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Effects of thermal pre-treatment and recuperative thickening on the fate of trace organic contaminants during anaerobic digestion of sewage sludge

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Abstract

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Abstract: This study examined the effects of thermal pretreatment and recuperative thickening on anaerobic digestion of sewage sludge on biogas production and removal of trace organic contaminants (TrOCs). Thermal pre-treatment and recuperative thickening resulted in approximately 15% increase in biogas production. However, the effects of thermal pretreatment and recuperative thickening on anaerobic digestion performance in respect to the removal of TrOCs were less obvious and varied widely depending on the molecular properties of each compound. In total, 16 TrOCs were detected in all primary sludge samples. Removal from the aqueous phase was negligible for most of these 16 TrOCs. Caffeine and paracetamol were the only two TrOCs with a high removal from the aqueous phase. In comparison to the aqueous phase, TrOC removal from the solid phase was considerably higher. Through a mass balance calculation, it was shown that thermal pre-treatment or a combination of thermal pre-treatment and recuperative thickening could enhance the biodegradation of five persistent TrOCs, namely TCEP, verapamil, clozapine, triclosan, and triclocarban by 17 to 50%.

Keywords: Anaerobic digestion, thermal pre-treatment, recuperative thickening, biogas, traces organic contaminants, biodegradation.

1. Introduction

Urbanization and continuous population growth have imposed an increasing demand on wastewater treatment plants (WWTPs) particularly in regard to the management of sewage sludge. In Australia, sewage sludge production (as dried solids) has increased from 0.30 to 0.33 million tonnes between 2010 to 2013 (Semblante et al., 2014). Sewage sludge contains biodegradable organics and an array of pathogens. Thus, sewage sludge treatment is necessary before any beneficial use or land disposal. Anaerobic digestion is currently the most widely used technique for sewage sludge treatment. Anaerobic digestion is a biological process in which microorganisms convert biodegradable materials in the absence of oxygen to biogas and more stable organics. It is well established that anaerobic digestion can efficiently stabilise organic materials and remove pathogenic agents in sewage sludge while simultaneously producing valuable biogas (Sawatdeenarunat et al., 2016; Sihuang et al., 2016; Tuyet et al., 2016). Biogas is a form of renewable fuel, which can be used to generate electricity and heat (Nghiem et al., 2017). The remaining and more stable solids are rich in nutrient and organics, thus, can be used for soil amendment (Nghiem et al., 2017).

Anaerobic digestion consists of four stages with hydrolysis being the first during which organic materials are transformed to fatty acids and other soluble organic compounds (Habiba et al., 2009). Since hydrolysis is the rate limiting step during anaerobic digestion, several pre-treatment methods, including thermal hydrolysis, biological treatment, ultrasonication, and ozonation, have been suggested to increase the digestion rate or improve the inherent degradability of sewage sludge (Carrère et al., 2010; Dhar et al., 2012). Thermal hydrolysis is a promising pre-treatment method to improve methane production during anaerobic processing ([Supplementary data Table S1](#)) since complex organic molecules can be transformed into short-chain fragments better suited for biological digestion (Liao et al., 2016; Mottet et al., 2009; Schieder et al., 2000). The effects of thermal pre-treatment at

temperature of up to 180 °C and duration up to 2 hours on anaerobic digestion performance have been evaluated by several recent studies (Bougrier et al., 2008; Kim et al., 2003; Pérez-Elvira and Fdz-Polanco, 2012; Phothilangka et al., 2008; Valo et al., 2004). The optimal temperature of thermal hydrolysis was reported to be 150-180 °C by Bougrier et al. (2008) for a pre-treatment duration of 30 to 60 minutes. Thermal hydrolysis has been successfully used at a full scale wastewater treatment plant (Kepp et al., 2000). The energy balance calculation showed the net electricity production due to enhanced biogas production increased by over 20%, which is more than the energy input for thermal hydrolysis.

In addition to thermal pre-treatment, recuperative thickening has also been identified as a potentially cost-effective and readily implementable method to improve anaerobic digester performance without the need to increase the size of the digester (Cobbledick et al., 2016). Recuperative thickening can increase the solids retention time (SRT) independently of the hydraulic retention time (HRT) by thickening a proportion of the digestate to remove water and then returning the thickened sludge back to the digester (Reynolds et al., 2001; Torpey and Melbinger, 1967; Yang et al., 2015). The increase in SRT helps to improve the conversion of organics to methane and increase the volatile solid (VS) reduction (Sieger et al., 2004; Yang et al., 2015). Recuperative thickening has been successfully applied in a few WWTPs in North America and Australia. Full scale monitoring data suggest that recuperative thickening can improve both biogas production and VS reduction by 15-30% (Greer, 2011; Reynolds et al., 2001).

