changes in weekly emission rates at > 99 % confidence levels. The N<sub>2</sub>O  
701 tracer-ratio technique recorded the lowest week-to-week increases of these techniques,  
702 of 11 and 16 %, respectively, which were just under half of the increases shown by the  
703 SF<sub>6</sub> technique. For the first step change, this can be explained by the overestimate in  
704 Week 1 due to a lack of night-time data. For the second step change, this is  
705 computationally caused by low day-time emission rates in Week 3, yet the ultimate  
706 cause for these is unclear.

707 Since the 95 % confidence intervals are about twice and 99 % confidence  
708 intervals roughly 3 times as large as the standard errors in Fig. 8, one may infer that (in  
709 all instances where data availability was sufficient) all techniques would probably

710 detect changes of order 20 % with 99 % confidence, and changes of order 10 % with  
711 95 % confidence. Such inferences should of course be tested with separate  
712 experiments.

713

### 714 **5.5 Sensitivity of emission estimates to animal distribution**

715 All herd-scale techniques except for the N<sub>2</sub>O tracer-ratio technique derived the CH<sub>4</sub>  
716 emission rate on a per-area basis, and for that required specification of the shape and  
717 extent of the emissions' source area (as determined by the fence lines). The per-area  
718 emissions were then converted into per-animal emission rates, with the same  
719 conversion factor for all techniques (61.8 m<sup>2</sup> animal<sup>-1</sup>). However, as noted before, for  
720 most of the time the steers did not spread evenly across the fenced area. A sensitivity  
721 study was undertaken to assess how the uneven spread affected the emission rate  
722 estimates from the different techniques, by defining a "cattle-preference" area that  
723 excluded the rarely-visited NE and SE corners of the enclosure (Fig. 9).

724 One effect of this changed area specification is common to all techniques: it  
725 reduces the conversion factor by 16.2 %, to 51.8 m<sup>2</sup> animal<sup>-1</sup>. Further effects differ  
726 between the techniques. For the mass-budget technique, the source area contact  
727 distance upwind of the mast, traversed by the air flow, is reduced, but only for the  
728 marginally acceptable wind directions more than ±40 ° off W, not for the wind  
729 directions around W. For the BLS technique, the "touchdown" statistics (where the  
730 backwards-tracked air parcels are in contact with the ground) remain unaltered, but the  
731 classification which touchdowns contribute to the downwind concentration elevation  
732 changes, which leads to altered three-dimensional concentration fields (normalised by  
733  $Q_c$ , which is assumed homogeneous within the source area). The net effect of that is  
734 expected to differ for different arrays of sensor locations (vertical profile, two-path  
735 system, and four-path system).

736 The results of the sensitivity tests are evaluated by taking the ratio of the  
737 emission rate with the "cattle-preference" area prescribed as the source to the emission  
738 rate with the fenced enclosure prescribed as the source. In Fig. 10, this ratio is  
739 displayed against wind direction. Distinct relationships are observed, for each  
740 technique. They display some scatter that stems from the variability of the turbulent  
741 flow parameters, mainly stability and the standard deviation of the crosswind  
742 fluctuations. Fig. 10 also implicitly provides an overview how the accepted runs, for  
743 each technique, were distributed with wind direction.

744 The relationships for the mass-budget and the BLS technique using the same  
745 vertical [CH<sub>4</sub>] profile are virtually identical. This is reassuring because it means the  
746 way in which crosswind fluctuations are accounted for in the mass-budget approach of  
747 Laubach and Kelliher (2004) is in agreement with the Lagrangian simulation statistics.  
748 For these two techniques, the reduced conversion factor determines the result in the  
749 sector from 230 to 310 °, while outside this sector there is a steep rise in the ratio with  
750 increasing angular distance from 270 °. As W winds dominated throughout the  
751 campaign, the net effect of prescribing the “cattle-preference” area on the weekly mean  
752  $Q_c$  would be a reduction by 10 to 16 % in each week. The reduced  $Q_c$  would be in  
753 better agreement with the two tracer-ratio techniques. The high sensitivity to source  
754 area choice is the biggest disadvantage of the vertical-profile techniques. It could be  
755 reduced by increasing the distance between sources and sensors, as recommended by  
756 Flesch et al. (2005), but at the cost of further narrowing the range of acceptable wind  
757 directions. Here, the sensitivity to source area choice, especially for the marginal wind  
758 directions, may go some way to explain why the step change from Week 1 to 2 was not  
759 detected by the vertical-profile techniques.

760 For the two-path FTIR system, the changed source area prescription leads to an  
761 increase of  $Q_c$  for westerly winds, by 10 to 20 %, and to a decrease for easterly winds,  
762 by 5 to 10 % (Fig. 10). The result is not symmetrical because of the different distances  
763 between FTIR path and nearest enclosure fence in W and E, respectively. The increase  
764 for W directions can be qualitatively understood as follows. The same downwind line-  
765 integrated [CH<sub>4</sub>] as before is now interpreted to be caused by an emitting area that is  
766 (in its N and S parts) farther away from the receptor path than with the original source  
767 area prescription. The longer distance would cause enhanced dilution by turbulent  
768 dispersion in all dimensions. To counter that dilution effect, the model must assign an  
769 increased emission rate to the source area, in higher proportion than that necessary to  
770 compensate for its reduced size. For E wind directions, the situation is different: the  
771 distance between the W boundary of the source area and the then-downwind W  
772 receptor path is unchanged. The source area extent is reduced in the corners farthest  
773 away from the receptor path, which had the smallest touchdown density and thus made  
774 the smallest relative contribution to the [CH<sub>4</sub>] signal. Consequently, the BLS model  
775 computes an emission rate for the reduced area that is increased less than  
776 proportionally, compared to the result for the total fenced area, and a smaller emission  
777 rate per animal results. Because W winds were far more frequent than E winds  
778 (Fig. 10), the effect of the changed source area prescription on the weekly mean  $Q_c$

779 from the FTIR would be an increase, of order 10 %. This would decrease the  
780 differences to the vertical-profile techniques, but would on the other hand increase the  
781 differences to the two tracer-ratio techniques.

782 The pattern for the four-path laser-system is quite different to that for the two-  
783 path FTIR system (Fig. 10). For most wind directions, the change caused by the re-  
784 definition of the source area is within  $\pm 10\%$ . Systematic decreases larger than that  
785 occur around 120, 225 and 320 °, which are “diagonal” directions towards corners  
786 poorly covered by downwind receptor paths (see Fig. 1); longer paths reaching past the  
787 corners would probably have reduced the extent of these decreases. Overall, it appears  
788 that the four-path setup is the least sensitive to the exact source area specification,  
789 because the touchdown coverage achieved with multiple paths is more evenly  
790 distributed than for setups with a single downwind path or point (mast). Fig. 10  
791 suggests that the effect of the re-defined source area on the weekly mean  $Q_c$  would  
792 have been within  $\pm 5\%$ , so despite the poorer precision of the laser system, compared  
793 to the other CH<sub>4</sub> instruments, the average results were robust.

794

795

## 796 **6. Conclusions**

797 Weekly-averaged CH<sub>4</sub> emission rates from a herd of steers fed with pasture baleage  
798 correlated strongly with dry-matter intake, yet less than proportionally. According to  
799 the enteric tracer-ratio technique, the increases in emission rate from one week to the  
800 next were 27 and 33 %, respectively. Comparing the five herd-scale techniques against  
801 this animal-scale technique and against each other yielded the following results and  
802 insights:

- 803 • All techniques were able to detect the increase from Week 2 to 3 with  
804 > 99 % confidence. The open-path-based techniques and the animal-scale  
805 technique also detected the increase from Week 1 to 2 with > 99 %  
806 confidence.
- 807 • The techniques based on vertical [CH<sub>4</sub>] profiles were carried out with the  
808 most accurate and well-calibrated CH<sub>4</sub> instrument, but they were prone to  
809 the lowest data yield, due to wind direction restrictions, and were the  
810 most sensitive to inhomogeneous spatial distribution of sources.

