



UNIVERSITY
OF WOLLONGONG
AUSTRALIA

University of Wollongong
Research Online

Faculty of Engineering and Information Sciences -
Papers: Part A

Faculty of Engineering and Information Sciences

2013

Effects of feed and draw solution temperature and transmembrane temperature difference on the rejection of trace organic contaminants by forward osmosis

Ming Xie

University of Wollongong, mx504@uowmail.edu.au

William E. Price

University of Wollongong, wprice@uow.edu.au

Long D. Nghiem

University of Wollongong, longn@uow.edu.au

Menachem Elimelech

Yale University

Publication Details

Xie, M., Price, W. E., Nghiem, L. D. & Elimelech, M. (2013). Effects of feed and draw solution temperature and transmembrane temperature difference on the rejection of trace organic contaminants by forward osmosis. *Journal of Membrane Science*, 438 57-64.

Research Online is the open access institutional repository for the University of Wollongong. For further information contact the UOW Library:
research-pubs@uow.edu.au

Effects of feed and draw solution temperature and transmembrane temperature difference on the rejection of trace organic contaminants by forward osmosis

Abstract

The effects of feed and draw solution temperature and transmembrane temperature difference on the rejection of 12 trace organic contaminants (TrOCs) by two forward osmosis (FO) membranes were investigated. The membrane structure parameter (S) and the reverse salt (NaCl) flux selectivity (RSFS) were constant over the temperature range of 20-40 °C, suggesting that within this range, the solution temperature did not significantly influence the membrane polymeric structure. Draw solution properties, including diffusivity, viscosity, and osmotic pressure, played an important role in water and reverse salt (NaCl) flux behaviour and TrOC rejection. Pure water and salt (NaCl) permeability coefficients of the two forward osmosis membranes increased as both the feed and draw solution temperatures increased from 20 to 40 °C due to an increase in solute diffusivity and a decrease in water viscosity. Rejection of charged TrOCs was higher than that of neutral TrOCs and their rejection was insensitive to temperature variation. On the other hand, rejection of neutral TrOCs decreased significantly when the feed and draw solution temperatures were 40 and 20 °C, respectively, due to the increase in their diffusivity at an elevated temperature. By contrast, rejection of neutral TrOCs increased when the feed and draw solution temperatures were 20 and 40 °C, respectively. The reverse salt (NaCl) flux increased due to an increase in the draw solute diffusivity. In addition, at a higher draw solution temperature, the dilution effect induced by higher water flux and the hindrance effect enhanced by a higher reverse salt (NaCl) flux led to the increase in the rejection of neutral TrOCs.

Keywords

solution, draw, feed, effects, transmembrane, temperature, difference, osmosis, rejection, trace, forward, contaminants, organic

Disciplines

Engineering | Science and Technology Studies

Publication Details

Xie, M., Price, W. E., Nghiem, L. D. & Elimelech, M. (2013). Effects of feed and draw solution temperature and transmembrane temperature difference on the rejection of trace organic contaminants by forward osmosis. *Journal of Membrane Science*, 438 57-64.

**Effects of feed and draw solution temperature and
transmembrane temperature difference on the rejection of
trace organic contaminants by forward osmosis**

Fresh manuscript submitted to

Journal of Membrane Science

March 2013

Ming Xie ¹, William E. Price ², Long D. Nghiem ^{1,*}, and Menachem Elimelech ³

¹ Strategic Water Infrastructure Laboratory, School of Civil, Mining and
Environmental Engineering, University of Wollongong, Wollongong, NSW 2522,
Australia

² Strategic Water Infrastructure Laboratory, School of Chemistry, University of
Wollongong, Wollongong, NSW 2522, Australia

³ Department of Chemical and Environmental Engineering, Yale University, New
Haven, Connecticut 06520-8286, USA

* Corresponding author: Long Duc Nghiem, Email: longn@uow.edu.au; Ph +61 2 4221 4590

1 **Abstract**

2 The effects of feed and draw solution temperature and transmembrane temperature
3 difference on the rejection of 12 trace organic contaminants (TrOCs) by two forward osmosis
4 (FO) membranes were investigated. The membrane structure parameter (S) and the reverse
5 salt (NaCl) flux selectivity ($RSFS$) were constant over the temperature range of 20 to 40 °C,
6 suggesting that within this range, the solution temperature did not significantly influence the
7 membrane polymeric structure. Draw solution properties, including diffusivity, viscosity and
8 osmotic pressure, played an important role in water and reverse salt (NaCl) flux behaviour
9 and TrOC rejection. Pure water and salt (NaCl) permeability coefficients of the two forward
10 osmosis membranes increased as both the feed and draw solution temperatures increased
11 from 20 to 40 °C due to an increase in solute diffusivity and a decrease in water viscosity.
12 Rejection of charged TrOCs was higher than that of neutral TrOCs and their rejection was
13 insensitive to temperature variation. On the other hand, rejection of neutral TrOCs decreased
14 significantly when the feed and draw solution temperatures were 40 and 20 °C, respectively,
15 due to the increase in their diffusivity at an elevated temperature. By contrast, rejection of
16 neutral TrOCs increased when the feed and draw solution temperatures were 20 and 40 °C,
17 respectively. The reverse salt (NaCl) flux increased due to an increase in the draw solute
18 diffusivity. In addition, at a higher draw solution temperature, the dilution effect induced by
19 higher water flux and the hindrance effect enhanced by a higher reverse salt (NaCl) flux led
20 to the increase in the rejection of neutral TrOCs.

21 *Keywords:* Forward osmosis; rejection; temperature; transmembrane temperature difference;
22 trace organic contaminants (TrOCs).

23

24 **1. Introduction**

25 Water scarcity exacerbated by population growth, industrialization, and increasingly
26 irregular weather patterns presents a major challenge to the sustainable development of
27 mankind [1]. Extraction of clean water from unconventional sources, such as municipal
28 wastewater, is arguably feasible from both technical and economic points of view [1-2].
29 These unconventional water sources require extensive treatment for the protection of public
30 health. For example, trace organic contaminants (TrOCs) are ubiquitous in secondary treated
31 effluent and sewage-impacted water bodies; ranging from a few nanograms per litre (ng/L) to
32 several micrograms per litre ($\mu\text{g/L}$) [3-5]. Uncontrolled release of these TrOCs presents a
33 threat to the aquatic environment, with effects such as acute and chronic toxicity to aquatic
34 organisms, accumulation in the ecosystem and loss of habitats and biodiversity, as well as a
35 range of possible adverse effects on human health. Numerous studies have been conducted to
36 enhance the removal capacity of current treatment processes or develop new technologies for
37 better removal of these TrOCs from domestic wastewater and other impaired water resources
38 [2].

39 Forward osmosis (FO), a membrane-based separation technology, has received renewed
40 attention in recent years [6-7]. In lieu of hydraulic pressure, FO utilizes a highly concentrated
41 draw solution to induce the driving force for separation. The transport of water molecules is
42 osmotically driven and contaminants in the feed solution can be rejected by the active layer
43 of the FO membrane. When the draw solute can add value to the extracted water, the diluted
44 draw solution can be directly consumed without any further treatment and FO can be applied
45 as a stand alone process [8]. FO can also be applied in conjunction with a draw solution
46 recovery process, such as reverse osmosis and thermal separation (e.g. conventional column
47 distillation [9-10] and membrane distillation (MD) [11-12]).

48 Temperature is an important factor governing mass transfer in membrane separation
49 processes, including the FO process. In several practical applications of FO, there can be
50 significant temporal and spatial variation in the temperature of feed solutions, such as
51 secondary treated effluent or seawater. Similarly, draw solutions can be at higher
52 temperatures than the feed solution as a result of thermal separation and recycling of the draw
53 solution or using higher temperatures to increase the solubility of the draw solute. Such
54 temperature variations could substantially impact the rejection of TrOCs by the FO process,
55 as also observed in the NF and RO processes [13-14].