A major issue associated with beneficial reuse of reclaimed water and biosolids from sewage treatment is the ubiquitous occurrence of trace organic contaminants (TrOCs) in municipal wastewater. These TrOCs include several groups of widely used compounds including pharmaceuticals and personal care products, steroid hormones, industrial chemicals, pesticides, phytoestrogens, and UV filters. Their toxicological effects on human and other

biota even at a very low concentration (less than 1 µg/L) remain largely unknown but are generally suspected (Luo et al., 2014). Some TrOCs can partition from the aqueous phase in wastewater to the solid phase in sludge during wastewater treatment (Citulski and Farahbakhsh, 2010; Semblante et al., 2015). When applied to farm land, these TrOCs may accumulate in soil, presenting a potential risk to human health and the ecosystem (Citulski and Farahbakhsh, 2010). However, to date, there have been only a few investigations on the removal of TrOCs from sewage sludge by anaerobic treatment.

Of a particular note, little is known about the impact of pre-treatment on the removal of TrOCs from sewage sludge by anaerobic digestion. In a systematic lab-scale study, McNamara et al. (2012) observed no discernible impact of thermal hydrolysis on the degradation of nonylphenol ethoxylates by anaerobic digestion. Similarly, Carballa et al. (2006) reported that thermal pre-treatment of sewage sludge had no observable impact on the removal of several pharmaceuticals, musks, and steroid hormones. By contrast, Hamid and Eskicioglu (2013) observed a notable increase in the removal of estrone and estradiol by anaerobic treatment following microwave-assisted pre-treatment (80 to 160 °C, 2.45 GHz, 1200 W). Given the paucity of information on this important issue, the present study aims to evaluate the influence of thermal hydrolysis and recuperative thickening on the fate of TrOCs in sewage sludge during anaerobic digestion. The influence of thermal hydrolysis and recuperative thickening on anaerobic digestion performance in terms of biogas production and organics removal is also investigated.

2. Materials and Methods

2.1 Lab-scale anaerobic digester and sludge

Three lab-scale anaerobic digesters previously described by Yang et al. (2016) were used in this study (Fig 1). Briefly, each digester consisted of a 28 L stainless steel reactor (Core

Brewing Concepts, Victoria, Australia), a peristaltic hose pump (DULCO®flex from ProMinent Fluid Controls, Australia), a temperature control unit (Neslab RTE 7), a thermal couple with temperature gauge, a biogas counter, and a gas trap for biogas sampling. One digester (denoted as D1) was operated as the control system without thermal pre-treatment and recuperative thickening. One digester (denoted as D2) was operated with thermal pre-treatment. The last reactor (denoted as D3) was operated with both thermal pre-treatment and recuperative thickening. All three reactors were operated in parallel and were each seeded with 20 L anaerobically digested sludge. The digested sludge and primary sludge were all sampled from a full scale wastewater treatment plant in New South Wales, Australia, with average total solid (TS) content of 29.0 ± 1.0 g/L and 22.2 ± 2.2 g/L, respectively. All sludge samples were stored at 4 °C until use or else discarded within two weeks.

Fig 1: The schematic diagram of the three lab-scale anaerobic digesters.

All anaerobic digesters were operated under the same HRT of 20 d by wasting 1 L of digestate and the feeding with 1 L of primary sludge each day. Raw sludge, thermally

pretreated sludge, and digested sludge were collected weekly for characterisation. Parameters that were regularly measured include TS, volatile solid (VS), chemical oxygen demand (COD), alkalinity and pH.

2.2 Thermal pre-treatment

The feed sludge to digester D2 and D3 was thermally pretreated at 150 °C and 500 kPa for 30 minutes using a New Tek Machinery pressure vessel (Changzhou, China) with a heating jacket. At the conclusion of the process, the pressure inside the vessel was released and the sludge was allowed to cool to room temperature (ca. 25 °C) before feeding to the digester.

2.3 Recuperative thickening

Digester D3 was operated with recuperative thickening to achieve an SRT of 30 d with the HRT at 20 d i.e., same as the other digesters. A thickening ratio of 1.33 (which is the ratio of the total TS from primary sludge feed and return thickened sludge over the TS from primary sludge feed) was used. Each day, 2 L of the digestate was withdrawn from digester D3 and dosed with thickening polymer (Zetag 8169, BASF) at 7.5 g/Kg dry sludge. The sludge was gently mixed and allowed to settle by gravity for at least 10 minutes. 1 L of thickened sludge was then mixed with the thermally pretreated (150 °C, 30 min) primary sludge (1 L) to form 2 L of feed to return to the digester. The excess thickened sludge and supernatant were discarded.