- 811           • The techniques based on path-averaged [CH<sub>4</sub>] measurements generally  
812           gave accurate weekly emission rates (within 10 % of the enteric tracer-  
813           ratio technique).
- 814           • The external tracer-ratio technique, which is inherently insensitive to  
815           uneven source distribution, provided emission rates with the smallest run-  
816           to-run variability of all techniques. This advantage in precision came at  
817           the cost of considerable extra effort to provide suitable tracer release at  
818           the exact CH<sub>4</sub> source locations. However, on a week-to-week basis, this  
819           technique possibly underestimated the changes in mean emission rate.
- 820           • With the BLS technique, four measurement paths offer significant  
821           advantages over two, not only in providing higher data yield (by  
822           including all wind directions), but also in being less sensitive to uneven  
823           source distribution. This was demonstrated by the laser system delivering  
824           unbiased weekly emission rates, despite the poorer accuracy of its  
825           individual mixing-ratio measurements.

826   These findings suggest as the ideal herd-scale technique one that combines the  
827   strengths of the accurate closed-path analyser with the strengths of a path-averaging  
828   approach. This could be done either by using horizontally-aligned arrays of many  
829   intakes, drawing air simultaneously to a closed-path analyser (e.g. perforated intake  
830   tubes), or by using high-quality open-path instruments, like the FTIR, in conjunction  
831   with regular cross-checks against such an analyser.

832           Of interest from the perspectives of animal nutrition and greenhouse gas  
833   emissions mitigation is that this experiment, using grass baleage as feed, recorded far  
834   lower CH<sub>4</sub> yields per feed intake than ever found with fresh grass. It should be  
835   followed up in further studies which properties of the baleage effected this result.

836

837

### 838   **Acknowledgements**

839   The experiment and the involvement of the NZ-based authors were funded by New  
840   Zealand's Foundation of Research, Science and Technology (FRST) until 2011, and by  
841   Core Funding for Crown Research Institutes from the Ministry of Business, Innovation  
842   and Employment's Science and Innovation Group since. The participation of the  
843   University of Wollongong was funded by ARC Linkage Grant LP0561000. E. A.

844 Cárdenas R. received a fellowship from the Livestock Emissions Abatement Research  
845 Network (LEARN) of NZ's Ministry for Agriculture and Forestry. The experiment  
846 would not have been possible without the logistical backing of Steve Lees (Aorangi  
847 Farm) and his staff. We thank Dan Robinson for drafting the animals and weighing  
848 and distributing their feed supply, Edgar Sandoval for help with the animal handling,  
849 and Sarah McLean and Grant Taylor for laboratory analyses (all AgResearch). Tony  
850 McSeveny and Peter Berben (Landcare Research) were instrumental in setting up the  
851 scientific equipment. We thank Dr Frank Kelliher for the initial idea of this experiment  
852 and valuable advice throughout the planning stages.

853

854

### 855 **Appendix: Temperature dependence of the open-path methane laser**

856 The open-path CH<sub>4</sub> laser system (Boreal Laser, 2005) reports its measurement as a  
857 mole fraction  $\chi$  (ppm), despite the fact that the absorption along the measurement path  
858 depends primarily on the absolute density of CH<sub>4</sub> molecules on the path,  $\rho_c$  (mol m<sup>-3</sup>  
859 or kg m<sup>-3</sup>). The reported mole fraction is only correct at an assumed reference  
860 temperature,  $T_{ref}$ , and pressure,  $p_{ref}$ . In general, though, actual temperature,  $T$ , and  
861 pressure,  $p$ , along the path will differ from these reference values, in which case the  
862 CH<sub>4</sub> mole fraction requires correction. There are three mechanisms by which  $T$  and  $p$   
863 affect the mole fraction.

864 The first of these mechanisms is the universal gas law. If this was the only one,  
865 then the correct mole fraction,  $\chi_{corr}$ , could be obtained from the reported mole fraction,  
866  $\chi_{meas}$ , as:

$$867 \chi_{corr} = \chi_{meas} (T p_{ref}) / (T_{ref} p) \quad (A1)$$

868 with temperatures in K. The relative sensitivity of  $\chi$  to temperature change (for  $p = p_{ref}$ )  
869 would be obtained as  $\chi^{-1} \partial\chi/\partial T = T_{ref}^{-1} = 3.37 \cdot 10^{-3} \text{ K}^{-1}$ , at  $T = T_{ref} = 296 \text{ K}$  (the  
870 choice of value is explained below). For example, a deviation of  $T$  from  $T_{ref}$  by 3 K  
871 would produce 1 % relative error in  $\chi$ , equivalent to 17 ppb absolute error at the global  
872 background mole fraction. A similar error would result from a difference of 1 kPa  
873 between  $p$  and  $p_{ref}$ , assuming the standard value  $p_{ref} = 101.3 \text{ kPa}$ . When determining  
874 emission rates, mole fraction differences between two measurement locations are  
875 crucial input variables. Assuming that temperature and pressure at the two locations

876 were equal, the mole fraction difference would be subject to the same relative error as  
877 the mole fraction itself, and so would be the computed emission rate.

878 The second mechanism is the dependence of the spectral intensity of the  
879 absorption line,  $S$ , on temperature. In first-order approximation, this can be written as:

$$880 \quad S(T) = S(T_{ref}) \exp(-b E_0/T) / \exp(-b E_0/T_{ref}) \quad (\text{A2})$$

881 where  $E_0$  is the energy of the lower state of the transition (in  $\text{cm}^{-1}$ ),  $b = 1.438775 \text{ K cm}$   
882 is a collection of fundamental constants (Planck constant times speed of light in  
883 vacuum divided by Boltzmann constant), sometimes referred to as the “second  
884 radiation constant”, and  $T_{ref}$  is a reference temperature, which is set to 296 K in the  
885 spectroscopic database HITRAN (Rothman et al. 1998, 2005). The arguments of the  
886 exponential function in (A2) quantify the probability with which the lower state is  
887 occupied for a given temperature. In (A2), the effects of the partition function and of  
888 stimulated emission are neglected; the full equation can be found e.g. in Rothman et al.  
889 (1998, their Eq. A11). Differentiating (A2) and normalising by  $S$  results in:

$$890 \quad S^{-1} dS/dT = b E_0/T^2 \quad (\text{A3})$$

891 The absorption line feature used by the GasFinder is the R3 triplet in the  $2\nu_3$  band,  
892 centred at the wavenumber  $6046.953 \text{ cm}^{-1}$  (J. Tulip, Boreal Laser, pers. comm., 2003).  
893 The HITRAN database (Rothman et al., 2005) gives  $E_0$  for each line of the triplet with  
894 6 digits precision; each  $E_0$  can be rounded to  $62.88 \text{ cm}^{-1}$ . Using this value to evaluate  
895 (A3) at  $T = T_{ref}$  gives  $S^{-1} dS/dT = 1.03 \cdot 10^{-3} \text{ K}^{-1}$ . Since observed absorbance is  
896 proportional to  $S$ , this relative temperature sensitivity of spectral intensity also applies  
897 to the reported mole fraction. It is about 0.3 of the gas-law temperature sensitivity.

898 The third mechanism is band-broadening of the absorption line. Some of this is  
899 caused by collisions of the absorbing molecules with other molecules which subtly  
900 alter the energy of the transition between the quantum states; this increases with  
901 increasing pressure (increasing the probability for collisions). There is also band-  
902 broadening due to the Doppler effect; this increases with increasing temperature  
903 (Hollas, 2004). While the HITRAN database provides parameters to estimate  
904 broadening effects, these do not include the effects of potentially highly variable water  
905 vapour concentration (Tuzson et al., 2010). Also, Frankenberg et al. (2008) point out  
906 that the catalogued broadening parameters are by no means certain, and for multiplets  
907 in the  $2\nu_3$  band of  $\text{CH}_4$  almost impossible to model. Further, the net effect of band-

908 broadening on the line-shape retrieval algorithm would have to be modelled or  
909 experimentally determined, which is beyond the scope of this study.