56 Several studies have examined the effect of temperature on the permeation of water [15-
57 16] and inorganic salts [17] in the FO process. Generally, it was observed that water and salt
58 permeabilities increased with increasing temperature in the FO process [16-20]. Recent
59 studies have also focused on the impact of the temperature difference between the feed and
60 draw solutions on water and draw solute permeation across FO membranes. Phuntsho et al.
61 [17] examined the water flux behaviour with feed and draw solutions of different temperature
62 and found that water flux increased significantly by increasing draw solution temperature.
63 You et al. [15] proposed that the heat flux generated by the temperature difference between
64 the feed and draw solutions could enhance the water flux due to the decrease in feed solution
65 viscosity and the increase in water diffusivity. However, no studies to date have investigated
66 the effect of temperature and temperature difference between feed and draw solutions on the
67 rejection of contaminants in the feed solution, which is a critical aspect to the deployment of
68 the FO process in wastewater reclamation. Elucidating the impact of temperature on the FO
69 process can be useful for the management of thermal draw solution recovery processes, such
70 as column distillation and MD, and optimization of FO performance with regard to solute
71 rejection and water flux.

72 In this study, an asymmetric cellulose-based FO membrane and a thin-film composite
73 polyamide FO membrane were used to investigate the rejection of 12 TrOCs under four feed
74 and draw solution temperature combinations. Membrane intrinsic properties, namely pure
75 water and salt (NaCl) permeability coefficients and membrane structural parameter, were
76 determined to better elucidate the impact of temperature on water and reverse salt (NaCl)
77 fluxes and TrOC rejection. The implications of the results for FO process performance and
78 optimization are elucidated and discussed.

79 **2. Materials and methods**

80 *2.1. Forward osmosis membranes*

81 Two commercially available FO membranes were used in this study: an asymmetric
82 cellulose triacetate FO membrane (CTA membrane) acquired from Hydration Technology
83 Innovations (Albany, OR) and a thin-film composite polyamide FO membrane (TFC
84 membrane) obtained from Oasys Water (Boston, MA). The CTA membrane has been the
85 subject of numerous previous FO studies and is composed of a cellulose triacetate layer with
86 an embedded woven support mesh [6, 21]. On the other hand, the TFC membrane is a

87 relatively new product. It is reported to have a thin selective polyamide active layer on top of
88 a porous polysulfone support layer [22].

89 2.2. *Forward osmosis system*

90 A bench-scale cross-flow FO system was used (Supplementary Data, Figure S1). The
91 membrane cell was made of acrylic plastic and had channel dimensions of 13 cm long, 9.5
92 cm wide, and 0.2 cm deep. The total effective membrane area was 123.5 cm². Two variable
93 speed gear pumps (Micropump, Vancouver, WA) were used to circulate the feed and draw
94 solutions. Flow rates of the feed and draw solutions were monitored using rotameters and
95 kept constant at 1 L/min (corresponding to a cross flow velocity of 9 cm/s). The draw
96 solution reservoir was placed on a digital balance (Mettler Toledo Inc., Hightstown, NJ) and
97 weight changes were recorded by a computer to calculate the permeate water flux. The
98 conductivity of the draw solution (0.5 M NaCl) was continuously measured using a
99 conductivity probe with a cell constant of 1/cm (Cole-Parmer, Vernon Hills, IL). To maintain
100 constant draw solution concentration, a peristaltic pump was regulated by a conductivity
101 controller to intermittently dose a small volume of a concentrated draw solution (6 M NaCl)
102 into the draw solution reservoir (control accuracy was ± 0.1 mS/cm). The concentrated draw
103 solution makeup reservoir was also placed on the same digital balance. This setup ensured
104 that the transfer of liquid between the two reservoirs did not interfere with the measurement
105 of permeate water flux and that the system could be operated at a constant osmotic pressure
106 driving force during the experiment. Details about the design and operation of this FO system
107 are available elsewhere [23].

108 2.3. *Membrane characterisation*

109 Changes in key properties of the two membranes, including pure water permeability
110 coefficient A , salt (NaCl) permeability coefficient B , and membrane structural parameter of
111 support layer S at different feed and draw solution temperatures were determined using the
112 standard experimental procedure recently proposed by Cath et al. [24]. Briefly, pure water
113 and salt permeability coefficients were measured using deionized water and 2000 mg/L NaCl,
114 respectively. The cross-flow RO filtration system used for this measurement has been
115 described in details elsewhere [23]. Experiments were conducted at 20 ± 0.1 °C and $40 \pm$
116 0.1 °C. The cross-flow velocity was maintained at 25 cm/s. Prior to each measurement, the
117 membranes were compacted at 15 bar with deionised water for at least 12 hours until the
118 permeate water flux had been stabilized. The pure water permeability coefficient was

119 measured at 10 bar (or 145 psi). NaCl was then added to the feed solution at a concentration
 120 of 2000 mg/L to determine the salt (NaCl) permeability coefficient at 10 bar (or 145 psi). The
 121 RO system was stabilised for 2 hours before recording permeate water flux with 2000 mg/L
 122 NaCl solution, J_w^{NaCl} , and taking feed and permeate samples to determine the observed NaCl
 123 rejection, R_o .

124 The water permeability coefficient, A , was determined by dividing the pure water
 125 permeate flux (J_w^{RO}) by the applied hydraulic pressure, ΔP :

$$126 \quad A = J_w^{RO} / \Delta P \quad (1)$$

127 The observed salt (NaCl) rejection, R_o , was calculated from the difference between the bulk
 128 feed (c_b) and permeate (c_p) salt concentrations, $R_o = 1 - c_p/c_b$, and then the salt (NaCl)
 129 permeability coefficient, B , was determined from [25-26]:

$$130 \quad B = J_w^{NaCl} \left(\frac{1 - R_o}{R_o} \right) \exp \left(- \frac{J_w^{NaCl}}{k_f} \right) \quad (2)$$

131 where k_f is the mass transfer coefficient for the crossflow channel of the RO membrane cell.

132 The mass transfer coefficient (k_f) was experimentally determined using a protocol
 133 described in our previous publication [23]. Using the permeate and feed salt concentrations
 134 (and thus, the corresponding osmotic pressures based on the van't Hoff equation, π_p and π_b ,
 135 respectively), the applied pressure (ΔP), the pure water flux (J_w), and the permeate flux with
 136 the 2000 mg/L NaCl solution (J_{salt}) enables the evaluation of the salt concentration at the
 137 membrane surface. This membrane surface concentration is used in the film model for
 138 concentration polarization to determine the mass transfer coefficient k_f [27]:

$$139 \quad k_f = \frac{J_{salt}}{\ln \left[\frac{\Delta P}{\pi_b - \pi_p} \left(1 - \frac{J_{salt}}{J_w} \right) \right]} \quad (3)$$

140 The membrane structural parameter of support layer S was evaluated in the bench-
 141 scale cross-flow FO system described in section 2.2. The water flux, J_w^{FO} , using a 0.5 M
 142 NaCl draw solution and deionized water feed solution was measured with the membrane in
 143 FO mode (i.e., active layer facing the feed solution) under four different feed and draw
 144 solution temperature scenarios (i.e., feed and draw solution temperatures of 20 °C and 20 °C;

145 40 °C and 40 °C; 40 °C and 20 °C; and 20 °C and 40 °C). The membrane structural
 146 parameter S was determined using [28]:

$$147 \quad S = \frac{D_s}{J_w^{FO}} \ln \left(\frac{B + A\pi_{D,b}}{B + J_w + A\pi_{F,m}} \right) \quad (4)$$

148 where D_s is the bulk solution diffusivity of the draw solute, $\pi_{D,b}$ is the bulk osmotic pressure
 149 of the draw solution, and $\pi_{F,m}$ is the osmotic pressure at the membrane surface on the feed
 150 side (zero for deionized water feed). A and B in Eq. 4 were calculated using Eqs. 1 and 2.