2.4 Analytical methods

2.4.1 Anaerobic digestion performance

Biogas production rate was monitored daily by a custom-made gas counter (Yang et al., 2016). The biogas composition was detected weekly by a portable gas analyser (GA5000 gas analyser, Geotechnical Instruments Ltd, UK) (Nghiem et al., 2014). Additionally, samples

from primary sludge (before and after thermal treatment) and digested sludge were taken weekly to analyse sludge characters such as TS, VS, total COD (tCOD), soluble COD (sCOD), pH and alkalinity. TS, VS, and alkalinity were measured in accordance to the standard methods. COD was measured following the US-EPA Method 8000 using high range plus COD vials (HACH, USA). The supernatant used for measurement of sCOD was obtained by centrifuging sludge sample at 3720xg for 10 minutes (Allegra X-12R centrifuge, Beckman Coulter, Australia), and then filtering through 1 μ m glass microfiber filter paper (Filtech, Australia).

2.4.2 TrOC sample preparation and analysis

Primary and digested samples were collected every 7 days to prepare duplicate samples for TrOC concentration analysis ([Wijekoon et al., 2015](#)). Sludge samples were centrifuged at 3720xg for 10 minutes (Allegra X-12R, Beckman Coulter, USA) to obtain solid pellets and supernatant for further processing. Supernatant from sludge sample (50 mL) was diluted with Milli-Q to 500 mL. Then the obtained aqueous samples were filtered by 1 μ m and then 0.7 μ m pore size glass fiber filter paper. The filtered samples were spiked with surrogate (50 μ L per sample) containing 40 isotopically labelled standards for method recovery and determine TrOC concentration before proceeding to solid phase extraction (SPE). During the SPE, HLB cartridges were conditioned with 5 mL methyl tert-butyl ether, 5 mL methanol, and 2 x 5 mL Milli-Q water before the liquid samples were loaded to the cartridges at the flow rate of approximately 15 mL/min. After concentrating to 1 mL, eluted samples were subjected to high performance liquid chromatography-tandem mass spectrometry analysis (HPLC-MSMS) ([Alturki et al., 2013](#)). In this study, a spectrum of 40 TrOCs was used to prepare the surrogate and screen the TrOC concentration of sludge samples.

The solid pellets were freeze-dried using an Alpha 1-2 LDplus Freeze dryer (Christ GmbH, Germany). The dried samples were ground to powder and 0.5 g was transferred to a 13 mL glass vial (with cap) for extraction. Methanol (10 mL) was added to the vial, mixed with the powder by a vortex mixer (VM1, Ratek, Australia), and ultrasonicated for 10 minutes at 40 °C. The solution was then centrifuged at 3720xg for 10 minutes to obtain a supernatant. The residual solid was extracted using 10 mL solvent made of dichloromethane and methanol (1:1, v/v) by repeating the previous steps. The supernatants from these steps were combined and diluted to 500 mL by Milli-Q water. The liquid samples were then filtered, spiked with surrogate, loaded to the SPE cartridges and analysed following the same procedure for sludge supernatant samples described before.

2.4.3 TrOC mass balance

Mass balance calculations were conducted for each TrOC to determine their fate in the aqueous and solid phase (Wijekoon et al., 2015). The total mass of each TrOC fed into the system can be described as:

$$M_{in} = X_{in} \times TS_{PS} + S_{in} \quad (1)$$

where M_{in} is the total mass of TrOC in 1 L of feed (ng), X_{in} is the TrOC concentration in the solid phase of primary sludge (ng/g dry sludge), TS_{PS} is the total solid concentration of primary sludge (g/L), and S_{in} is the TrOC concentration in the aqueous phase of primary sludge (ng/L). The mass of TrOC (M_{aq}) in the aqueous phase in 1 L of the digestate can be measured experimentally. The mass of TrOC in the solid phase of the digestate can be described as:

$$M_{solid} = X_{solid} \times TS_{DS} \quad (2)$$

where M_{solid} is the mass of TrOC in the solid phase (ng), X_{solid} is the TrOC concentration in the solid phase of digested sludge (ng/g dry sludge), TS_{DS} is the total solid concentration of digested sludge (g/L). Thus the mass balance for TrOC concentration can be presented as

$$M_{in} = M_{aq} + M_{solid} + M_{bio} \quad (3)$$

where M_{bio} is the mass of TrOC that has been biodegraded.