910 To complicate matters further, instrument-specific parameters also influence the  
911 temperature and pressure sensitivity of the GasFinder (J. Terry, Boreal Laser, pers.  
912 comm., 2012). According to correction curves determined by the manufacturer, the net  
913 sensitivity of  $\chi$  to pressure in the range 97 to 105 kPa is typically within  $\pm 0.3\%$  kPa<sup>-1</sup>.  
914 Over the pressure range of the present experiment (100.0 to 102.7 kPa), this sensitivity  
915 has only minor effect, compared to the temperature dependence from all mechanisms  
916 combined. For the GasFinder system used here, the total temperature sensitivity was  
917 found about 2.5 times that from the gas law alone. The interaction of spectroscopic  
918 effects with instrument parameters is therefore important, however not fully  
919 predictable. Because of that, the temperature sensitivities of any two GasFinder  
920 systems will probably differ, and individual calibration is recommended.

921

## 922 **References**

- 923 Bai, M., 2010. Methane emissions from livestock measured by novel spectroscopic techniques.  
924 PhD thesis, University of Wollongong, Australia, 303 pp.
- 925 Benchaar, C., Pomar, C., Chiquette, J., 2001. Evaluation of dietary strategies to reduce  
926 methane production in ruminants: A modelling approach. *Can. J. Anim. Sci.*, 81: 563-  
927 574.
- 928 Blaxter, K.L., Clapperton, J.L., 1965. Prediction of the amount of methane produced by  
929 ruminants. *British J. of Nutrition*, 19: 511-22.
- 930 Boreal Laser, 2005. GasFinderMC Operation Manual. Boreal Laser Inc., Spruce Grove,  
931 Alberta, Canada, 79 pp.
- 932 CSIRO, 2007. Nutrient Requirements of Domesticated Ruminants. CSIRO Publishing,  
933 Collingwood, Victoria, Australia, 270 pp.
- 934 Flesch, T. K., Wilson, J. D., Yee, E., 1995. Backward-time Lagrangian stochastic dispersion  
935 models, and their application to estimate gaseous emissions. *J. Appl. Meteorol.*, 34:  
936 1320-1332.
- 937 Flesch, T. K., Wilson, J. D., Harper, L. A., Crenna, B. P., and Sharpe, R. R., 2004. Deducing  
938 ground-air emissions from observed trace gas concentrations: a field trial. *J. Appl.*  
939 *Meteorol.*, 43: 487-502.
- 940 Flesch, T. K., Wilson, J. D., Harper, L. A., and Crenna, B. P., 2005. Estimating gas emissions  
941 from a farm with an inverse-dispersion technique. *Atmos. Environ.*, 39: 4863-4874.
- 942 Frankenberg, C., Warneke, T., Butz, A., Aben, I., Hase, F., Spietz, P., and Brown, L. R., 2008.  
943 Pressure broadening in the 2 $\nu_3$  band of methane and its implication on atmospheric  
944 retrievals. *Atmos. Chem. Phys.*, 8: 5061-5075.
- 945 Gao, Z., Desjardins, R. L., and Flesch, T. K., 2009. Comparison of a simplified  
946 micrometeorological mass difference technique and an inverse dispersion technique for  
947 estimating methane emissions from small area sources. *Agric. For. Meteorol.*, 149: 891-  
948 898.

- 949 Gao, Z., Yuan, H., Ma, W., Liu, X., and Desjardins, R. L., 2011. Methane emissions from a  
950 dairy feedlot during the fall and winter seasons in Northern China. *Environmental*  
951 *Pollution*, 159: 1183-1189.
- 952 González, J., Faría-Mármol, J., Rodríguez, C.A., Martínez, A., 2007. Effects of ensiling on  
953 ruminal degradability and intestinal digestibility of Italian rye-grass. *Anim. Feed Sci.*  
954 *Technol.*, 136: 38-50.
- 955 Grainger, C., Clarke, T., McGinn, S. M., Auldist, M. J., Beauchemin, K. A., Hannah, M. C.,  
956 Waghorn, G. C., Clark, H., and Eckard, R. J., 2007. Methane emissions from dairy cows  
957 measured using the sulfur hexafluoride (SF<sub>6</sub>) tracer and chamber techniques. *J. Dairy*  
958 *Sci.*, 90: 2755-2766.
- 959 Griffith, D. W. T., 1996. Synthetic calibration and quantitative analysis of gas-phase FT-IR  
960 spectra. *Appl. Spectrosc.*, 50: 59-70.
- 961 Griffith, D. W. T., Bryant, G. R., Hsu, D., and Reisinger, A. R., 2008. Methane emissions from  
962 free-ranging cattle: Comparison of tracer and integrated horizontal flux techniques. *J.*  
963 *Environ. Qual.*, 37: 582-591.
- 964 Griffith, D. W. T., Deutscher, N. M., Caldwell, C. G. R., Kettlewell, G., Riggensbach, M., and  
965 Hammer, S., 2012. A Fourier transform infrared trace gas and isotope analyser for  
966 atmospheric applications. *Atmos. Meas. Tech.*, 5: 2481-2498.
- 967 Hammond, K. J., Muetzel, S., Waghorn, G. C., Pinares-Patiño, C. S., Burke, J. L., and Hoskin,  
968 S. O., 2009. The variation in methane emissions from sheep and cattle is not explained  
969 by the chemical composition of ryegrass. *Proceedings of the New Zealand Society of*  
970 *Animal Production*, 69: 174-178.
- 971 Harper, L. A., Denmead, O. T., Freney, J. R., and Byers, F. M., 1999. Direct measurements of  
972 methane emissions from grazing and feedlot cattle. *J. Anim. Sci.*, 77: 1392-1401.
- 973 Harper, L. A., Denmead, O. T., and Flesch, T. K., 2011. Micrometeorological techniques for  
974 measurement of enteric greenhouse gas emissions. *Anim. Feed Sci. Technol.*, 166-167:  
975 227-239.
- 976 Hollas, J. M., 2004. *Modern Spectroscopy*, 4<sup>th</sup> ed. John Wiley & Sons, Chichester, England,  
977 452 pp.
- 978 Huhtanen, P., Jaakkola, S., 1993. The effects of forage preservation method and proportion of  
979 concentrate on digestion of cell wall carbohydrates and rumen digesta pool size in cattle.  
980 *Grass Forage Sci.*, 48: 155-165.
- 981 IPCC, 2006. 2006 IPCC Guidelines for National Greenhouse Gas Inventories, Volume 4:  
982 Agriculture, Forestry and Other Land Use. Available online at [http://www.ipcc-](http://www.ipcc-nggip.iges.or.jp/public/2006gl/vol4.html)  
983 [nggip.iges.or.jp/public/2006gl/vol4.html](http://www.ipcc-nggip.iges.or.jp/public/2006gl/vol4.html).
- 984 Johnson, K., Huylar, M., Westberg, H., Lamb, B., Zimmerman, P., 1994. Measurement of  
985 methane emissions from ruminant livestock using a SF<sub>6</sub> tracer technique. *Environ. Sci.*  
986 *Technol.*, 28: 359-362.
- 987 Lassey, K. R., 2007. Livestock methane emission: from the individual grazing animal through  
988 national inventories to the global methane cycle. *Agric. For. Meteorol.*, 142: 120-132.
- 989 Lassey, K. R., Pinares-Patiño, C. S., Martin, R. J., Molano, G., and McMillan, A. M. S., 2011.  
990 Enteric methane emission rates determined by the SF<sub>6</sub> tracer technique: Temporal  
991 patterns and averaging periods. *Anim. Feed Sci. Technol.*, 166-167: 183-191.
- 992 Laubach, J., 2010. Testing of a Lagrangian model of dispersion in the surface layer with cattle  
993 methane emissions. *Agric. For. Meteorol.*, 150: 1428-1442.
- 994 Laubach, J., and Kelliher, F. M., 2004. Measuring methane emission rates of a dairy cow herd  
995 by two micrometeorological techniques. *Agric. For. Meteorol.*, 125: 279-303.