151 Reverse salt flux selectivity ($RSFS$) is defined as the ratio of water flux, J_w^{FO} , to
 152 reverse salt (NaCl) flux, J_s^{FO} , in the FO process. The $RSFS$ is independent of the membrane
 153 support layer properties and can quantitatively describe the FO membrane performance [29]:

$$154 \quad RSFS = \frac{A}{B} nRT \quad (5)$$

155 where n is number of dissolved species created by the draw solute (2 for NaCl), T is the draw
 156 solution temperature, and R is the ideal gas constant.

157 2.4. Model trace organic contaminants

158 A total of 12 TrOCs, including nine pharmaceuticals and personal care products, and
 159 three pesticides, were selected for this investigation (Table 1). These TrOCs are frequently
 160 detected in secondary treated effluent and sewage-impacted water bodies at trace levels. They
 161 were also selected to represent a diverse range of physicochemical properties (e.g., charge,
 162 hydrophobicity, and molecular weight). The model TrOCs are small molecular weight
 163 compounds (less than 362 g/mol) with effective hydrophobicity measured by distribution
 164 coefficient ($\log D$) at neutral pH in the range from -0.96 to 5.28. The TrOCs were purchased
 165 as analytical grade standards. A combined stock solution containing 1 g/L of each compound
 166 was prepared in pure methanol. The stock solution was kept at -18 °C in the dark and was
 167 used within one month.

168 **Table 1:** Key physicochemical properties of selected trace organic contaminants (TrOCs).

Compound	Category	Molecular weight (g/mol)	Log D^a (pH 7)	p K_a^a	Diffusion coefficient b ($\times 10^{-6}$ cm 2 /s)	
					20°C	40°C
Amitriptyline	<u>Hydrophilic,</u>	277	2.28	9.18	4.82	7.83

Trimethoprim	charged	290	0.27	7.04	4.99	8.11
Sulfamethoxazole		253	-0.96	5.18	5.99	9.73
Diclofenac		296	1.77	4.18	5.24	8.52
Bezafibrate		362	-0.93	3.29	4.45	7.23
Caffeine		194	-0.63	0.52	7.23	11.75
Atrazine		216	2.64	2.27	5.75	9.34
Primidone	Hydrophilic, neutral	218	0.83	12.26	5.98	9.71
Carbamazepine		236	1.89	13.94	5.84	9.49
Pentachlorophenol		266	2.85	4.68	6.72	10.92
Linuron	Hydrophobic, neutral	249	3.12	12.13	5.9	9.58
Triclosan		290	5.28	7.8	5.58	9.06

169 ^a Values for pK_a and $\log D$ were obtained from the SciFinder Scholar (ACS) database

170 ^b Calculated using USEPA On-line Tools - “Estimated Diffusion Coefficients in Air and
171 Water” (<http://www.epa.gov/athens/learn2model/part-two/onsite/estdiffusion.html>)

172

173 2.5. Trace organic contaminant rejection experiments

174 The TrOCs stock solution was added to a background electrolyte solution (20 mM
175 NaCl and 1 mM NaHCO₃) to obtain a feed solution concentration of 2 µg/L. Either HCl (1
176 M) or NaOH (1 M) was introduced to the feed tank to adjust the initial pH value of the feed
177 solution. A draw solution of 0.5 M NaCl was prepared in Milli-Q water in a volumetric flask.

178 Trace organic contaminant rejection experiments were conducted in the FO mode
179 where the active layer of the membrane faced the feed solution. The initial volumes of the
180 feed and draw solutions were 4 L and 1 L, respectively. Feed and draw solution tanks and
181 pipelines were covered by thermal insulation foam to minimize the water evaporation loss
182 and heat loss. A new FO membrane coupon was used for each experiment. The experiment
183 was concluded when 1 L water permeated through the membrane (i.e., 25% water recovery).
184 Volumes of feed and draw solutions were checked and compared with water flux data at the
185 conclusion of each experiment to make sure that water evaporation from the feed and draw
186 solution tanks was negligible (the difference between measured volume and water flux data
187 less than 3%). Samples from both the feed and draw solutions were collected at the beginning
188 and after 1 L of water had permeated through the FO membrane for solid phase extraction
189 (SPE) and subsequent LC-MS analysis.

190 TrOC rejection is calculated by taking into account the dilution of the draw
191 solution using a mass balance calculation. A dilution factor (DF) is introduced to
192 calculate the concentration of TrOCs in the permeate sample, which is defined as

$$DF = \frac{V_{ds,f}}{V_p} \quad (6)$$

193

194 where $V_{ds,f}$ is the final volume of the draw solution and V_p is the volume of permeate. The
195 TrOC rejection, R (%), is calculated from

$$R = \left(1 - \frac{DF \times C_{ds,f}}{C_{f,0}} \right) \times 100 \quad (7)$$

196

197 Here, DF is the dilution factor obtained from Eq. 6, $C_{ds,f}$ is the final draw solution
198 concentration of the TrOC, and $C_{f,0}$ is the initial feed concentration of the TrOC.

199 2.6. Analytical methods

200 The feed and draw solution samples were extracted using Oasis HLB cartridges (Waters,
201 Milford, MA, USA) prior to LC-MS analysis to determine the concentration of TrOCs. The
202 cartridges were pre-conditioned with 7 mL dichloromethane and methanol (1:1, v/v), 7 mL
203 methanol, and 7 mL reagent water. The sample was 500 mL in volume and was first adjusted
204 to pH 2 – 3 and then loaded onto the cartridges at a flow rate of approximately 2 mL/min.
205 The cartridges were then rinsed with 20 mL of Milli-Q water and dried with a gentle stream
206 of high purity nitrogen for 30 min. The TrOCs were eluted from the cartridges using 7 mL
207 methanol followed by 7 mL dichloromethane and methanol (1:1, v/v) at about 2 mL/min. The
208 eluted samples were evaporated to dryness in a water bath at 40 °C for two to three hours
209 under a gentle stream of high purity nitrogen gas. The extracted residues were then
210 redissolved in 200 μ L methanol solution containing 5 μ g carbamazepine- d_{10} and transferred
211 into 2 mL vials for LC-MS analysis.

212 Analyses of the trace organic contaminants were conducted using a Shimadzu LC-MS
213 system (LC-MS 2020) equipped with an electrospray ionization (ESI) interface. A
214 Phenomenex Kinetex 2.6 μ m C8 column (50 mm \times 4.6 mm) was used as the chromatography
215 column and was maintained at 26 °C inside a column oven (CTO-20A). The mobile phase
216 was Milli-Q water buffered with 0.1% (v/v) formic acid and acetonitrile. The details about
217 the gradient elution are provided in Supplementary Data Table S1. The mobile phase flow
218 rate was 0.5 mL/min and the sample injection volume was 10 μ L. The analytes from the
219 HPLC system were fed directly into a quadrupole mass spectrometer via an ESI source. ESI
220 positive ionization $[M+H]^+$ mode was adopted for caffeine, primidone, trimethoprim,
221 sulfamethoxazole, carbamazepine, bezafibrate, atrazine, linuron and amitriptyline while ESI

222 negative ionization $[M-H]^-$ mode was used for pentachlorophenol, diclofenac and triclosan.
 223 All mass spectra were acquired in selective ion monitoring mode with the detector voltage of
 224 0.9 kV, desolvation line temperature of 250 °C, and heating block temperature of 200 °C.
 225 High purity nitrogen was used as both the nebulizing and drying gas at a flowrate of 1.5 and
 226 10 L/min, respectively. Standard solutions of the analytes were prepared at 1, 10, 50, 100,
 227 500 and 1000 ng/mL, and an internal instrument calibration was carried out with
 228 carbamazepine-d₁₀ as the internal standard. The calibration curves for all the analytes had a
 229 correlation coefficient of 0.99 or higher.