3. Results and discussion

3.1 Thermal pre-treatment and recuperative thickening

Thermal pre-treatment and recuperative thickening (Digester D3) resulted in approximately 15% increase in biogas production in comparison to the control digester (D1) (Fig 2). The combination of thermal pre-treatment and recuperative thickening (Digester D3) did not lead to any additional increase in biogas production compared to only thermal pre-treatment (D2). According to Pilli et al., (2015), thermal pre-treatment causes the disintegration and solubilisation of some solid sludge particles, thus, enhancing the hydrolysis step and hence biogas production. Indeed, in this study, in which approximately 10% of the tCOD of primary sludge was converted to sCOD after thermal treatment. On the other hand, recuperative thickening can extend the residence time of sludge in the reactor and recapture soluble macro-organic molecules for further digestion. Biogas production-increase by up to 30% has been reported in previous laboratory scale and full scale studies (Cobbledick et al., 2016; Reynolds et al., 2001). Results from Fig 2 suggest that the benefits of thermal pre-treatment and recuperative thickening are mutually exclusive. It is also noteworthy that thermal pre-treatment and recuperative thickening did not exert any observable impact on biogas composition. Throughout this study, biogas composition from all three digesters was stable with approximately 60% CH₄ and 40% CO₂.

Fig 2: Average biogas production from digester D1 (Control), D2 (Thermal pre-treatment (TP)) and D3 (Thermal pre-treatment and recuperative thickening (TP+RT)). Error bars show the standard deviation of 7 measurements (one per week).

The sludge composition varied quite significantly throughout the course of this study. Since organic removal in terms of TS, VS, tCOD and sCOD was determined on a weekly basis, there were some notable variations. TS and VS removals ranged from 50 to 80% and 70 to 90%, respectively. Due to these significant variations in TS and VS, the effects of thermal pre-treatment and recuperative thickening were not observable in this study. Nevertheless, some enhancement in the removal of tCOD and sCOD could be observed in Fig 3. With the exception of day 49, the removal of tCOD by Digester 2 (thermal pre-treatment) and Digester 3 (thermal pretreatment and recuperative thickening) was comparable or higher than that of the control digester (Fig 3a).

Fig 3: (a) tCOD removal and (b) sCOD removal by the control digester (D1), digester 2 with thermal pre-treatment (TP), and digester 3 with thermal pre-treatment and recuperative thickening (TP+RT).

The effect of pre-treatment and recuperative thickening on removal performance was most notable in terms of sCOD removal. Digester D2 showed comparable sCOD removal to that by the control digester (D1). On the other hand, digester D3 showed a notable increase in sCOD removal (Fig 3b). As noted above, thermal pre-treatment led to the solubilisation of some tCOD into sCOD. On the other hand, due to sludge thickening, soluble organics can be retained for further digestion. Thus, recuperative thickening could improve the removal of sCOD.

Several other parameters including pH and alkalinity were also monitored. The mixed liquor pH value of all three digesters was stable between 7.0 – 7.5 and the alkalinity was over 2600 mg CaCO₃/L (Supplementary data Fig S4). These results confirm stable operation of all three digesters in this study.

3.2 TrOC occurrence in primary sludge

243 In good agreement with a previous study by Yang et al. (2017), of the 40 TrOCs monitored in
244 this study, 16 compounds were prevalently detected in all primary sludge samples (Fig 4).
245 The concentrations in the aqueous and solid phase were in the range from 50 to 40,000 ng/L
246 and from 20 to nearly 9,000 ng/g dry sludge, respectively. The occurrence of these TrOCs in
247 primary sludge is well related to their usage in daily life. For examples, caffeine (which is a
248 stimulant in coffee and tea) and paracetamol (which is a widely used pain killer) were
249 detected at the highest concentration in the aqueous phase (40,000 and 38,000 ng/L,
250 respectively). At the TS content of 29 g/L, it can also be inferred from Fig 4 that these TrOCs
251 occurred mostly in the solid phase (i.e. 70 to 100% in the total mass in primary sludge).
252 Caffeine and ibuprofen are the only two exceptions. The mass distributions of caffeine and
253 ibuprofen in the solid phase were 24 and 41%, respectively, possibly because of their
254 hydrophilicity. These results highlight the need for specific investigation of the removal of
255 TrOCs from the solid phase and that data from previous studies considering only the aqueous
256 phase may not be valid in the context of anaerobic digestion.

Fig 4: TrOC concentrations in (a) aqueous phase and (b) solid phase of primary sludge. 12 samples were taken during the experimental period.