- 996 Laubach, J., and Kelliher, F. M., 2005a. Measuring methane emission rates of a dairy cow herd  
997 (II): results from a backward-Lagrangian stochastic model. *Agric. For. Meteorol.*, 129:  
998 137-150.
- 999 Laubach, J., and Kelliher, F. M., 2005b. Methane emissions from dairy cows: Comparing open-  
1000 path laser measurements to profile-based techniques. *Agric. For. Meteorol.*, 135: 340-345.
- 1001 Laubach, J., Kelliher, F. M., Knight, T., Clark, H., Molano, G., and Cavanagh, A., 2008.  
1002 Methane emissions from beef cattle – a comparison of paddock- and animal-scale  
1003 measurements. *Aust. J. Exp. Agric.*, 48: 132-137.
- 1004 Leuning, R., Baker, S. K., Jamie, I. M., Hsu, C. H., Klein, L., Denmead, O. T., and Griffith, D.  
1005 W. T., 1999. Methane emission from free-ranging sheep: a comparison of two  
1006 measurement methods. *Atmos. Environ.*, 33: 1357-1365.
- 1007 Loh, Z., Chen, D., Bai, M., Naylor, T., Griffith, D., Hill, J., Denmead, T., McGinn, S., and  
1008 Edis, R., 2008. Measurement of greenhouse gas emissions from Australian feedlot beef  
1009 production using open-path spectroscopy and atmospheric dispersion modelling. *Aust. J.*  
1010 *Exp. Agric.*, 48: 244-247.
- 1011 McDonald, P., Henderson, A.R., Heron, S.J.E., 1991. *The Biochemistry of Silage*, 2<sup>nd</sup> ed.  
1012 Chalcombe Publications, Marlow, Bucks., England, 340 pp.
- 1013 McGinn, S. M., Beauchemin, K. A., Iwaasa, A. D., and McAllister, T. A., 2006. Assessment of  
1014 the sulfur hexafluoride (SF<sub>6</sub>) tracer technique for measuring enteric methane emissions  
1015 from cattle. *J. Environ. Qual.*, 35: 1686-1691.
- 1016 Pinares-Patiño, C.S., Ulyatt, M.J., Lassey, K.R., Barry, T.N., Holmes, C.W., 2003. Rumen  
1017 function and digestion parameters associated with differences between sheep in methane  
1018 emissions when fed chaffed lucerne hay. *J. Agric. Sci. Camb.*, 140: 205-214.
- 1019 Pinares-Patiño, C. S., Waghorn, G. C., Machmüller, A., Vlaming, B., Molano, G., Cavanagh,  
1020 A. and Clark, H., 2007. Methane emissions and digestive physiology of non-lactating  
1021 dairy cows fed pasture forage. *Can J. Anim. Sci.*, 87: 601-613.
- 1022 Rothman, L. S., Rinsland, C. P., Goldman, A., Massie, S. T., Edwards, D. P., Flaud, J.-M.,  
1023 Perrin, A., Camy-Peyret, C., Dana, V., Mandin, J.-Y., Schroeder, J., McCann, A.,  
1024 Gamache, R. R., Wattson, R. B., Yoshino, K., Chance, K. V., Jucks, K. W., Brown, L.  
1025 R., Nemtchinov, V., and Varanasi, P., 1998. The HITRAN molecular spectroscopic  
1026 database and HAWKS (HITRAN Atmospheric Workstation): 1996 edition. *J. Quant.*  
1027 *Spectrosc. Radiat. Transfer*, 60: 665-710.
- 1028 Rothman, L. S., Jacquemart, D., Barbe, A., Benner, D. C., Birk, M., Brown, L. R., Carleer, M.  
1029 R., Chackerian Jr, C., Chance, K., Coudert, L. H., Dana, V., Devi, V. M., Flaud, J.-M.,  
1030 Gamache, R. R., Goldman, A., Hartmann, J.-M., Jucks, K. W., Maki, A. G., Mandin, J.-  
1031 Y., Massie, S. T., Orphal, J., Perrin, A., Rinsland, C. P., Smith, M. A. H., Tennyson, J.,  
1032 Tolchenov, R. N., Toth, R. A., Vander Auwera, J., Varanasi, P., and Wagner, G., 2005.  
1033 The HITRAN 2004 molecular spectroscopic database. *J. Quant. Spectrosc. Radiat.*  
1034 *Transfer*, 96: 139-204.
- 1035 Smith, T. E. L., Wooster, M. J., Tattaris, M., and Griffith, D. W. T., 2011. Absolute accuracy  
1036 and sensitivity analysis of OP-FTIR retrievals of CO<sub>2</sub>, CH<sub>4</sub> and CO over concentrations  
1037 representative of “clean air” and “polluted plumes”. *Atmos. Meas. Tech.*, 4: 97-116.
- 1038 Tuzson, B., Hiller, R. V., Zeyer, K., Eugster, W., Neftel, A., Ammann, C., and Emmenegger,  
1039 L., 2010. Field intercomparison of two optical analyzers for CH<sub>4</sub> eddy covariance flux  
1040 measurements. *Atmos. Meas. Tech.*, 3: 1519-1531.
- 1041  
1042

1043 **Tables**

1044

1045

1046

1047

1048 **Table 1** Overview of the techniques used to determine CH<sub>4</sub> emission rates from  
 1049 61 cattle, and the associated instrumentation for each technique.

1050

Technique	Instrumentation	Institution
<i>Animal scale:</i>		
Enteric tracer-ratio	SF <sub>6</sub> release in rumen, yokes on cattle, GC analysis	AgRes
<i>Herd scale:</i>		
External tracer-ratio	N <sub>2</sub> O release canisters on cattle, open-path FTIR	UoW
Mass-budget	Los Gatos with 7 intakes, cup anemometer profile	LCR
BLS from profile	Los Gatos with 7 intakes, cup & sonic anemometer	LCR
BLS from 4 paths	GasFinder MC, cup & sonic anemometer	LCR
BLS from 2 paths	open-path FTIR, cup & sonic anemometer	UoW

1051 AgRes = AgResearch, LCR = Landcare Research, UoW = University of Wollongong

1052

1053

1054

1055 **Table 2** The steers' average liveweight, estimated feed on offer, feed intake, and  
 1056 methane emissions (as obtained with the SF<sub>6</sub> tracer-ratio technique) at the  
 1057 three feeding levels. Values in parentheses are standard deviations, unless  
 1058 indicated otherwise <sup>1)</sup>.  
 1059

Period	Week 1	Week 2	Week 3	P value <sup>4)</sup>
Feeding level	Low	Medium	High	
Liveweight (kg)	331 (21)	339 (21)	352 (22)	n.d.
Feed (kg DM d <sup>-1</sup> animal <sup>-1</sup> ):				
offered <sup>1)</sup>	4.30 (0.09)	6.78 (0.15)	10.32 (0.23)	n.d.
refused (soiled) <sup>2)</sup>	0.04	0.06	1.03 (0.33)	n.d.
intake <sup>1)</sup>	4.26 (0.10)	6.72 (0.16)	9.29 (0.40)	n.d.
CH <sub>4</sub> emissions:				
g CH <sub>4</sub> d <sup>-1</sup> animal <sup>-1</sup>	70.8 (13.5)	89.7 (11.1)	119.1 (16.4)	<0.001
g CH <sub>4</sub> d <sup>-1</sup> (kg LW) <sup>-1</sup>	0.213 (0.036)	0.265 (0.028)	0.338 (0.039)	<0.001
g CH <sub>4</sub> (kg DMI) <sup>-1</sup>	16.6	13.4	12.8	n.d.
Y <sub>m</sub> (% of GEI) <sup>3)</sup>	4.6	3.7	3.6	n.d.