230 3. Results and discussion

231 3.1. Membrane properties

232 The *A* and *B* values (i.e. pure water and salt (NaCl) permeability coefficients) of both
 233 the CTA and TFC membranes increased with an increase in feed solution temperature (Table
 234 2). Results reported here are consistent with a previous study conducted by Wong et al. [30]
 235 and can be attributed to an increase in solute diffusivity and a decrease in viscosity of water
 236 as the temperature increases. While the *A* and *B* values of the TFC membrane were
 237 substantially different from those of the CTA membrane, their responses to temperature
 238 variation are similar. It is noteworthy that the membrane structural parameter *S* was largely
 239 unchanged when the temperature of both the feed and draw solutions increased from 20 to
 240 40 °C (Figure 1). This finding indicates that the membrane polymer structure does not
 241 change when solution temperature increases from 20 to 40 °C. Similarly, no statistically
 242 significant changes in the *S* value of the two membranes could be observed when the feed and
 243 draw solution temperatures were 40 and 20 °C, respectively. The p-values of the one-sample
 244 t-test on the structural parameter *S* of the CTA and TFC membranes (n=4) were 0.08 and 0.07,
 245 respectively. In addition, the calculation of the *S* value using Eq. 4 can result in some inherent
 246 error when temperatures of feed and draw solutions are different [30]. Because there was heat
 247 transfer induced by the temperature difference between the feed and draw solution, the *A* and
 248 *B* values of the membrane active layer were likely not the same as the values used in Eq. 4,
 249 which were obtained from RO experiments at the same feed solution temperature.

250 **Table 2:** Water and salt (NaCl) permeability coefficients of CTA and TFC FO membranes at
 251 different temperatures (mean value ± standard deviation from two membrane samples).

Temperature	Membrane	Water permeability	Salt (NaCl) permeability
-------------	----------	--------------------	--------------------------

(°C)		coefficient, <i>A</i> ($\times 10^{-12}$ m/s·Pa)	coefficient, <i>B</i> ($\times 10^{-8}$ m/s)
20	CTA	1.81 ± 0.27	6.81 ± 0.11
40		1.94 ± 0.09	9.92 ± 0.25
20	TFC	13.1 ± 0.07	4.56 ± 0.37
40		21.7 ± 0.31	8.06 ± 0.14

252 **Table 3:** Key properties of the 0.5 M NaCl draw solution.

Thermodynamic property	Unit	Temperature (°C)	
		20	40
Osmotic pressure ^a	bar	24.05	25.69
Viscosity ^[39]	μPa·s	1043.2	684.8
Diffusion coefficient ^[40]	$\times 10^{-9}$ m ² /s	1.2	2.51

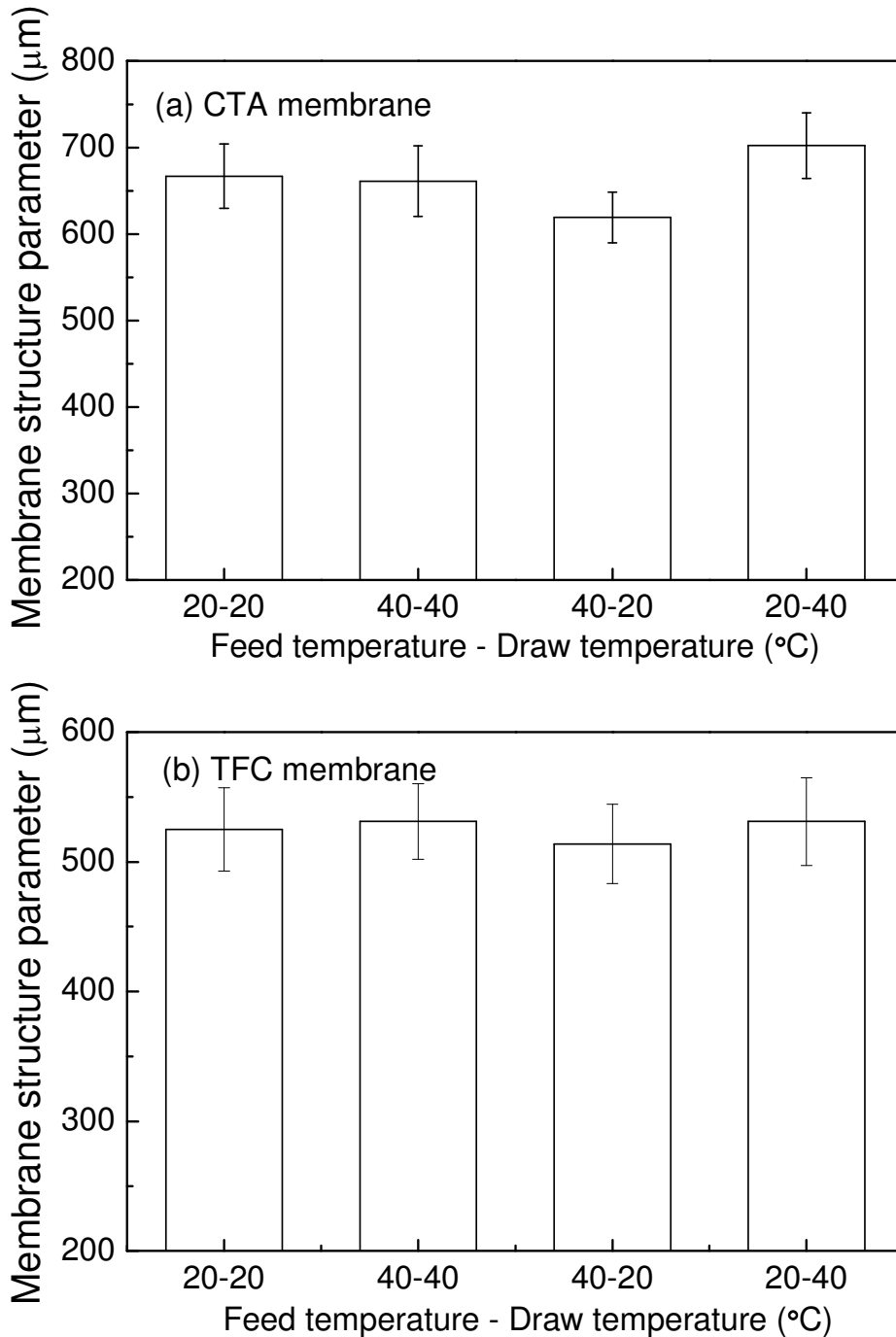
253 ^a Calculated using OLI Stream Analyser (OLI Systems, Inc., Morris Plains, NJ

254

255

256

257



258

259 **Figure 1:** Membrane support layer structural parameter for (a) CTA and (b) TFC membranes
 260 at varying feed and draw solution temperatures. Experimental conditions were as follows: FO
 261 mode (i.e. feed solution facing membrane active layer), deionized water as feed solution,
 262 draw solution = 0.5 M NaCl, and cross-flow rate = 1 L/min for both sides (corresponding to
 263 cross flow velocity = 9 cm/s). Four different feed and draw solution temperature scenarios
 264 were used: under the condition of same feed and draw solution temperature at 20 and 40 °C;
 265 under the conditions of feed temperature at 40 °C and draw solution temperature at 20 °C,