3.3 TrOC removal in the aqueous and solid phase

TrOC concentrations in the aqueous and solid phase of the feed and digestate from the three reactors are shown in Figs. 4 and 5, respectively. In these Figs, the TrOCs were listed in the order of increasing hydrophobicity. Under all experimental conditions, caffeine and paracetamol were almost completely removed (98 – 99%) from the aqueous phase (Fig 5). Moderate removals from the aqueous phase were observed for trimethoprim and amitriptyline, especially when pre-treatment and recuperative thickening were applied together (D3). However, all other TrOCs were not significantly removed from the aqueous phase as can be observed with all three digesters (Fig 5). In fact, in the case of ibuprofen, gemfibrozil, and diuron, their concentrations in the aqueous phase of the digestate (after anaerobic treatment) were even higher than the corresponding values of the feed primary sludge (Fig 5). It is possible that the anaerobic condition could facilitate the transfer of some TrOCs from the solid to aqueous phase. This is probably because of the transfer of TrOCs from the solid phase to the aqueous phase during anaerobic digestion. It is also noteworthy from section 3.1 that most of these TrOCs are in the solid phase.

TrOC removal from the solid phase was notably higher in comparison to that from the aqueous phase. As can be seen in Fig 6, several hydrophilic TrOCs including caffeine, sulfamethoxazole, trimethoprim and paracetamol were well removed from the solid phase by anaerobic digestion. The hydrophilicity of compounds appears to be an important factor for their high removal from solid phase since hydrophilic compounds would easily desorb from sludge granules. However, similar to the removal from aqueous phase, there is no obvious evidence that thermal pre-treatment and recuperative thickening could improve the removal of all of these TrOCs from the solid phase (Fig 6).

Fig 5: Average concentrations of TrOCs in aqueous phase of primary sludge (PS), digested sludge from digester D1 (Control), D2 (TP) and D3 (TP+RT) (mean $\pm$ standard deviation of 12 samples).

Several previous studies have also shown no discernible changes in TrOC removal after thermal pre-treatment. For example, McNamara et al. (2012) reported that nonylphenol, diethoxylate and nonylphenol monoethoxylate were not removed from the influent by anaerobic treatment with and without thermal treatment (150 °C, 2 h). Similarly, Carballa et al. (2006) also reported that thermal pre-treatment of mixed sludge by autoclaving at 130 °C for 1 h had no impact on the removal of various pharmaceuticals, musks, and hormones by anaerobic treatment. However, it is noteworthy that these previous studies focused on the anaerobic treatment of wastewater and only considered the aqueous phase. Thus, their results cannot readily correlate to the anaerobic digestion of wastewater sludge. As discussed above, during anaerobic digestion of sludge, the transfer of TrOCs between the aqueous and solid

phase can influence the overall removal efficiency. Thus, it is important to conduct a mass balance to elucidate the contribution of biodegradation and the fate of TrOCs in the aqueous and solid phase.

Fig 6: Average concentrations of TrOCs in solid phase of primary sludge (PS), digested sludge from digester D1 (Control), D2 (TP) and D3 (TP+RT) (mean $\pm$ standard deviation of 12 samples).

3.4 Fate of TrOCs during anaerobic digestion

Fig 7 shows the fate of each TrOC amongst the three possible domain namely biodegradation, partitioning to the solid phase, and partitioning in the aqueous phase. Several readily biodegradable TrOCs can be identified from Fig 7. They include caffeine, sulfamethoxazole, trimethoprim and paracetamol (Fig 7). Likewise, four TrOCs including ibuprofen, carbamazepine, diuron and clozapine were not biodegraded under any experimental conditions in this study (Fig 7).

It has been established that a compound's molecular structure is a major factor governing their degradability (Tadkaew et al., 2011; Wijekoon et al., 2015; Yang et al., 2016). TrOCs with strong electron donating functional groups (Supplementary data Table S6) such as amine (caffeine, sulfamethoxazole and trimethoprim), amino (paracetamol and sulfamethoxazole), hydroxyl (paracetamol) and ether (trimethoprim) are known to be readily biodegradable. On the other hand, TrOCs with strong electron withdrawing functional groups tend to be persistent to biological treatment. Examples of these electron withdrawing functional groups are carboxyl (gemfibrozil and ibuprofen), amide group (carbamazepine), and chloro (diuron). Indeed, as can be seen in Fig 7, all TrOCs with electron withdrawing functional groups were not effectively biodegraded.