1060 <sup>1)</sup> Feed values per individual animal were calculated by dividing the estimated total for the herd  
 1061 by the number of animals, hence no s.d. is available. The indicated uncertainties are  
 1062 estimated weighing errors for the feed offered, and propagated from weighing error and  
 1063 visual estimation of refused amounts for feed intake.

1064 <sup>2)</sup> The amount of soiled feed was estimated by eye as a proportion of feed on offer: < 1 %,  
 1065 < 1 % and 10 % at low, medium and high feeding levels, respectively.

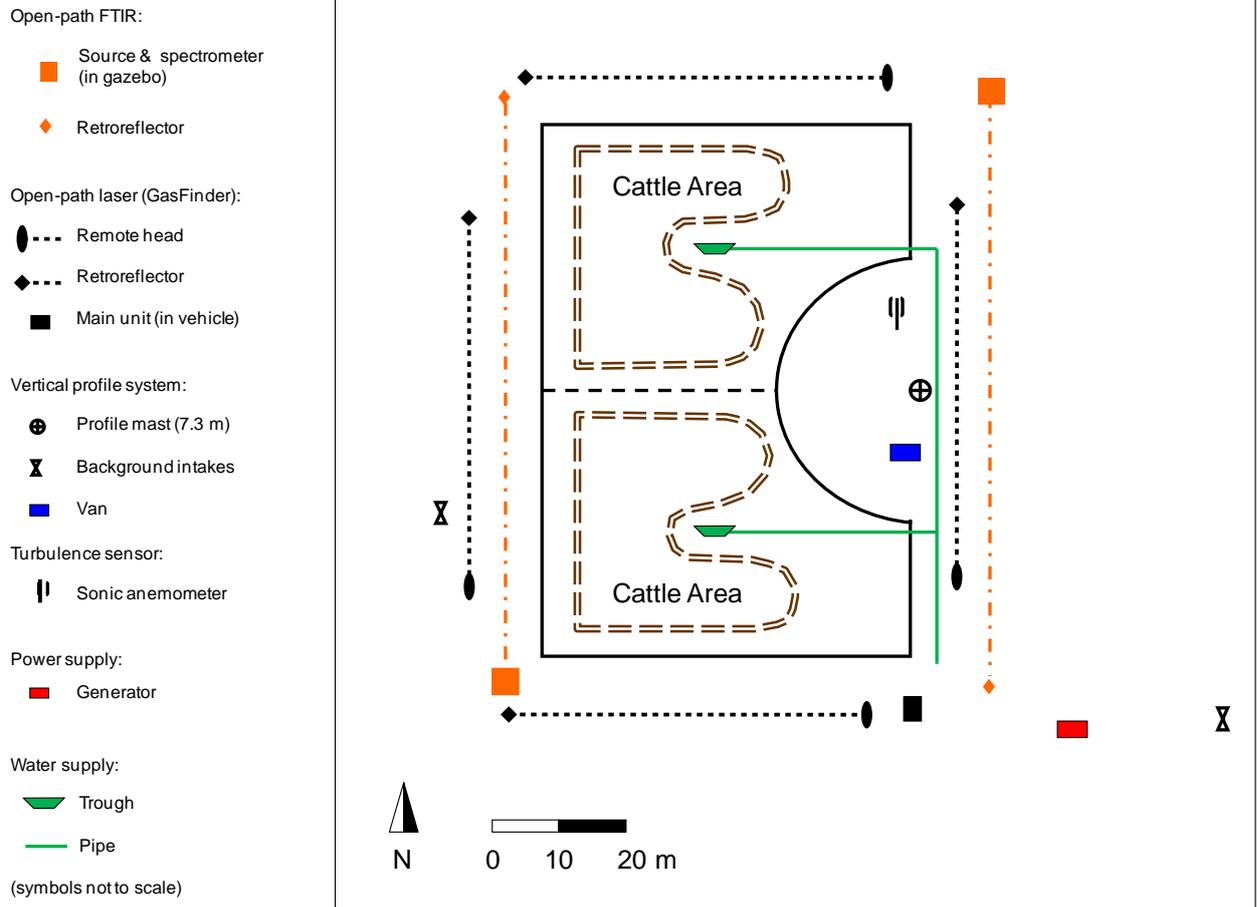
1066 <sup>3)</sup> Methane yield: methane energy as percentage of the gross energy intake (GEI).

1067 <sup>4)</sup> Probability value for feeding level effect; n.d. = not determined.  
 1068  
 1069

1070 **Figures**

1071

1072



1073

1074 **Fig. 1** Schematic of the experimental setup. The cattle area was subdivided into two  
1075 equal-sized enclosures (dashed line), occupied by 30 and 31 animals,  
1076 respectively. Every morning, baleage feed was laid out approximately as  
1077 indicated by the “B”-shaped double-dashed lines.  
1078

1079

1080

1081 a)



1082

1083 b)



1084

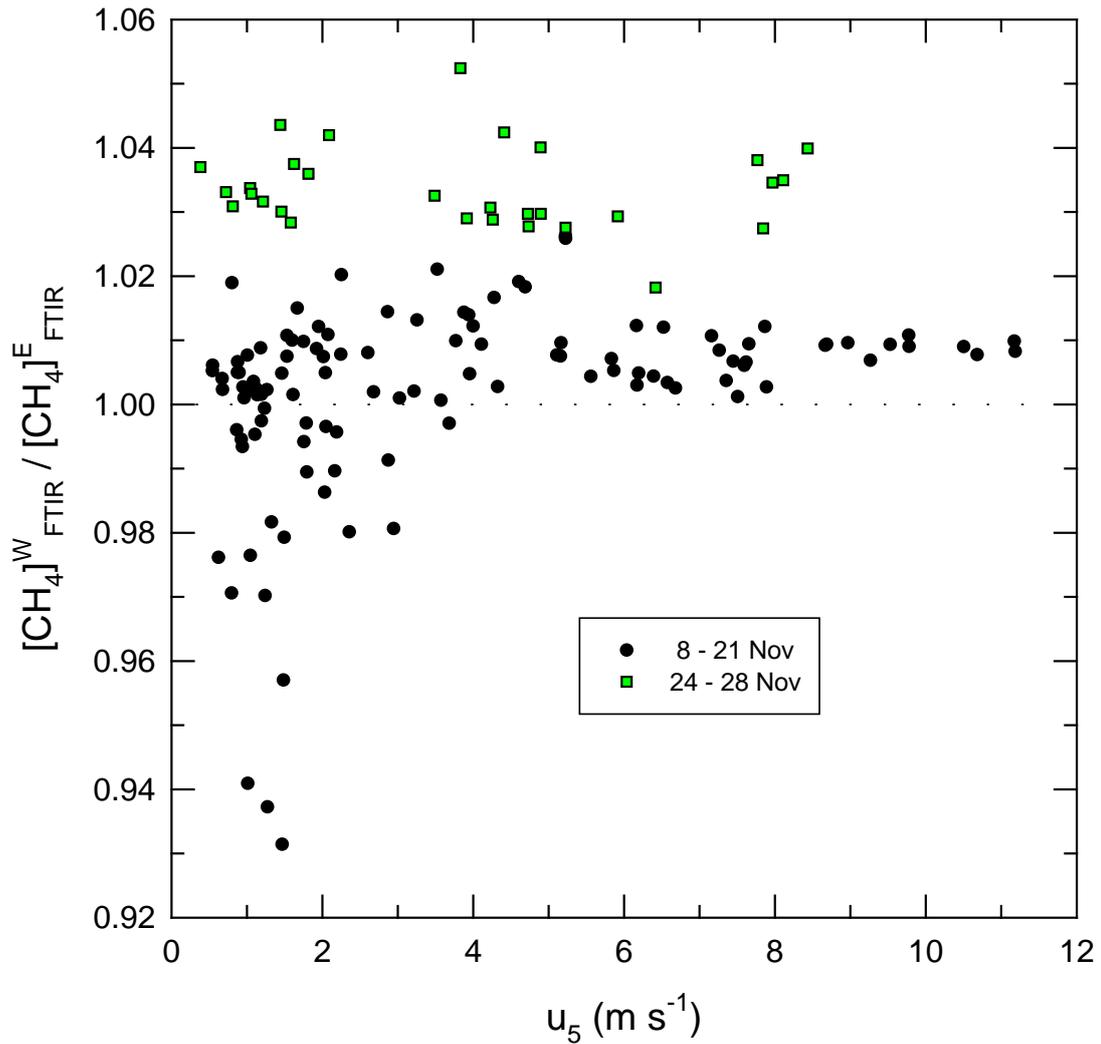
1085 **Fig. 2** The source/detector arrays of the two types of open-path instruments used:  
1086 a) GasFinder MC infrared laser (Boreal Laser Inc.), b) FTIR spectrometer  
1087 (University of Wollongong).  
1088