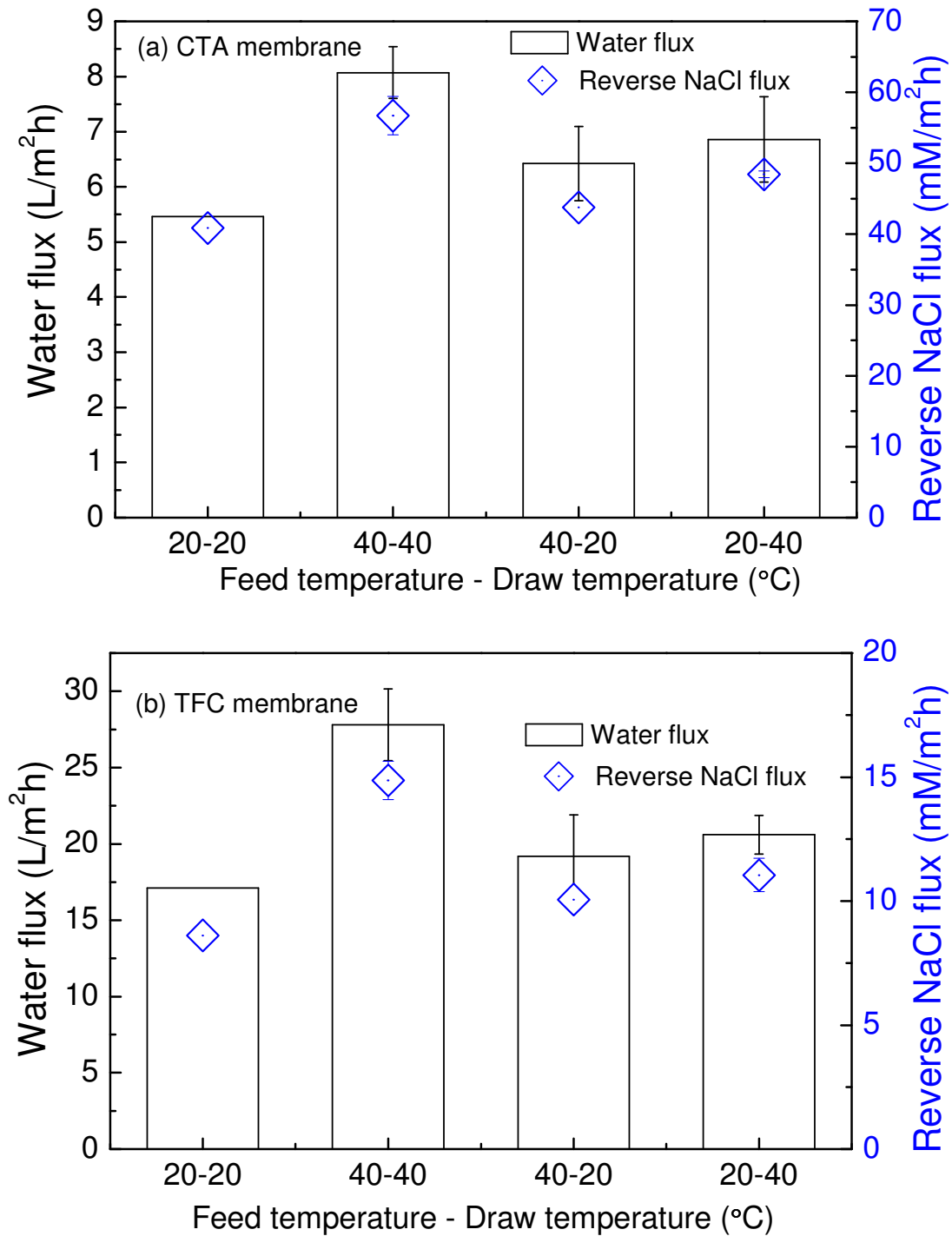
266 and feed temperature at 20 °C and draw solution temperature at 40 °C. Error bars represent
267 standard deviation of data obtained from two repeated experiments.

268

269 3.2. *Water and reverse salt (NaCl) fluxes*

270 Both water and reverse salt (NaCl) fluxes were significantly impacted by feed and
271 draw solution temperatures (Figure 2). When the feed and draw solution temperatures were
272 the same (denoted 20-20 and 40-40 in Figure 2), the water and reverse salt (NaCl) fluxes of
273 the CTA and TFC membranes substantially increased as the solution temperature increased
274 from 20 to 40°C. This observation is in good agreement with the increase in the membrane *A*
275 and *B* values reported in Section 3.1 and the literature [17, 31].

276 In addition to the isothermal conditions investigated by these two previous studies [17,
277 31], the effects of transmembrane temperature difference between the feed and draw
278 solutions on solute and water transport were also examined in the current study. Water and
279 reverse salt (NaCl) fluxes of both membranes increased slightly when either the feed or draw
280 solution temperature increased to 40 °C and the other remained at 20 °C (denoted 40-20 and
281 20-40 in Figure 2) compared to the isothermal condition where the feed and draw solution
282 temperatures were both at 20 °C. The increase in feed solution temperature from 20 to 40 °C
283 enhanced the diffusivity of water molecules, thereby increasing the water and reverse salt
284 (NaCl) fluxes. On the other hand, the increase of draw solution temperature from 20 to 40 °C
285 decreased draw solution viscosity and increased the draw solute diffusivity (Table 3), thereby
286 increasing the water and reverse salt (NaCl) fluxes.



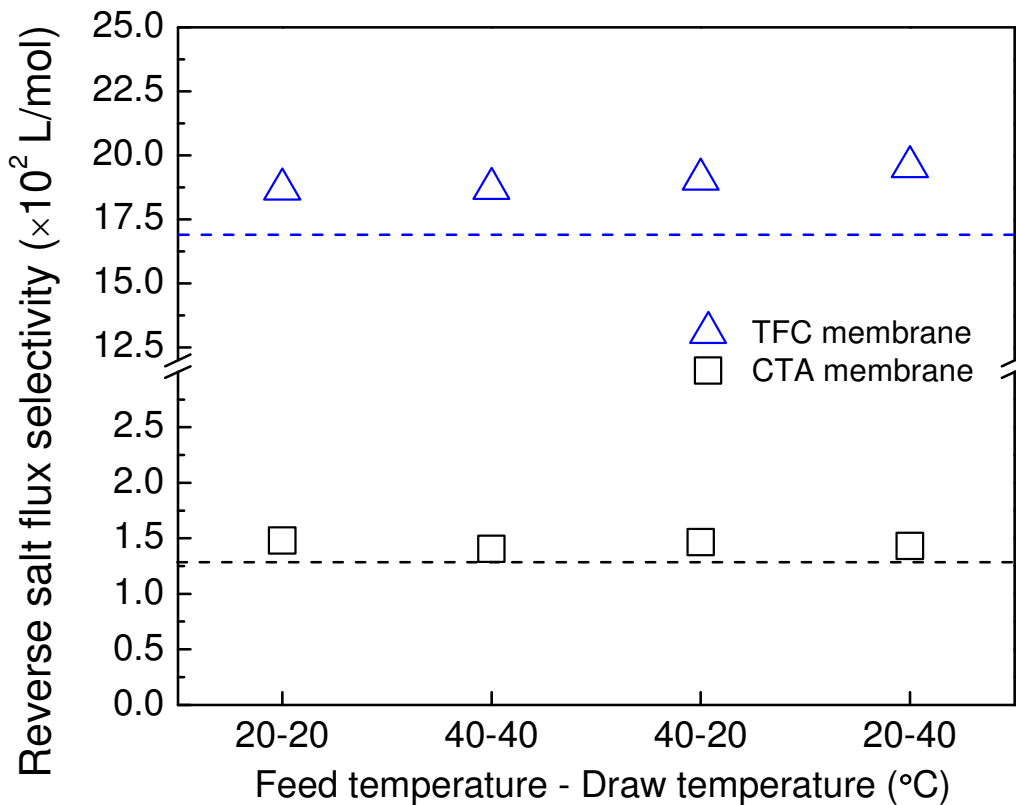
287

288 **Figure 2:** Water and reverse NaCl fluxes of (a) CTA and (b) TFC membranes at varying feed
 289 and draw solution temperatures. Experimental conditions were described in Figure 1.