Results from this study are consistent with several previous studies. Caffeine (Narumiya et al., 2013; Yang et al., 2016), trimethoprim (Malmborg and Magnér, 2015; Narumiya et al., 2013) and sulfamethoxazole (Carballa et al., 2007; Narumiya et al., 2013) have been reported to be well removed by anaerobic digestion. By contrast, carbamazepine (Carballa et al., 2007; Malmborg and Magnér, 2015; Narumiya et al., 2013), diuron (Carballa et al., 2007; Tadkaew et al., 2011) and ibuprofen (Alvarino et al., 2014; Malmborg and Magnér, 2015) were resistant to anaerobic digestion.

Of particular note, enhanced biodegradation due to either thermal pre-treatment and/or recuperative thickening was observed with five TrOCs (denoted in Fig 7 with #). The biodegradation of triclosan and triclocarban were improved by approximately 10% due to thermal pre-treatment (Fig 6a and b) and further improved (by about 15%) when recuperative thickening was also applied (Fig 6c). Verapamil and clozapine were approximately 20% more biodegraded when both thermal pre-treatment and recuperative thickening were applied (Fig 6a and c). However, with thermal pre-treatment and recuperative thickening, TCEP biodegradation increased to approximately 40% and 60%, respectively.

Fig 7: Overall fate of each compound by anaerobic digestion in digester (a) D1 (Control), (b) D2 (TP) and (c) D3 (TP+RT).

The positive impact of thermal pre-treatment and recuperative thickening does not seem to be governed by the compound hydrophobicity. Indeed, of the 16 TrOCs in Fig 7, TCEP is highly hydrophilic while triclosan and triclocarban are the most hydrophobic. The removal of TrOCs with electron withdrawing functional groups (thus these TrOCs are inherently persistent to biodegradation) is likely to benefit from thermal pre-treatment and recuperative thickening. These TrOCs have at least one electron withdrawing functional group in their molecular structure and are known to be persistent to biodegradation.

4. Conclusions

The effects of thermal pretreatment and recuperative thickening on anaerobic digestion performance were examined in terms of biogas production and the removal of trace organic contaminants (TrOCs). Thermal pre-treatment and recuperative thickening resulted in approximately 15% increase in biogas production. In total, 16 TrOCs were detected in all primary sludge samples. The effects of thermal pretreatment and recuperative thickening on TrOC removal varied significantly. Removal from the aqueous phase was negligible for most of the 16 TrOCs detected in the primary sludge samples. Caffeine and paracetamol were the only two TrOCs with an appreciable level of removal from the aqueous phase. In comparison to the aqueous phase, TrOC removal from the solid phase was considerably higher. Through a mass balance calculation, it was shown that thermal pre-treatment or a combination of thermal pre-treatment and recuperative thickening could enhance the biodegradation of five persistent TrOCs, namely TCEP, verapamil, clozapine, triclosan, and triclocarban by 17 to 50%.”

5. References

- Alturki, A.A., McDonald, J.A., Khan, S.J., Price, W.E., Nghiem, L.D., Elimelech, M., 2013. Removal of trace organic contaminants by the forward osmosis process. *Separation and Purification Technology* 103, 258-266.
- Alvarino, T., Suarez, S., Lema, J.M., Omil, F., 2014. Understanding the removal mechanisms of PPCPs and the influence of main technological parameters in anaerobic UASB and aerobic CAS reactors. *Journal of Hazardous Materials* 278, 506-513.
- Bougrier, C., Delgenès, J.P., Carrère, H., 2008. Effects of thermal treatments on five different waste activated sludge samples solubilisation, physical properties and anaerobic digestion. *Chemical Engineering Journal* 139, 236-244.
- Carballa, M., Omil, F., Lema, J.M., Alder, A.C., 2006. Comparison between the conventional anaerobic digestion of sewage sludge and its combination with a chemical or thermal pre-treatment concerning the removal of pharmaceuticals and personal care products. *Water Science and Technology* 53, 109-117.
- Carballa, M., Omil, F., Ternes, T., Lema, J.M., 2007. Fate of pharmaceutical and personal care products (PPCPs) during anaerobic digestion of sewage sludge. *Water Research* 41, 2139-2150.
- Carrère, H., Dumas, C., Battimelli, A., Batstone, D.J., Delgenès, J.P., Steyer, J.P., Ferrer, I., 2010. Pretreatment methods to improve sludge anaerobic degradability: A review. *Journal of Hazardous Materials* 183, 1-15.
- Citulski, J.A., Farahbakhsh, K., 2010. Fate of Endocrine-Active Compounds during Municipal Biosolids Treatment: A Review. *Environmental Science & Technology* 44, 8367-8376.