1090

1091 **Fig. 3** A steer wearing the collection yoke – the white pipe around his neck – for the  
1092 enteric tracer-ratio ( $\text{SF}_6$ ) technique, as well as the  $\text{N}_2\text{O}$  release canister for the  
1093 external tracer-ratio technique, in a grey bag attached to the left front of the  
1094 yoke.  
1095

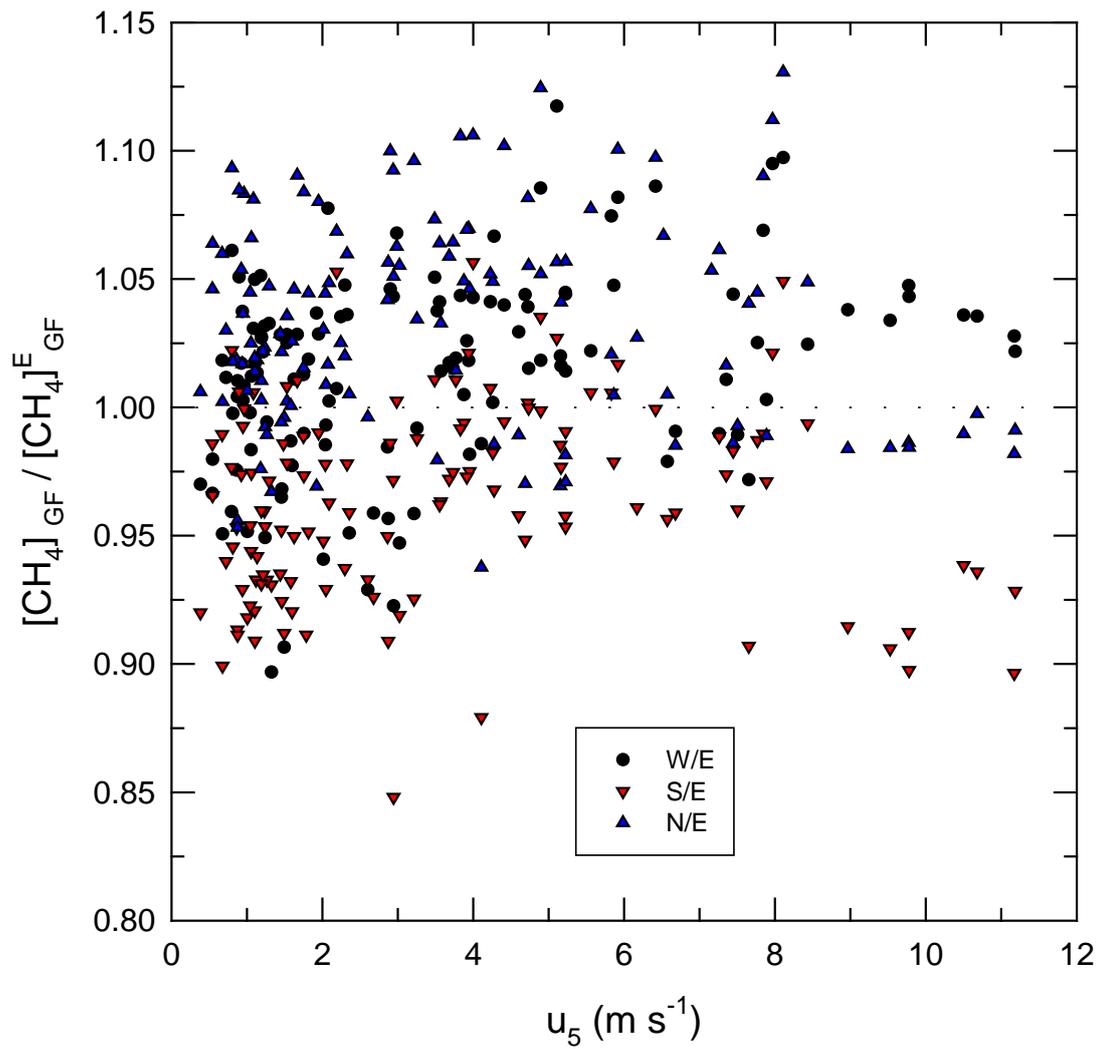
1096



1097

1098 **Fig. 4** Ratio of FTIR-measured uncorrected  $\text{CH}_4$  mole fractions, from W path to E  
 1099 path, for times of cattle absence, against wind speed at 7.18 m. There were no  
 1100 periods of cattle absence on 22 to 23 Nov. From 22 to 24 Nov, the W FTIR  
 1101 experienced an electronics failure. Rectifying this necessitated a realignment,  
 1102 after which some calibration parameters had changed.  
 1103