290

291 It is noteworthy that there was no discernible variation in the *RSFS* value of either
 292 membrane regardless of the feed or draw solution temperatures (Figure 3). In addition, the
 293 determined *RSFS* values obtained from the FO experiments (symbols) are almost identical to

294 those calculated from the intrinsic properties of the membranes (dashed line). The *RSFS* was
 295 independent of the membrane support layer properties and reflected the polymer structure of
 296 the membrane active layer. As a result, the insignificant variation in *RSFS* and the membrane
 297 *S* value (section 3.1) suggests that the membrane polymer structure did not change
 298 significantly within the temperature range of 20 to 40 °C. The water and reverse salt (NaCl)
 299 fluxes behaviour at different feed and draw solution temperature conditions was attributed
 300 mostly to the temperature-dependent properties of feed and draw solutions.



301 3.3.

302 *TrOC rejection performance*

303 Overall, TrOC rejections by the TFC membrane were considerably higher than those by
 304 the CTA membrane. This can be attributed to the intrinsic properties of the TFC and CTA
 305 membranes. The TFC membrane has a smaller *B* value (Table 2) and higher *RSFS* value than
 306 the CTA membrane (Figures 3). Nevertheless, with respect to the rejection of TrOCs, both
 307 the CTA and TFC membranes responded to the variation in temperature and transmembrane
 308 temperature difference in a similar manner (Figure 4).

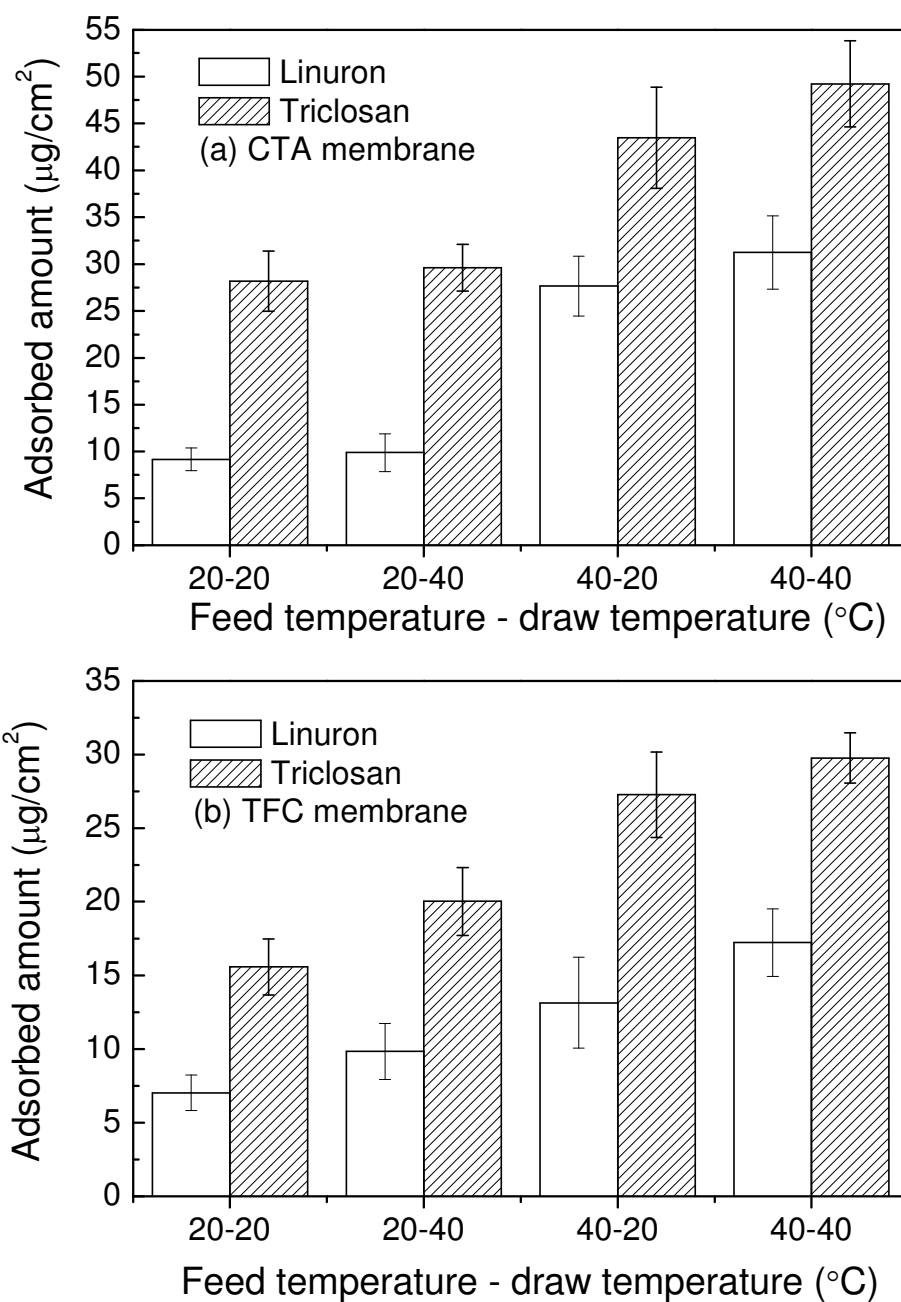
309 **Figure 4:** Rejections of 12 model TrOCs by the (a) CTA and (b) TFC FO membranes
 310 at varying feed and draw solution temperatures. The experimental conditions were as follows:
 311 FO mode (i.e. feed solution facing membrane active layer), initial concentrations of 12 trace

312 organic contaminants in the feed = 2 $\mu\text{g/L}$, pH = 7, background electrolyte contained 20 mM
313 NaCl and 1 mM NaHCO_3 , draw solution = 0.5 M NaCl, and cross flow rate = 1 L/min for
314 both sides (corresponding to cross flow velocity = 9 cm/s). Four different feed and draw
315 solution temperature scenarios were used: under the condition of same feed and draw solution
316 temperature at 20 and 40 $^\circ\text{C}$ (denoted as F40-D40 and F20-D20); under the conditions of feed
317 temperature at 40 $^\circ\text{C}$ and draw solution temperature at 20 $^\circ\text{C}$ (denoted as F40-D20), and feed
318 temperature at 20 $^\circ\text{C}$ and draw solution temperature at 40 $^\circ\text{C}$ (denoted as F20-D40). Error
319 bars represent standard deviation of four measurements in two repeated experiments. The
320 rejection behaviours of charged and neutral TrOCs (Table 1) significantly differ from each
321 other (Figure 4). In an aqueous solution, charged TrOCs are hydrated and the hydration of
322 charged TrOCs significantly increases their apparent molecular sizes [32]. In addition, at the
323 experimental pH value used in this study (pH 6.5), the CTA and TFC membranes are both
324 negatively charged and electrostatic interaction is an important rejection mechanism of
325 charged solutes [33-34]. Thus, rejections of charged TrOCs were notably higher than those of
326 neutral TrOCs.