389 Cobbledick, J., Aubry, N., Zhang, V., Rollings-Scattergood, S., Latulippe, D.R., 2016. Lab-
390 scale demonstration of recuperative thickening technology for enhanced biogas production
391 and dewaterability in anaerobic digestion processes. *Water Research* 95, 39-47.

392 Dhar, B.R., Nakhla, G., Ray, M.B., 2012. Techno-economic evaluation of ultrasound and
393 thermal pretreatments for enhanced anaerobic digestion of municipal waste activated sludge.
394 *Waste Management* 32, 542-549.

395 Greer, D., 2011. Municipal and industry synergies boost biogas production. *BioCycle* 52, 43.

396 Habiba, L., Hassib, B., Moktar, H., 2009. Improvement of activated sludge stabilisation and
397 filterability during anaerobic digestion by fruit and vegetable waste addition. *Bioresource*
398 *Technology* 100, 1555-1560.

399 Hamid, H., Eskicioglu, C., 2013. Effect of microwave hydrolysis on transformation of
400 steroidal hormones during anaerobic digestion of municipal sludge cake. *Water Research* 47,
401 4966-4977.

402 Kepp, U., Machenbach, I., Weisz, N., Solheim, O.E., 2000. Enhanced stabilisation of sewage
403 sludge through thermal hydrolysis - three years of experience with full scale plant. *Water*
404 *Science and Technology* 42, 89-96.

405 Kim, J., Park, C., Kim, T.-H., Lee, M., Kim, S., Kim, S.-W., Lee, J., 2003. Effects of various
406 pretreatments for enhanced anaerobic digestion with waste activated sludge. *Journal of*
407 *Bioscience and Bioengineering* 95, 271-275.

408 Liao, X., Li, H., Zhang, Y., Liu, C., Chen, Q., 2016. Accelerated high-solids anaerobic
409 digestion of sewage sludge using low-temperature thermal pretreatment. *International*
410 *Biodeterioration and Biodegradation* 106, 141-149.

411 Luo, Y., Guo, W., Ngo, H.H., Nghiem, L.D., Hai, F.I., Zhang, J., Liang, S., Wang, X.C.,
412 2014. A review on the occurrence of micropollutants in the aquatic environment and their

413 fate and removal during wastewater treatment. *Science of The Total Environment* 473–474,
 414 619-641.

415 Malmborg, J., Magnér, J., 2015. Pharmaceutical residues in sewage sludge: Effect of
 416 sanitization and anaerobic digestion. *Journal of Environmental Management* 153, 1-10.

417 McNamara, P.J., Wilson, C.A., Wogen, M.T., Murthy, S.N., Novak, J.T., Novak, P.J., 2012.
 418 The effect of thermal hydrolysis pretreatment on the anaerobic degradation of nonylphenol
 419 and short-chain nonylphenol ethoxylates in digested biosolids. *Water Research* 46, 2937-
 420 2946.

421 Mottet, A., Steyer, J.P., Déléris, S., Vedrenne, F., Chauzy, J., Carrère, H., 2009. Kinetics of
 422 thermophilic batch anaerobic digestion of thermal hydrolysed waste activated sludge.
 423 *Biochemical Engineering Journal* 46, 169-175.

424 Narumiya, M., Nakada, N., Yamashita, N., Tanaka, H., 2013. Phase distribution and removal
 425 of pharmaceuticals and personal care products during anaerobic sludge digestion. *Journal of*
 426 *Hazardous Materials* 260, 305-312.

427 Nghiem, L.D., Koch, K., Bolzonella, D., Drewes, J.E., 2017. Full scale co-digestion of
 428 wastewater sludge and food waste: Bottlenecks and possibilities. *Renewable and Sustainable*
 429 *Energy Reviews* 72, 354-362.

430 Nghiem, L.D., Manassa, P., Dawson, M., Fitzgerald, S.K., 2014. Oxidation reduction
 431 potential as a parameter to regulate micro-oxygen injection into anaerobic digester for
 432 reducing hydrogen sulphide concentration in biogas. *Bioresource Technology* 173, 443-447.

433 Pérez-Elvira, S.I., Fdz-Polanco, F., 2012. Continuous thermal hydrolysis and anaerobic
 434 digestion of sludge. Energy integration study. *Water Science & Technology* 65, 1839-1846.

435 Phothilangka, P., Schoen, M.A., Huber, M., Luchetta, P., Winkler, T., Wett, B., 2008.
 436 Prediction of thermal hydrolysis pretreatment on anaerobic digestion of waste activated
 437 sludge. *Water Science & Technology* 58, 1467-1473.