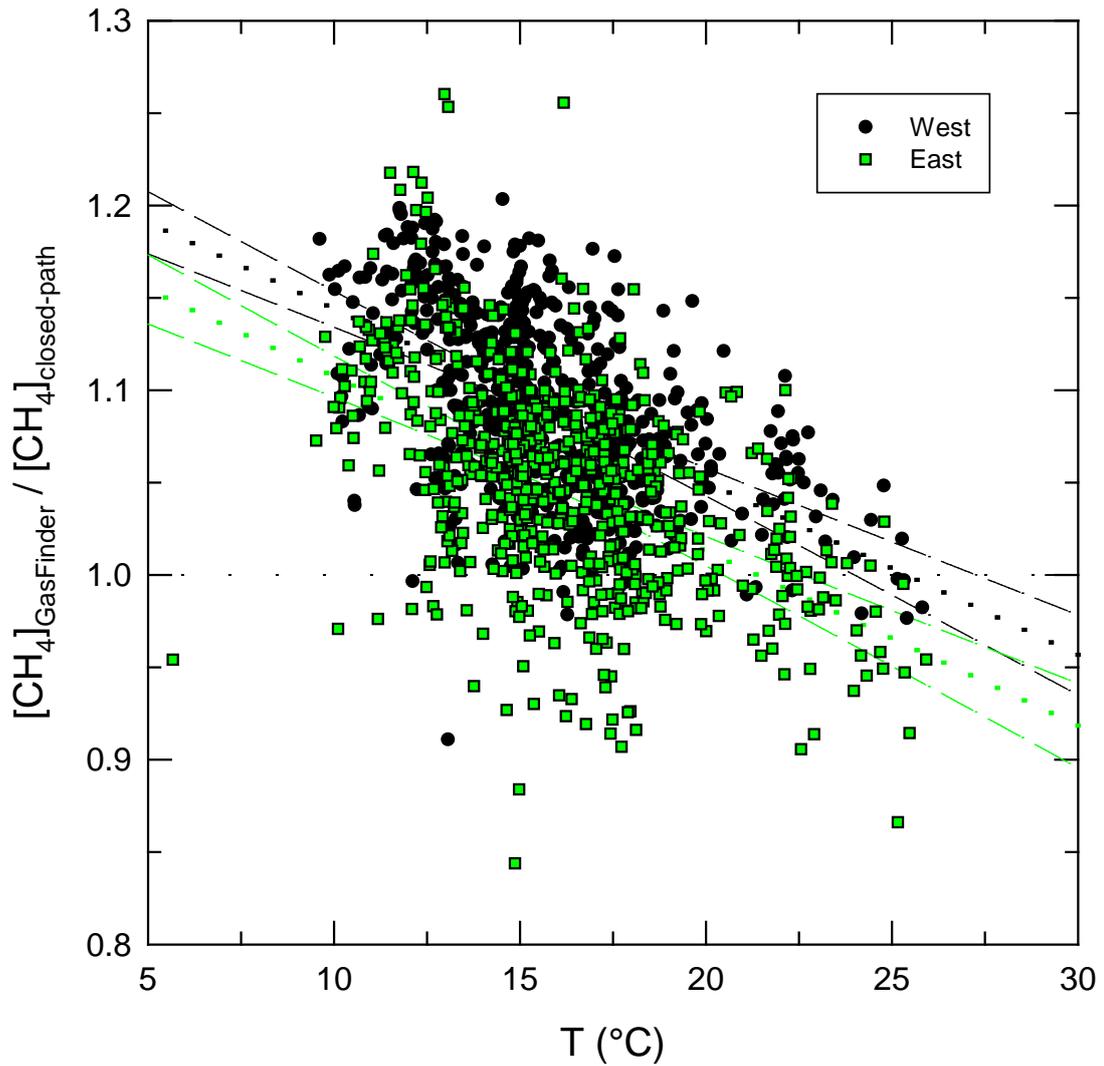
1104



1105

1106 **Fig. 5** Ratios of the open-path laser's  $\text{CH}_4$  mole fractions, from W, S and N path  
 1107 relative to E path, for times of cattle absence, against wind speed at 7.18 m.  
 1108

1109



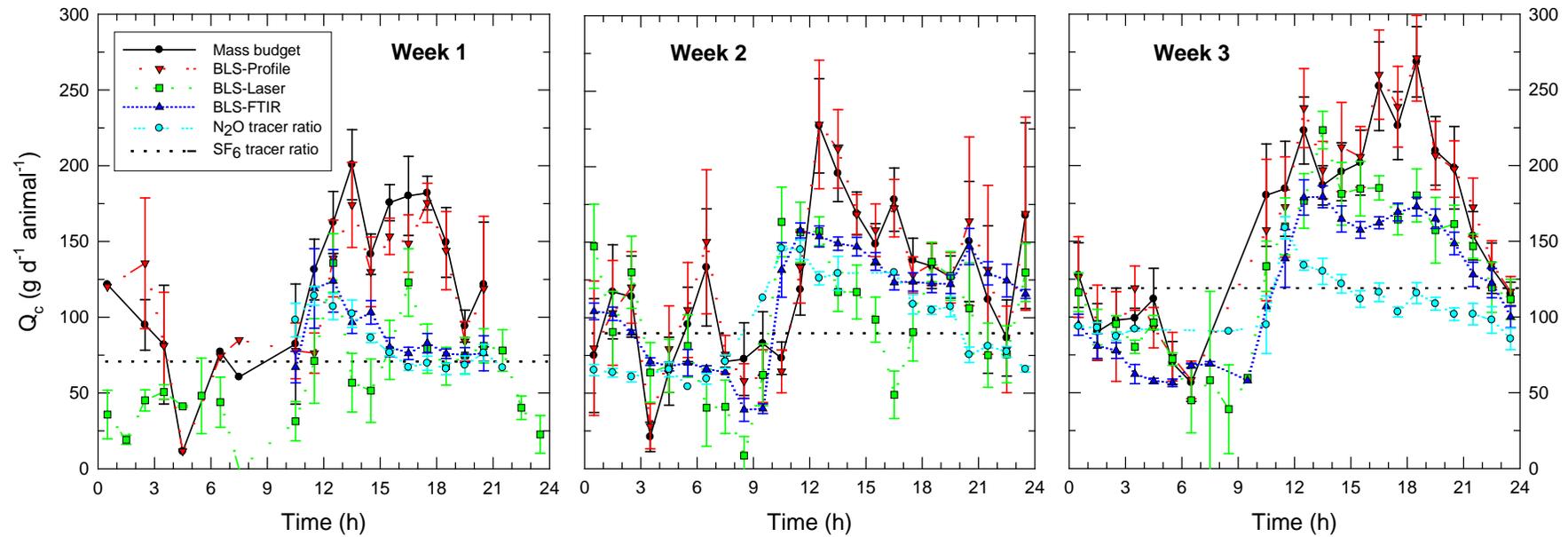
1110

1111 **Fig. 6** Ratio of CH<sub>4</sub> mole fractions from open-path laser (GasFinder) and closed-path  
 1112 analyser (Los Gatos) against temperature, for periods with  $u_5 > 2 \text{ m s}^{-1}$ . Short-  
 1113 dashed lines: linear regression curves separately for W path (dots, 621 points,  
 1114  $R^2 = 0.30$ ) and E path (squares, 649 points,  $R^2 = 0.26$ ), long-dashed lines:  
 1115 99 %-confidence intervals. The regression slope is the same for both,  
 1116  $-9.4 \times 10^{-3} \text{ K}^{-1}$ .

1117

1118

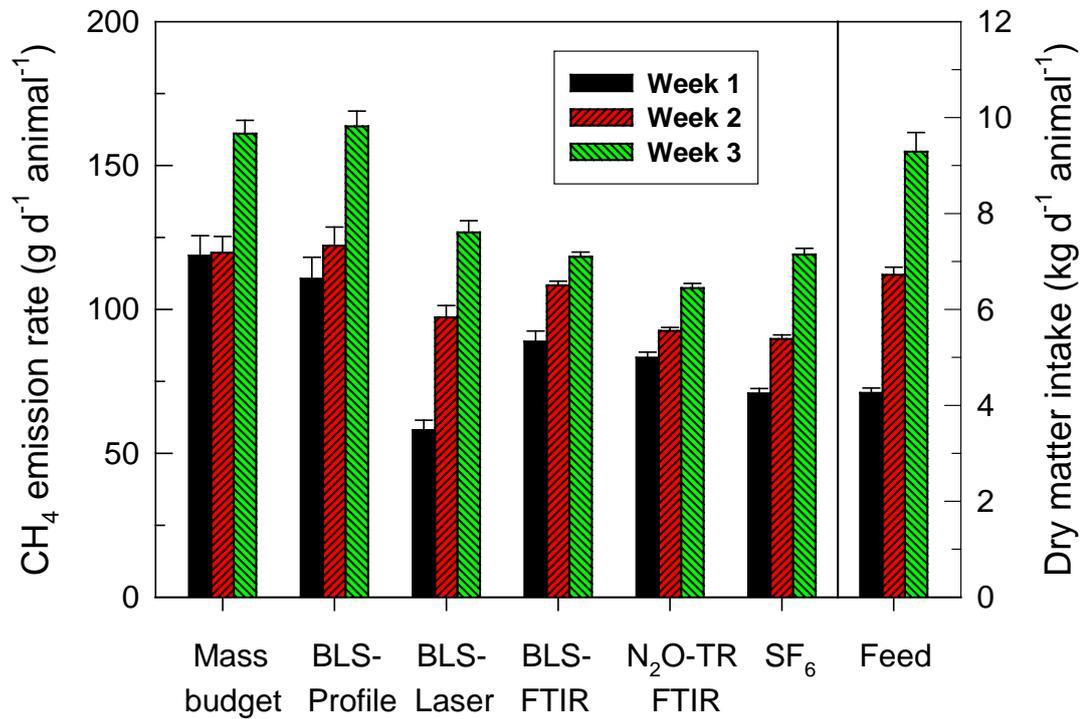
1119



1120

1121 **Fig. 7** Mean diurnal courses of  $\text{CH}_4$  emission rates obtained with the five herd-scale techniques, with error bars indicating standard error  
1122 of the mean for each hourly bin. The three panels are for low, medium and high feeding level, from left to right. The weekly mean  
1123 emission rate from the animal-scale  $\text{SF}_6$  technique is indicated as a horizontal dashed line in each panel.  
1124