327 Temperature and transmembrane temperature difference only exerted a small influence
328 on the rejection of charged TrOCs by the CTA and TFC membranes (Figure 4). By contrast,
329 while rejection of both hydrophobic and hydrophilic neutral TrOCs increased with their
330 molecular weight, their rejections varied significantly depending on the feed and draw
331 solution temperatures. Results reported here demonstrate an intricate relationship between
332 rejection of neutral TrOCs and temperature-dependent solvent and solute properties, such as
333 solution viscosity and solute diffusivity. Overall, rejection of neutral TrOCs increased in the
334 following order of feed and draw solution temperatures (in $^\circ\text{C}$): 40 – 40 < 40 – 20 < 20 – 20 <
335 20 – 40.

336 When the feed and draw solution temperatures are the same, the low rejection of neutral
337 TrOCs at high solution temperature can be attributed to an increase in solute partition into the
338 membrane and diffusion coefficient. The permeation of neutral TrOCs is governed by
339 “solution-diffusion” mechanism [35-36], where TrOCs first partition into the membrane
340 active layer and then diffuse through it. The diffusion of TrOCs can be described by Fick’s
341 law and it is proportional to the diffusion coefficient [37]. A higher solution temperature
342 favoured the partition of hydrophobic neutral TrOCs into the membrane. Notably, the
343 adsorbed amounts of linuron and triclosan (with their Log D values > 3) increased by one
344 order of magnitude with the increase of feed solution temperature from 20 to 40 $^\circ\text{C}$ (Figure 5).

345 At the same time, the diffusion coefficients of neutral TrOCs increase significantly (Table 1),
 346 thereby leading to a markedly decrease in their rejections (Figure 4).



347
 348 **Figure 5:** Adsorbed amount of linuron and triclosan by (a) CTA and (b) TFC membranes at
 349 varying feed and draw solution temperatures. The adsorption amount was calculated using
 350 mass balance. Experimental conditions were described in Figure 4.

351 Transmembrane temperature difference between the feed and draw solutions impacts
 352 the water and reverse salt flux, which can subsequently influence the rejection of neutral
 353 TrOCs. The diffusion coefficient of the draw solute increased with the increase in draw

354 solution temperature, resulting in an increase in both the water and reverse salt flux (Figure 2).
355 Consequently, the increase in water flux can directly contribute to an increase in rejection,
356 which is similar to that observed in the nanofiltration or reverse osmosis processes [38]. In
357 addition, the increase in reverse salt flux can retard the forward diffusion of neutral TrOCs
358 [23], thereby leading to higher rejection of these contaminants.

359 *3.4. Implications for FO systems*

360 The enhanced rejection of neutral TrOCs and improved water flux reported here have
361 important implications for the integration of the FO process with a thermally-driven
362 separation process, such as MD or conventional column distillation, for recovering the draw
363 solutes. These results highlight the potential of FO for water production from reclaimed
364 wastewater and other unconventional water sources that may be impaired with TrOCs. MD
365 has been widely recognised as a potential draw solution recovery process [11-12]. In the MD
366 process, solar thermal or low-grade heat can be utilised to increase the feed solution (i.e. the
367 diluted draw solution of the FO process) temperature for the extraction of water vapour
368 across a microporous membrane which is condensed to the liquid form. Thus, integrating the
369 MD process with FO can not only improve the water flux and TrOC rejection by FO process
370 but also reduce the carbon footprint of the overall treatment system. Similarly, the enhanced
371 performance at a high draw solution temperature can facilitate practical deployment of
372 thermolytic salts, such as ammonium carbonate, as the draw solutes for FO [9-10].

373 **4. Conclusions**

374 Results reported here demonstrate that feed and draw solution temperature and
375 transmembrane temperature difference have a significant impact on FO water and reverse salt
376 (NaCl) fluxes as well as TrOC rejection. The membrane structural parameter (S) and the
377 reverse salt flux selectivity ($RSFS$) did not change significantly in the temperature range of 20
378 to 40°C, indicating that any thermal-induced changes in the membrane polymer structure
379 would be negligible. The increase in water and solute diffusivities at higher temperatures and
380 the temperature-dependent draw solution properties governed the water and reverse salt
381 (NaCl) flux behaviour and TrOC rejection. Because electrostatic interaction was an important
382 rejection mechanism, rejection of charged TrOCs was higher than that of neutral TrOCs and
383 their rejection was insensitive to temperature variation. Rejection of neutral TrOCs decreased
384 significantly as the feed solution temperature increased from 20 to 40 °C. This decrease
385 resulted from the enhanced diffusivity of neutral TrOCs at an elevated temperature. By

386 contrast, rejection of neutral TrOCs increased when the feed and draw solution temperatures
387 were 20 and 40°C, respectively. This increase in the rejection of neutral TrOCs could be
388 attributed to the changes in properties of the draw solution. Water flux enhanced by higher
389 osmotic pressure led to a dilution effect. At the same time, an increase in the reverse salt
390 (NaCl) caused by a higher draw solute diffusivity further hinder the forward diffusion of the
391 neutral TrOCs from the feed to the draw solution.

392 **5. Acknowledgments**

393 The authors would like to thank Hydration Technology Innovations and Oasys Water for
394 providing membrane samples. University of Wollongong is acknowledged for the provision
395 of a doctoral scholarship to Ming Xie.

396 **6. References**

- 397 [1] M. Elimelech, W.A. Phillip, The Future of Seawater Desalination: Energy, Technology,
398 and the Environment, *Science*, 333 (2011) 712-717.
- 399 [2] M.A. Shannon, P.W. Bohn, M. Elimelech, J.G. Georgiadis, B.J. Marinas, A.M. Mayes,
400 Science and technology for water purification in the coming decades, *Nature*, 452 (2008)
401 301-310.
- 402 [3] T. Basile, A. Petrella, M. Petrella, G. Boghetich, V. Petruzzelli, S. Colasuonno, D.
403 Petruzzelli, Review of Endocrine-Disrupting-Compound Removal Technologies in Water and
404 Wastewater Treatment Plants: An EU Perspective, *Industrial & Engineering Chemistry*
405 *Research*, 50 (2011) 8389-8401.
- 406 [4] M. Carballa, F. Omil, J.M. Lema, M.a. Llompert, C. García-Jares, I. Rodríguez, M.
407 Gómez, T. Ternes, Behavior of pharmaceuticals, cosmetics and hormones in a sewage
408 treatment plant, *Water Research*, 38 (2004) 2918-2926.
- 409 [5] S.A. Snyder, P. Westerhoff, Y. Yoon, D.L. Sedlak, Pharmaceuticals, Personal Care
410 Products, and Endocrine Disruptors in Water: Implications for the Water Industry,
411 *Environmental Engineering Science*, 20 (2003) 449-469.
- 412 [6] T.Y. Cath, A.E. Childress, M. Elimelech, Forward osmosis: Principles, applications, and
413 recent developments, *Journal of Membrane Science*, 281 (2006) 70-87.
- 414 [7] S. Zhao, L. Zou, C.Y. Tang, D. Mulcahy, Recent developments in forward osmosis:
415 Opportunities and challenges, *Journal of Membrane Science*, 396 (2012) 1-21.
- 416 [8] L.A. Hoover, W.A. Phillip, A. Tiraferri, N.Y. Yip, M. Elimelech, Forward with Osmosis:
417 Emerging Applications for Greater Sustainability, *Environmental Science & Technology*, 45
418 (2011) 9824-9830.
- 419 [9] R.L. McGinnis, M. Elimelech, Energy requirements of ammonia-carbon dioxide forward
420 osmosis desalination, *Desalination*, 207 (2007) 370-382.
- 421 [10] J.R. McCutcheon, R.L. McGinnis, M. Elimelech, A novel ammonia-carbon dioxide
422 forward (direct) osmosis desalination process, *Desalination*, 174 (2005) 1-11.