438 Pilli, S., Yan, S., Tyagi, R.D., Surampalli, R.Y., 2015. Thermal Pretreatment of Sewage
 439 Sludge to Enhance Anaerobic Digestion: A Review. *Critical Reviews in Environmental*
 440 *Science and Technology* 45, 669-702.

441 Reynolds, D.T., Cannon, M., Pelton, T., 2001. Preliminary investigation of recuperative
 442 thickening for anaerobic digestion. *Proceedings of the Water Environment Federation* 2001,
 443 389-410.

444 Sawatdeenarunat, C., Nguyen, D., Surendra, K.C., Shrestha, S., Rajendran, K., Khanal, S.K.,
 445 Oechsner, H., Xie, L., 2016. Anaerobic biorefinery: Current status, challenges, and
 446 opportunities. *Bioresource Technology*.

447 Schieder, D., Schneider, R., Bischof, F., 2000. Thermal hydrolysis (TDH) as a pretreatment
 448 method for the digestion of organic waste. *Water Science and Technology* 41, 181-187.

449 Semblante, G.U., Hai, F.I., Huang, X., Ball, A.S., Price, W.E., Nghiem, L.D., 2015. Trace
 450 organic contaminants in biosolids: Impact of conventional wastewater and sludge processing
 451 technologies and emerging alternatives. *Journal of Hazardous Materials* 300, 1-17.

452 Semblante, G.U., Hai, F.I., Ngo, H.H., Guo, W., You, S.-J., Price, W.E., Nghiem, L.D., 2014.
 453 Sludge cycling between aerobic, anoxic and anaerobic regimes to reduce sludge production
 454 during wastewater treatment: Performance, mechanisms, and implications. *Bioresource*
 455 *Technology* 155, 395-409.

456 Sieger, R., Brady, P., Donovan, J., Shea, T., 2004. White Paper on High Performance
 457 Anaerobic Digestion, In: Federation, W.E., Committee, R.a.B., Subcommittee, B.T. (Eds.).

458 Sihuang, X., Hai, F.I., Xinmin, Z., Wenshan, G., Ngo, H.H., Price, W.E., Nghiem, L.D., 2016.
 459 Anaerobic co-digestion: a critical review of mathematical modelling for performance
 460 optimization. *Bioresource Technology* 222, 498-512.

461 Tadkaew, N., Hai, F.I., McDonald, J.A., Khan, S.J., Nghiem, L.D., 2011. Removal of trace
 462 organics by MBR treatment: The role of molecular properties. *Water Research* 45, 2439-2451.

463 Torpey, N.W., Melbinger, R.N., 1967. Reduction of Digested Sludge Volume by Controlled
 464 Recirculation. *Journal Water Pollution Control* 39, 1464-1474.

465 Tuyet, N.T., Dan, N.P., Vu, N.C., Trung, N.L.H., Thanh, B.X., De Wever, H., Goemans, M.,
 466 Diels, L., 2016. Laboratory-scale membrane up-concentration and co-anaerobic digestion for
 467 energy recovery from sewage and kitchen waste. *Water Science and Technology* 73, 597-606.

468 Valo, A., Carrère, H., Delgenès, J.P., 2004. Thermal, chemical and thermo-chemical pre-
 469 treatment of waste activated sludge for anaerobic digestion. *Journal of Chemical Technology*
 470 *& Biotechnology* 79, 1197-1203.

471 Wijekoon, K.C., McDonald, J.A., Khan, S.J., Hai, F.I., Price, W.E., Nghiem, L.D., 2015.
 472 Development of a predictive framework to assess the removal of trace organic chemicals by
 473 anaerobic membrane bioreactor. *Bioresource Technology* 189, 391-398.

474 Yang, S., Hai, F.I., Price, W.E., McDonald, J., Khan, S.J., Nghiem, L.D., 2016. Occurrence
 475 of trace organic contaminants in wastewater sludge and their removals by anaerobic digestion.
 476 *Bioresource Technology* 210, 153-159.

477 Yang, S., McDonald, J., Hai, F.I., Price, W.E., Khan, S.J., Nghiem, L.D., 2017. The fate of
 478 trace organic contaminants in sewage sludge during recuperative thickening anaerobic
 479 digestion. *Bioresource Technology*.

480 Yang, S., Nghiem, D.L., Bustamante, H., Rys, v.D., Murthy, N.S., 2015. Recuperative
481 thickening: A possible tool to improve anaerobic digestion of wastewater sludge. Proceedings
482 of the Ozwater 2015, Adelaide, Australia.

483