1125



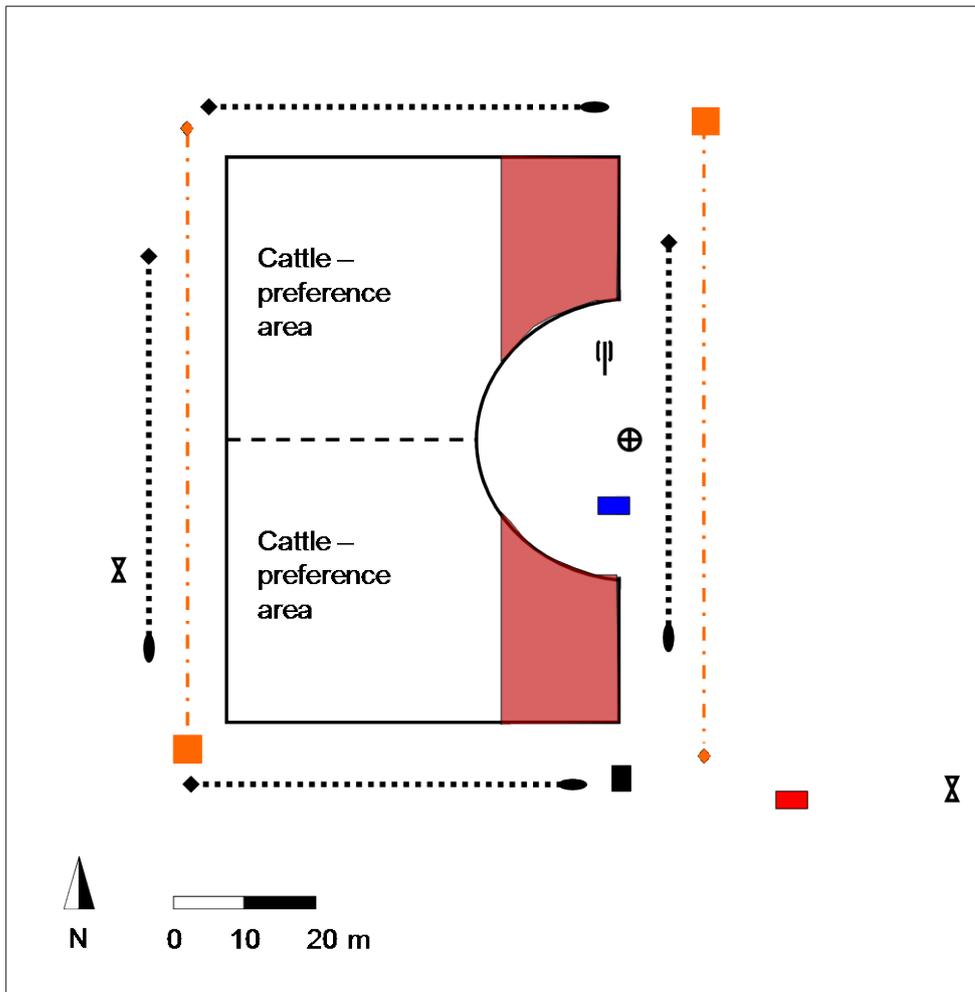
1127

1128 **Fig. 8** Mean weekly CH<sub>4</sub> emission rates for each technique, obtained by averaging the  
 1129 mean diurnal courses of Fig. 7. Error bars indicate propagated errors (from the  
 1130 hourly standard errors) for the five herd-scale techniques, and the standard error  
 1131 of the mean from 61 animals for the animal-scale SF<sub>6</sub> technique. The right-  
 1132 most set of columns shows the feed intake, with estimated errors from  
 1133 weighing and visual assessment of refusal.

1134

1135

1136

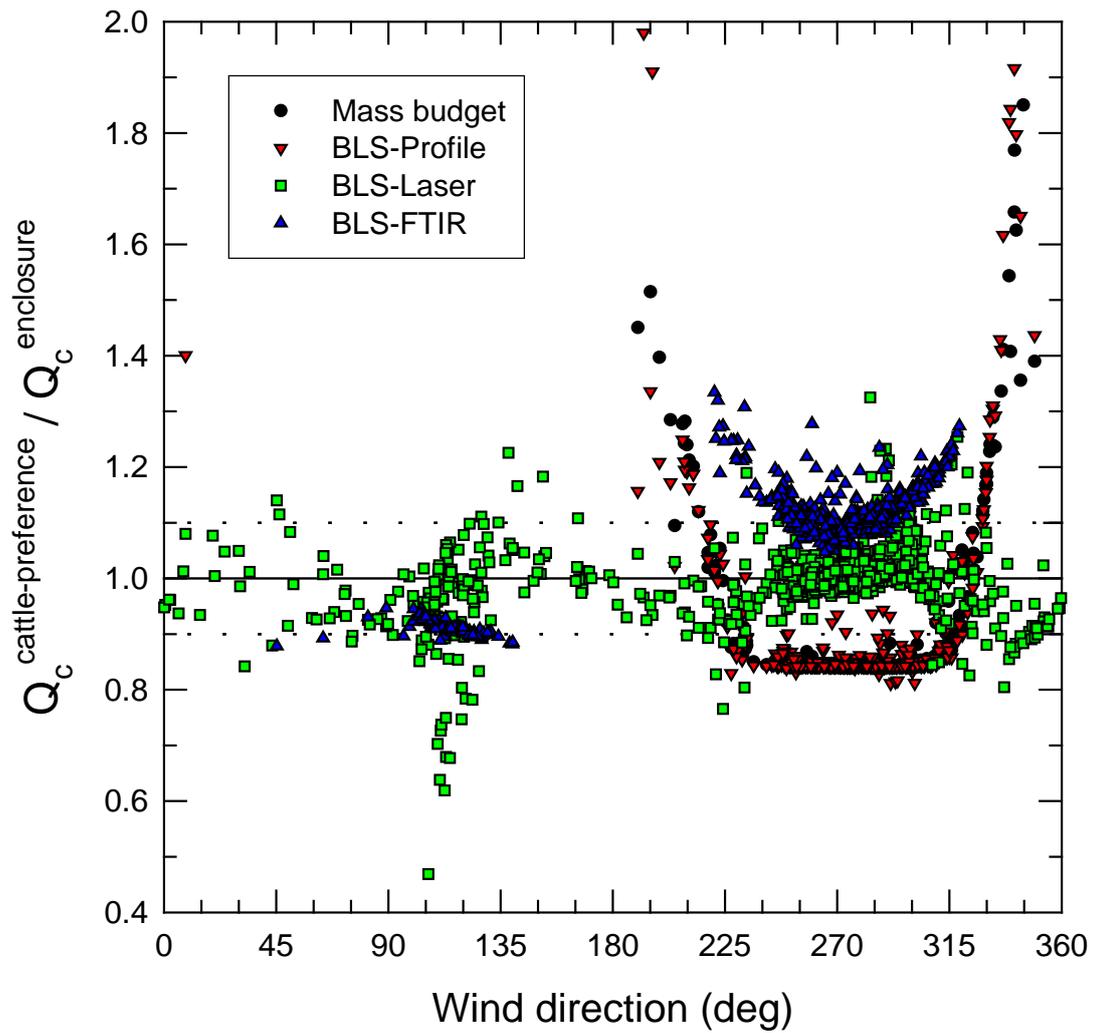


1137

1138 **Fig. 9** Schematic of the experimental setup as in Fig. 1, but with the eastern areas that  
 1139 were rarely visited by the cattle indicated in dark colour. These areas were not  
 1140 considered part of the source area when CH<sub>4</sub> emission rates were re-computed  
 1141 for the “cattle-preference area”.  
 1142

1143

1144



1145

1146 **Fig. 10**  $\text{CH}_4$  emission rates assuming the “cattle-preference area” as the source area  
 1147 divided by  $\text{CH}_4$  emission rates assuming the whole fenced enclosure as the  
 1148 source area, as functions of wind direction, for the four techniques sensitive to  
 1149 source area choice. The dashed lines indicate 10 % deviation from the ideal  
 1150 value 1 (solid line).  
 1151

1152

1153