- 423 [11] C.R. Martinetti, A.E. Childress, T.Y. Cath, High recovery of concentrated RO brines
424 using forward osmosis and membrane distillation, *Journal of Membrane Science*, 331 (2009)
425 31-39.
- 426 [12] T.Y. Cath, D. Adams, A.E. Childress, Membrane contactor processes for wastewater
427 reclamation in space: II. Combined direct osmosis, osmotic distillation, and membrane
428 distillation for treatment of metabolic wastewater, *Journal of Membrane Science*, 257 (2005)
429 111-119.
- 430 [13] J.M. Arsuaga, M.J. López-Muñoz, J. Aguado, A. Sotto, Temperature, pH and
431 concentration effects on retention and transport of organic pollutants across thin-film
432 composite nanofiltration membranes, *Desalination*, 221 (2008) 253-258.
- 433 [14] R.R. Sharma, S. Chellam, Temperature and concentration effects on electrolyte transport
434 across porous thin-film composite nanofiltration membranes: Pore transport mechanisms and
435 energetics of permeation, *Journal of Colloid and Interface Science*, 298 (2006) 327-340.
- 436 [15] S.-J. You, X.-H. Wang, M. Zhong, Y.-J. Zhong, C. Yu, N.-Q. Ren, Temperature as a
437 factor affecting transmembrane water flux in forward osmosis: Steady-state modeling and
438 experimental validation, *Chemical Engineering Journal*, 198–199 (2012) 52-60.
- 439 [16] S. Zhao, L. Zou, Effects of working temperature on separation performance, membrane
440 scaling and cleaning in forward osmosis desalination, *Desalination*, 278 (2011) 157-164.
- 441 [17] S. Phuntsho, S. Vigneswaran, J. Kandasamy, S. Hong, S. Lee, H.K. Shon, Influence of
442 temperature and temperature difference in the performance of forward osmosis desalination
443 process, *Journal of Membrane Science*, 415–416 (2012) 734-744.
- 444 [18] J.R. McCutcheon, M. Elimelech, Influence of concentrative and dilutive internal
445 concentration polarization on flux behavior in forward osmosis, *Journal of Membrane
446 Science*, 284 (2006) 237-247.
- 447 [19] B.S. Chanukya, S. Patil, N.K. Rastogi, Influence of concentration polarization on flux
448 behavior in forward osmosis during desalination using ammonium bicarbonate, *Desalination*.
- 449 [20] C.A. Nayak, N.K. Rastogi, Forward osmosis for the concentration of anthocyanin from
450 *Garcinia indica* Choisy, *Separation and Purification Technology*, 71 (2010) 144-151.
- 451 [21] J.R. McCutcheon, M. Elimelech, Influence of membrane support layer hydrophobicity
452 on water flux in osmotically driven membrane processes, *Journal of Membrane Science*, 318
453 (2008) 458-466.
- 454 [22] N.Y. Yip, A. Tiraferri, W.A. Phillip, J.D. Schiffman, M. Elimelech, High Performance
455 Thin-Film Composite Forward Osmosis Membrane, *Environmental Science & Technology*,
456 44 (2010) 3812-3818.
- 457 [23] M. Xie, L.D. Nghiem, W.E. Price, M. Elimelech, Comparison of the removal of
458 hydrophobic trace organic contaminants by forward osmosis and reverse osmosis, *Water
459 Research*, 46 (2012) 2683-2692.
- 460 [24] T.Y. Cath, M. Elimelech, J.R. McCutcheon, R.L. McGinnis, A. Achilli, D. Anastasio,
461 A.R. Brady, A.E. Childress, I.V. Farr, N.T. Hancock, J. Lampi, L.D. Nghiem, M. Xie, N.Y.
462 Yip, Standard Methodology for Evaluating Membrane Performance in Osmotically Driven
463 Membrane Processes, *Desalination*, 312 (2013) 31-38.
- 464 [25] R.W. Baker, *Membrane technology and applications*, 2nd edition ed., John Wiley &
465 Sons, Ltd, New York, NY, 2004.

- 466 [26] M. Mulder, Basic Principles of Membrane Technology, 2nd edition, 2nd edition ed.,
467 Kluwer Academic Publishers, Dordrecht, The Netherlands, 2001.
- 468 [27] I. Sutzkover, D. Hasson, R. Semiat, Simple technique for measuring the concentration
469 polarization level in a reverse osmosis system, *Desalination*, 131 (2000) 117-127.
- 470 [28] S. Loeb, L. Titelman, E. Korngold, J. Freiman, Effect of porous support fabric on
471 osmosis through a Loeb-Sourirajan type asymmetric membrane, *Journal of Membrane*
472 *Science*, 129 (1997) 243-249.
- 473 [29] W.A. Phillip, J.S. Yong, M. Elimelech, Reverse Draw Solute Permeation in Forward
474 Osmosis: Modeling and Experiments, *Environmental Science & Technology*, 44 (2010)
475 5170-5176.
- 476 [30] M.C.Y. Wong, K. Martinez, G.Z. Ramon, E.M.V. Hoek, Impacts of operating conditions
477 and solution chemistry on osmotic membrane structure and performance, *Desalination*, 287
478 (2012) 340-349.
- 479 [31] J.R. McCutcheon, R.L. McGinnis, M. Elimelech, Desalination by ammonia-carbon
480 dioxide forward osmosis: Influence of draw and feed solution concentrations on process
481 performance, *Journal of Membrane Science*, 278 (2006) 114-123.
- 482 [32] L.D. Nghiem, A.I. Schafer, M. Elimelech, Role of electrostatic interactions in the
483 retention of pharmaceutically active contaminants by a loose nanofiltration membrane,
484 *Journal of Membrane Science*, 286 (2006) 52-59.
- 485 [33] M. Xie, W.E. Price, L.D. Nghiem, Rejection of pharmaceutically active compounds by
486 forward osmosis: Role of solution pH and membrane orientation, *Separation and Purification*
487 *Technology*, 93 (2012) 107-114.
- 488 [34] A.A. Alturki, J.A. McDonald, S.J. Khan, W.E. Price, L.D. Nghiem, M. Elimelech,
489 Removal of trace organic contaminants by the forward osmosis process, *Separation and*
490 *Purification Technology*, 103 (2013) 258-266.
- 491 [35] N.T. Hancock, W.A. Phillip, M. Elimelech, T.Y. Cath, Bidirectional Permeation of
492 Electrolytes in Osmotically Driven Membrane Processes, *Environmental Science &*
493 *Technology*, 45 (2011) 10642-10651.
- 494 [36] L.D. Nghiem, A.I. Schäfer, M. Elimelech, Nanofiltration of Hormone Mimicking Trace
495 Organic Contaminants, *Separation Science and Technology*, 40 (2005) 2633-2649.
- 496 [37] E.L. Cussler, *Diffusion mass transfer in fluid systems*, 2009.
- 497 [38] L.D. Nghiem, A.I. Schäfer, M. Elimelech, Removal of Natural Hormones by
498 Nanofiltration Membranes: Measurement, Modeling, and Mechanisms, *Environmental*
499 *Science & Technology*, 38 (2004) 1888-1896.
- 500 [39] J. Kestin, H.E. Khalifa, R.J. Correia, Tables of the dynamic and kinematic viscosity of
501 aqueous NaCl solutions in the temperature range 20--150 [degree]C and the pressure range
502 0.1--35 MPa, *Journal of Physical and Chemical Reference Data*, 10 (1981) 71-88.
- 503 [40] V.M.M. Lobo, Mutual diffusion coefficients in aqueous electrolyte solutions (Technical
504 Report), *Pure and Applied Chemistry*, 65 (1993) 2613-2640.